

# Tools for monitoring aquatic environments to identify anthropic effects

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Received: 2 May 2017 / Accepted: 26 December 2017 / Published online: 5 January 2018  
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**Abstract** Anthropic activities are directly related to the contamination of aquatic ecosystems owing to the release of numerous chemicals from agricultural and urban waste. These contaminants cause environmental degradation and a decrease in the availability of water quality. The objective of this search was to evaluate the efficiency of physicochemical, chemical, and microbiological tests; extraction of chlorophyll *a*; and genetic parameters to identify anthropic activities and weather condition effects on the stream water quality and the consequences of its use by the population. The

physicochemical parameters were within the limits allowed by the Brazilian law. However, contamination by metals (Cd 0.510 mg L<sup>-1</sup>, Co 0.405 mg L<sup>-1</sup>, and Ni 0.316 mg L<sup>-1</sup>) has been found at various collection points to be more than the allowable values. The antibiotic oxytetracycline was detected in stream water in quantities of up to 89 µg L<sup>-1</sup>. In relation to microbiological contamination, *Escherichia coli* and *Pseudomonas* spp. have been isolated. The averages of chlorophyll *a* were up to 0.15558 mg cm<sup>-2</sup>. Genetic tools identified greater number of micronuclei and DNA damage in periods that showed lower rainfall rates and lower amounts of metals. The analysis used for monitoring was efficient to verify the interference that animal breeding and planting of different cultures have caused on that stream. Thus, the continued use of this water for drinking, irrigation of vegetables, and recreational activities makes the population susceptible to contamination by bacteria and creates conditions for the development of genetic alterations in the long run.

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**Keywords** Surface water · Microbiology · Genetic parameter · Environmental pollutants

## Introduction

The lack of basic sanitation, improper disposal of solid waste, excessive use of agro-chemicals, leaching in soil without a vegetation cover, and untreated effluent discharges, in large volumes, are the main problems

regarding the contamination of watercourses (Callisto and Gonçalves Jr 2002; Tiefenthaler et al. 2011).

Activities related to agriculture and animal breeding also contribute to increase the amount of chemicals released into the environment, in particular, into aquatic environments. The need for increased productivity, in search for better agricultural yields, promotes the indiscriminate use of chemical compounds, with consequences to the environment and public health (Ventura et al. 2008). In this context, Brazil is the world's fourth largest consumer of chemical substances used in agricultural activities, and the biggest expense is made with pesticides, mainly for soybean crops (approximately 35%) (Santos and Monteiro 2004).

The release of chemical compounds in water results in higher concentrations than those found naturally, and as a result, there are changes in the microenvironment and the natural physicochemical composition of the rivers, the vegetation that covers the banks, the color of the water, and the existing biota (Callisto et al. 2005). The use of different parameters in a biomonitoring study allows the evaluation of the impact on the environment of contaminated urban, industrial, and agricultural wastes (Rocha et al. 2015).

The evaluation of physicochemical and biological parameters in water bodies is an evidence of the anthropic, agrarian, and industrial interferences (Goulart and Callisto 2003). The physical and chemical parameters measured are temperature, pH, and oxygen saturation (Okeke et al. 2011; Gemmell and Schmidt 2013). For chemicals, tests are made on metals and organic compounds (Schipper et al. 2008). Moreover, for microbiological parameters, analyses of coliforms, *Escherichia coli*, *Salmonella* spp., and *Pseudomonas* spp. (Rowny and Stewart 2012; Kittinger et al. 2013) and extraction of chlorophyll *a* (Siqueira and Rodrigues 2009) are performed. Furthermore, for biological parameters, animal bioassays are carried out using the micronucleus piscine test with *Astyanax lacustris* (Silva et al. 2011) and the Comet Assay with *A. lacustris* (Gontijo and Tice 2003).

The “Curral de Arame” stream receives constant anthropic interference that does not exceed a length of 30 km and is located on the watershed of the sub-basins of Brilhante and Dourados Rivers, both belonging to the Ivinhema River Basin (de Lima 1999) and in a region of the Cwa climate type (humid mesothermal climate, warm summers, and dry winters), where the temperature of the coldest months (June and July) is below 18 °C and

that of the hottest month (January) is above 22 °C. In addition, the total rainfall in summer exceeds by more than ten times the lowest monthly rainfall (July) (Fietz and Fisch 2008). The study assessed the efficiency of analysis for monitoring of aquatic environments (physicochemical, chemical, and microbiological tests; extraction of chlorophyll *a*; and genetic parameters) to identify anthropic activities and weather condition interferences on the stream water quality.

## Materials and methods

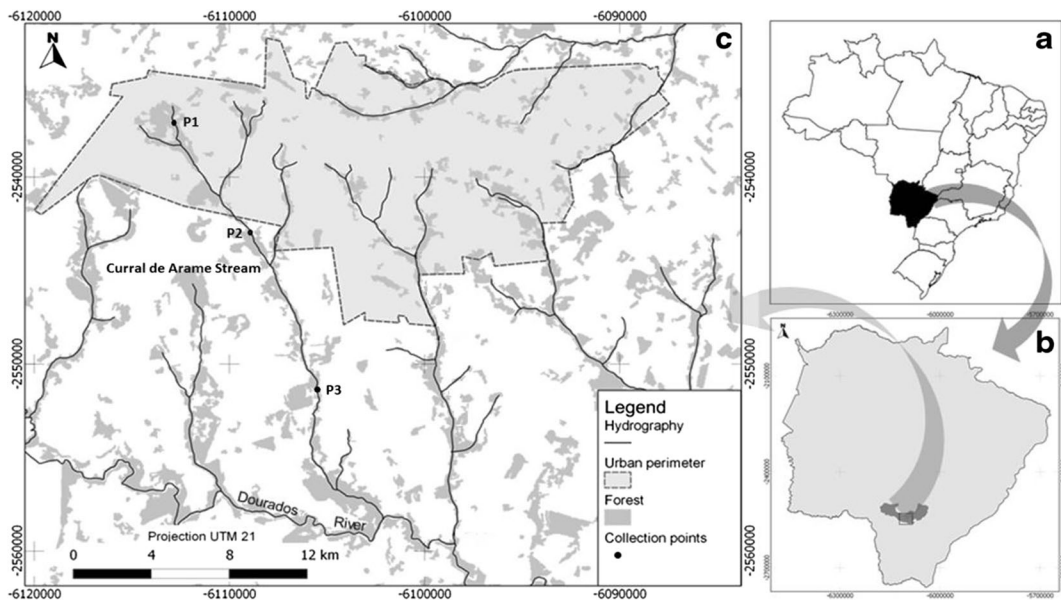
### Study area

The urban perimeter of Dourados is at an average altitude of 430 m above sea level (m a.s.l.), latitude of 22°13'16"S, and longitude of 54°48'20"W, with a total area of 4,086,237 km<sup>2</sup> and 212,870 inhabitants (IBGE—Instituto Brasileiro de Geografia e Estatística 2013). Among the streams located in the region of Grande Dourados (that covers 10 municipalities in its vicinity), “Curral de Arame” stream, located in the basin of the Dourados River (MS, Brazil) is influenced by local agriculture. The water source is located in a native bushland reserve, passing through places of intense agricultural activity (sugar cane, corn, and soy), and flowing into the Dourados River, where water collection is performed for public consumption of the entire city.

### Collection points

The monitoring of the “Curral de Arame” stream was conducted in the months of January, March, May, August, September, and November 2013, with a total of six collections in each sampling spot, namely Points 1, 2, and 3 (Fig. 1).

Point 1 (P1) (22°12'48.83"S and 054°54'41.06"W) is located in the source of the stream within the protected area, although inhabited. In addition, Point 2 (P2) (22°15'36.69"S and 054°52'40.81"W) is located 0.5 km of sugarcane plantation areas and receives agrochemical effluents from these. Furthermore, Point 3 (P3) (22°20'16.07"S and 054°50'48.23"W) is located 7.4 km away from the mouth of the Dourados River receiving interference from agrochemicals used in sugarcane plantation (Point 2) and corn and soybean crops planted near the banks.



**Fig. 1** Map of the State of Mato Grosso do Sul (a), Dourados municipality (b), and collection points (P1, P2, and P3) at the Curral de Arame Stream (c)

### Water collection procedures

Water samples for metal and organic compound analysis were collected in amber bottles of 1 L. The samples were frozen ( $-20\text{ }^{\circ}\text{C}$ ) for the determination of organic compounds, and for metals, the water were acidified with nitric acid (1 mL) and refrigerated (2 to  $8\text{ }^{\circ}\text{C}$ ).

Water samples were collected in sterile glass bottles of 500 mL by submerging to a 20 cm depth on the stream bank for microbiological testing. The samples were transported in a refrigerated cooler with an 8-h time interval between the collection of the water samples and the microbiological test.

For the evaluation of other parameters, water samples were collected in previously cleaned 20 L flasks.

### Physico-chemical analyses

A Hanna HI 9829 multi-parameter probe was used to measure in situ the following: water temperature ( $^{\circ}\text{C}$ ), hydrogen potential (pH), and dissolved oxygen ( $\text{mg L}^{-1}$ ). The rainfall index was calculated from data of means of 15 days of rain that preceded the collection. These data came from the Embrapa Centro Oeste (EMBRAPA 2014).

### Metal detection

#### *Instrumentation, sample preparation, and analysis*

An Agilent AA 240FS flame atomic absorption spectrometer (Agilent Technologies®, Santa Clara, CA, USA) equipped with a deuterium background corrector and mono element hollow cathode lamps was used throughout this study. High-purity acetylene (99.7%, White Martins®, Brazil) and nitrous oxide (99.9%, White Martins®, Brazil) were used for flame composition for atomization of metal analysis by air-acetylene flame. Considering that Cu, Fe, Mn, Cd, Co, and Ni are used in agrochemicals in the region of Grande Dourados, these compounds were selected for this study. The parameters, such as wavelength, hollow cathode lamp current, slit setting, and gas flow rate, were adjusted under optimum conditions, and the results are shown in Table 1. All measurements were carried out in five replicates.

High-purity de-ionized water obtained using a Millipore Milli-Q Academic® deionizer system (resistivity of  $18.2\text{ M}\Omega\text{ cm}$ , Millipore, Bedford, MA, USA) and nitric acid (65% (m/v), Sigma-Aldrich®, USA) were used to prepare the analytical solutions and samples. All solutions were stored in glassware bottles, cleaned by soaking in 10% (v/v)  $\text{HNO}_3$  for at least

**Table 1** Physicochemical analysis of water samples from P1, P2, and P3 in the “Curral de Arame” stream

| Physicochemical parameters             | P1   | P2   | P3   | CONAMA <sup>a</sup> parameters     |
|--|------|------|------|------------------------------------|
| Water temperature (°C)                 | 24.5 | 24.8 | 28.6 | Less than 40 °C                    |
| pH                                     | 7.4  | 7.4  | 7.6  | 6.0 to 9.0                         |
| Dissolved oxygen (mg L <sup>-1</sup> ) | 7.4  | 5.6  | 8.7  | Not less than 4 mg L <sup>-1</sup> |

<sup>a</sup> Brazilian Resolution CONAMA no. 357/2005

24 h, and thoroughly rinsed in de-ionized water before use.

For sample pretreatment procedure, 200 mL water was transferred to an Erlenmeyer flask followed by 10.0 mL HNO<sub>3</sub>. The mixture was submitted to a heating block system at 90 °C to reduce and concentrate the water samples to 30 mL. The final volume was adjusted up to 50 mL with 1.0% (v/v) HNO<sub>3</sub> solution. All samples were prepared in triplicate.

For the main atomic lines for Cu, Fe, Mn, Cd, Co, and Ni, 5.0 mL min<sup>-1</sup> sample flow rate and multi-element calibration curves within the 0.1–2.0 mg L<sup>-1</sup> Cu, 0.5–4.0 mg L<sup>-1</sup> Fe, 0.2–4.0 mg L<sup>-1</sup> Mn, 0.2–10.0 mg L<sup>-1</sup> Cd, 0.2–10.0 mg L<sup>-1</sup> Co, and 0.2–10.0 mg L<sup>-1</sup> Ni ranges were consistently obtained. All measurements were carried out in five replicates. The limit of detection (LOD) and limit of quantification (LOQ) were calculated according to the IUPAC.

#### Determination of organic compounds

##### *Extraction of water constituents*

The water samples (200 mL) were subjected to the solid phase extraction (SPE). The cartridge was conditioned with ultrapure water and subsequently, 200 mL of the sample was eluted by SPE. The constituents that adhered to the cartridge were eluted with 20 mL methanol followed by 20 mL ethyl acetate. The methanol and ethyl acetate fractions of each sample were combined and evaporated. Subsequently, the fractions were diluted in 100 µL methanol, filtered through a 0.20-µm membrane, and analyzed by high-performance liquid chromatography (HPLC).

##### *HPLC analysis*

The samples and standards were analyzed in an analytical HPLC (Varian 210) system with a ternary solvent delivery system equipped with an auto-sampler, a

photodiode array detector (PAD). The column was a C-18 (25 cm × 4.6 mm; particle size, 5 µm; Luna, Phenomenex, Torrance, CA, USA). In each analysis, the flow rate and the injected volume were set as 1.0 mL min<sup>-1</sup> and 20 µL, respectively. All chromatographic analyses were performed at 25 °C. Considering that carbendazim is a fungicide widely used in soybean crops and thiamethoxam is an insecticide used in corn and sugarcane crops, these agrochemicals were used for analyses. Elution for carbendazim (CAS no. 10605-21-7) and thiamethoxam (CAS no. 15-3719-23-4) was carried out using acetonitrile (A) and water (B) 65:35 for 5.50 min. The solvent gradient program was set at 0 min, 35% B for 5.50 min, 40% B for 15 min, and 20% B for 18 min and returned to the initial analysis time at 20 min. Oxytetracycline is an antibiotic widely used in treatment of respiratory diseases in breeding animals. The elution of oxytetracycline (CAS no. 79-57-2) was carried out using 28% acetonitrile and 72% 0.01 mol L<sup>-1</sup> oxalic acid.

##### *Linearity and detection limits*

In the region of Grande Dourados, there is intense cultivation of soybeans, maize, and sugar cane where the compounds carbendazim and thiamethoxam are widely used for these plantations. In this area are also developed activities related to creation of animals, where antibiotics can be constantly used, among these we can highlight oxytetracycline. This way the content estimation of the standards oxytetracycline, cabendazim, and tiametoxam in the samples was performed by external calibration by HPLC, and each determination was carried out five times (1–100 µg mL<sup>-1</sup>). The detection limits were determined by injecting ( $n = 5$ ) standard solutions of known concentration and then decreasing the concentrations of the samples until peak detection with a signal/noise ratio of 3.

## Microbiological analysis

The techniques used to quantify water microorganisms were based on the *Standard Methods for the Examination of Water and Wastewater*, published by the *American Public Health Association* (APHA, AWWA, and WEF 2005).

### *Research on total and thermo-tolerant coliforms and E. coli*

For the analysis of coliforms, the most probable number (MPN) technique was used, with serial dilution of the sample and using a triplicate of three tubes of lauryl tryptose broth (Merck, Darmstadt, Germany). Later, the brilliant green bile broth (BGBB) (HiMedia Laboratories Ltd., Mumbai, India) was used for confirmation of total coliforms and the EC Broth (Isofar, Rio de Janeiro, Brazil) for confirmation of thermo-tolerant coliforms. The number of total coliforms was determined by MPN. *E. coli* was isolated in EC broth on EMB (eosin methylene blue agar) agar plates (HiMedia Laboratories Ltd., Mumbai, India). The colonies with a metallic green sheen underwent biochemical identification (APHA, AWWA, and WEF 2005; Silva et al. 2010).

### *Research on Salmonella spp.*

A buffered peptone water pre-enrichment (HiMedia Laboratories Ltd.) and the selective enrichment in selenite cystine broth (SC) (Isofar) and Rappaport Vassiliadis broth (RV) (Isofar) were used for the detection of *Salmonella* spp. In addition, for the isolation of the microorganism, xylose-lysine-desoxycholate agar (XLD) (Isofar) was used (APHA, AWWA, and WEF 2005; Silva et al. 2010). The colonies with a transparent halo and a central black spot were selected by the biochemical methods TSI (triple sugar iron agar), MIO (motility, indole, and ornithine), and urea.

### *Research on Pseudomonas spp.*

For the analysis of *Pseudomonas* spp., 10 mL of the water sample was enriched in buffered peptone water (HiMedia Laboratories Ltd.) and, after 24 h, inoculated by using a discontinuous streak technique, in duplicate, using selective cetrimide agar (Merck) plates according to the methodology proposed by Oliveira et al. (2013). The plates with bacteriological growth were considered

positive for *Pseudomonas* spp. (APHA, AWWA, and WEF 2005; Silva et al. 2010).

### *Extraction of chlorophyll a*

Chlorophyll *a* was extracted from algae belonging to the periphytic community. The experiment was carried out in three plastic slides of polyethylene terephthalate (PET) having a size of  $2.5 \times 7.5$  cm, with approximately  $18.75 \text{ cm}^2$ , and an artificial substrate for the development of the periphytic community in it. The slides have been set into the stream 15 days before each collection and submerged at a 30 cm water depth. After this period, the slides were removed from the stream and placed in 50 mL tubes filled with stream water. The slides were then stored in a cooler with ice until the analysis began. The procedures followed the method described by Siqueira and Rodrigues (2009).

## Biological analysis

### *Animal bioassays*

The water samples from points P1, P2, and P3 were placed in aerated glass tanks ( $40 \times 30 \times 20$  cm) at room temperature ( $23^\circ\text{C}$ ) for 24 h. After this period, ten fishes (*A. lacustris*) from a commercial fish farm were placed in these tanks for 72 h, five were used for micronucleus test, and another five for Comet Assay.

The procedures for conducting the animal experiments were approved by the ethics committee on animal research of the UFGD, protocol n° 005/2013.

### *Micronucleus piscine test on A. lacustris*

Specimens of *A. lacustris* were removed from the tanks after a 72-h treatment and anesthetized with benzocaine 2% (soluble). Then, a cut was made in the caudal fin to collect blood for blood extension.

The micronucleus count of erythrocytes followed the protocol described by Schmid (1975) and Heddle et al. (1983). The blood was collected in the caudal region, and a peripheral blood smear was made on the surface of the slide (two for each fish). The slides were fixed in ethanol and stained with Rapid Panoptic LB (Laborclin, Pinhais, Brazil). A total of 2000 cells was counted for each fish. Only red blood cells with intact cell and cytoplasmic membrane were considered. The counting of micronuclei was made on a Nikon optical microscope



(400×), and the frequency of micronuclei was calculated (total number of cells with micronucleus / total number of cells observed × 100).

#### *Comet Assay on A. lacustris*

The Comet Assay was adapted from the methodology of Ramsdorf et al. (2009). Six microliters of blood was collected by gill puncture and diluted in 2000 µL of saline solution (PBS). Two slides from each fish were made of 20 µL of cell suspension and 120 µL of low melting point 0.5% agarose (*v/v*) at 37 °C. The slides remained in a lysis solution at 4 °C for 1 h. After lysis, the slides were stored on a 0.3 mol L<sup>-1</sup> NaOH buffer and 0.001 mol L<sup>-1</sup> EDTA (pH > 13) for 20 min and subjected to electrophoresis at 25 V, 300 mA, for 20 min. Then, the slides were neutralized with 0.4 mol L<sup>-1</sup> Tris for 15 min, fixed in ethanol, and stained with 0.02 mol L<sup>-1</sup> ethidium bromide. A total of 100 nucleoids from each slide were observed with a fluorescence microscope (Labomed—T121100) with a 400× lens.

The nucleoids were classified according to the size of the “tail” as Class 0 (no damage), Class 1 (slightly damaged), Class 2 (intermediate damage), Class 3 (damaged), and Class 4 (high damage). Slide analysis was always performed by the same technician. For DNA damage evaluation, cell score (CS) was calculated using the following formula: (percentage of cells in Class 0 × 0) + (percentage of cells in Class 1 × 1) + (percentage of cells in Class 2 × 2) + (percentage of cells in Class 3 × 3) + (percentage of cells in Class 4 × 4).

#### Statistical analysis

Data on extraction of chlorophyll *a* and the frequency of micronuclei (MCN) and cell score (CS), from the animal bioassays, were subjected to analysis of variance and then to Duncan’s test at a 0.05 probability to compare the averages using the BioEstat 4.0 program (Ayres et al. 2005).

## Results and discussion

#### Physicochemical analysis

The values for physicochemical analysis were within the allowable values according to the National

Environmental Council (CONAMA) (357/2005) for P1, P2, and P3 (Table 1).

#### Metal detection

The concentrations of Cd, Co, and Ni were above the permissible limits at all collection points (P1, P2, and P3) according to the CONAMA (357/2005) (Table 2), which places the stream into a Class 3 water classification (intended to supply for human consumption after conventional or advanced treatment; irrigation of trees, cereals, and fodder crops, amateur fishing, secondary contact recreation, and livestock watering). In addition, these metals could affect aquatic biodiversity (Ribeiro et al. 2012).

Bermudez et al. (2009) reported that human activities, including industry, mining, and agriculture, can be cited as sources of pollution for the incorporation of metals in the environment. We can mention that the main sources of metal contamination in the “Curral de Arame” stream are related to agricultural activities by percolation of waste from insecticides, herbicides, fungicides, and fertilizers, considering that these metals are present in these products.

The copper found were lesser than the quantification limit of procedure (0.026 mg L<sup>-1</sup>) but higher than the CONAMA parameter (0.013 mg L<sup>-1</sup>). This fact suggested that the concentration of Cu could be higher than the parameter allowed, considering that this is a micronutrient used in chemical fertilization.

P3 showed a higher concentration of metals, and among the metals tested, both Cd (0.510 mg L<sup>-1</sup>) and Ni (0.316 mg L<sup>-1</sup>) were above the limits allowed by the Brazilian legislation (Ministerio do Meio Ambiente—CONAMA 2005). This may be related to the dragging power of water near the mouth of the stream and receiving interference from agrochemicals used in agriculture.

Bayen (2012) reported the difficulty to identify the source of metal pollution into the water, because it can be a combination of natural and anthropogenic sources, diffuse or occasional. High level of copper, cadmium, nickel, and other metals was found in the spots; the detection of lead, arsenic, cadmium, nickel, and manganese is related to agricultural activities, as these elements are present in agricultural supplies (Schipper et al. 2008; Rocha et al. 2015).

**Table 2** Metal concentrations (mean and standard deviation (SD), mg sL<sup>-1</sup>) in water from the “Curral de Arame” stream

|                    | Elements (mg L <sup>-1</sup> ) |               |                |               |               |               |
|--------------------|--------------------------------|---------------|----------------|---------------|---------------|---------------|
|                    | Copper (Cu)                    | Iron (Fe)     | Manganese (Mn) | Cadmium (Cd)  | Cobalt (Co)   | Nickel (Ni)   |
| P1                 | <LQ                            | 0.651 ± 0.018 | 0.026 ± 0.002  | 0.071 ± 0.003 | 0.280 ± 0.024 | 0.236 ± 0.009 |
| P2                 | <LQ                            | 1.464 ± 0.041 | 0.092 ± 0.004  | 0.085 ± 0.004 | 0.405 ± 0.008 | 0.263 ± 0.016 |
| P3                 | <LQ                            | 1.610 ± 0.044 | 0.115 ± 0.010  | 0.510 ± 0.002 | 0.303 ± 0.021 | 0.316 ± 0.011 |
| CONAMA* parameters | 0.013                          | 5.0           | 0.5            | 0.01          | 0.2           | 0.025         |

LQ limit of quantification (mg L<sup>-1</sup>): Cu 0.026, Fe 0.049, Mn 0.017, Cd 0.029, Co 0.073, and Ni 0.062. LOD limit of detection (mg L<sup>-1</sup>): Cu 0.008, Fe 0.015, Mn 0.005, Cd 0.008, Co 0.021, and Ni 0.019. \*Brazilian Resolution CONAMA no. 357/2005

Determination of organic compounds

The antibiotic oxytetracycline was found in greater concentration in P3 (89 µg L<sup>-1</sup>), followed by P2 (53 µg L<sup>-1</sup>) and P1 (51 µg L<sup>-1</sup>). The drug detection analysis in natural environments is used as a marker of human activities, since these drugs are found in aquatic environments due to the action of man (Kasprzyk-Hordern et al. 2009).

Research conducted in freshwater environments in Spain (Vazquez-Roig et al. 2011; Andreu et al. 2016) evaluated the presence of antibiotics (ciprofloxacin, norfloxacin, ofloxacin, oxytetracycline, sulfamethoxazole, tetracycline, and trimethoprim), but oxytetracycline has not been quantified. However, in the present study, this drug was detected in all collection spots. Such contamination may be related to the contamination of urban waste water, operations in animal feeding, industrial waste effluents, and uncontrolled landfills.

Microbiological analysis

*Total coliforms and E. coli*

The highest values found for total coliforms occurred in the August collection at P2 (240 NMP/100 mL), May at P3 (460 NMP/100 mL), and November at P3 (460 NMP/100 mL) (Fig. 2). According to the CONAMA’s Resolution 020/1986 (Ministério do Meio Ambiente—CONAMA 1986) and WHO—World Health Organization 2006), such water could be directed to the irrigation of crops as it presented values ≤ 1000 total coliforms per 100 mL of water, which is the amount allowed for this activity.

Even with lower values for total coliforms in P1 and P2 in both studies, *E. coli* was reported in stream water. P1 was located near the native bushland (a refuge for

several species of animals), allowing direct contact of wild animals with the watercourse, and P2 is the nearest from the urban area (Gomes et al. 2007; Mussury et al. 2008).

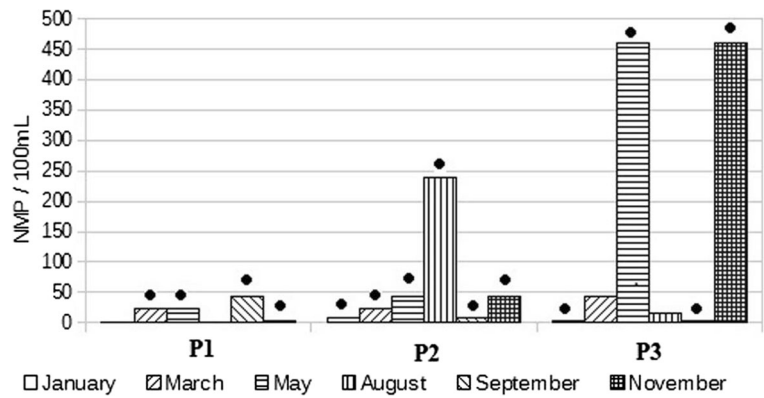
P3, near the mouth of the “Curral de Arame” stream, showed the highest values for total coliforms and the lowest number of *E. coli* isolated per collection. Studies reported that the collection sites near the mouth have high concentrations of total coliforms, possibly due to the water draining (Poma et al. 2012; Sassoma et al. 2015).

Lee et al. (2014) have studied the fecal human and animal contamination of a watercourse by means of molecular markers. In the present study, they found that such tool, by itself, did not generate enough information to precisely identify the contamination of fecal origin. In order to obtain more reliable results regarding the source of pollution, the markers and rainfall values were assessed together.

The results of total coliforms and rainfall indexes in the collection periods presented a directly proportional relationship. The month when the higher rainfall index occurred, May, was also the month that showed higher values for coliforms (total and thermo-tolerant) (Fig. 3). The month of January showed lower values for coliforms (total and thermo-tolerant) and also a higher rainfall index. According to Gentry-Shields et al. (2012), the effects of rain were studied as a tool to assist fecal contamination detection in river basins of the urban interior, because rain, while flowing into the watercourse, carries along all the contamination present on the streets of the cities and on soils.

The oxytetracycline may also have influenced the proliferation of *E. coli*, because P3 was the one that presented the lowest number of isolated *E. coli* and the highest concentration of antibiotics (Sato et al. 2015).

**Fig. 2** Total coliforms (NMP/100 mL) of the six samples taken from the “Curral de Arame” stream, P1, P2, and P3. (●) represents *E. coli* found at the collection point



Research on *Salmonella* spp.

No *Salmonella* spp. was detected in any of the samples. Several authors studied the presence of *Salmonella* spp. in natural aquatic environments, as it causes various diseases such as gastroenteritis, bacteremia, and typhoid fever (Abakpa et al. 2015; Masarikova et al. 2016).

Research of *Pseudomonas* spp.

In all collection points in the months of March, May, August, and November, *Pseudomonas* spp. was identified in the water samples of the “Curral de Arame” stream. However, in September, *Pseudomonas* spp. was observed only in P3, and, in January, there were no *Pseudomonas* spp. in all the analyzed points.

The proliferation of *Pseudomonas* spp. in natural environments has become a problem to human health, since it is responsible for opportunistic and nosocomial infections. This is an alarming fact because *Pseudomonas* spp. presents an intrinsic and acquired resistance due to overuse of

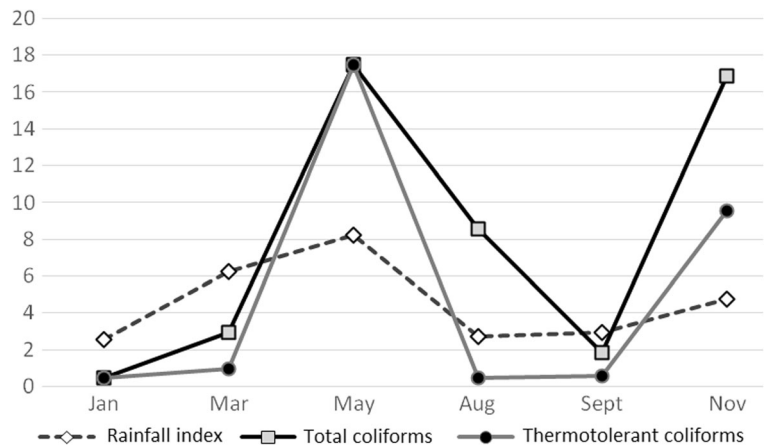
antibiotics (Master et al. 2011). Studies have isolated *Pseudomonas* spp. in aquatic freshwater environments and described the concern with multidrug resistance to antibiotics as exhibited by microorganisms. *Pseudomonas* spp. presents a high metabolic versatility and can adapt to different conditions and different habitats (Suzuki et al. 2013; Rocha et al. 2015).

Extraction of chlorophyll *a*

The month of March showed a higher average of chlorophyll *a* (0.266), differing from the months of September and May when the lowest average (0.081) was verified. The collection points presented no different values among them as to the average of chlorophyll *a* (Table 3).

The results of the present study corroborate the findings of Ortiz et al. (2008) who have identified an inversely proportional relationship between the number of nutrients and dissolved oxygen. We observed that the point with the highest chlorophyll *a* (P1) had the lowest dissolved oxygen

**Fig. 3** Relationship between rainfall index (mm) and coliforms (total and thermo-tolerant) in months of collection at P1, P2, and P3





**Table 3** Values of chlorophyll *a* (mg cm<sup>-2</sup>) of the “Curral de Arame” stream expressed on average per month and averages per collection points and standard deviation (SD)

| Values of chlorophyll <i>a</i> (mg cm <sup>-2</sup> ) | Months            | Averages           |
|---|-------------------|--------------------|
|   | March             | 0.2669 ± 0.0425 A  |
|   | May               | 0.0818 ± 0.0276 C  |
|   | September         | 0.1522 ± 0.1360 BC |
|   | November          | 0.2139 ± 0.1085 AB |
| Points  | Averages          |                    |
| P1  | 0.1818 ± 0.1078 A |                    |
| P2  | 0.2065 ± 0.1284 A |                    |
| P3  | 0.1555 ± 0.1024 A |                    |

Averages followed by the same letter in columns are statistically equal according to Duncan’s test at a 0.05 probability

content (Table 1) and the point with the lowest chlorophyll *a* (P3) had the highest dissolved oxygen index when compared to the other points (Table 1).

Biological analysis

*Micronucleus*

The results for the micronucleus test (MCN) were presented in Table 4. In January and May, the biggest number of MCN in erythrocytes of *A. lacustris* was observed in P2 (12.0 and 15.2, respectively) and the lowest in P3 (1.800 and 7.600, respectively). In addition, in August and November, P3 showed the highest number of MCN (17.0 and 15.0, respectively) and P1 the lowest (6.2 and 8.0, respectively), being different only in August. On the other hand,

in September (14.3) and August (13.0), greater numbers of micronuclei could be found.

Comparing the collection points, the highest frequency of MCN in P1 was in September (12.0) and the lowest in August (6.2), being different between them. In P2, the largest number of MCN happened in August (15.8); however, there was no difference between the months. The highest number of MCN, in P3, was found in August (17.0), statistically equal the frequency observed in the months of September and November and different from January and May, with lower micronuclei frequencies. In general, the highest average of micronuclei occurred in P2 (14.2), followed by P3 (11.4) and P1 (8.5), being different among them.

Some authors (Duarte et al. 2012; Ghisi and Oliveira 2013) suggested that the induction of micronuclei is related

**Table 4** Frequency of micronuclei (MCN), cell score (CS), and standard deviation (SD) in P1, P2, and P3 spots of the “Curral de Arame” stream in Dourados, MS, Brazil along six months of collection in 2013

| Points         | MCN             |                 |                  |                 |                 |                  |                 | Averages/month |
|----------------|-----------------|-----------------|------------------|-----------------|-----------------|------------------|-----------------|----------------|
|                | Jan             | Mar             | May              | Aug             | Sept            | Nov              |                 |                |
| P1             | 8.2 ± 4.1 abB   | –               | 8.2 ± 6.3 abA    | 6.2 ± 3.3 bB    | 12.0 ± 1.5 aA   | 8.0 ± 1.8 abB    | 8.5 ± 4.0 C     |                |
| P2             | 12.0 ± 3.7 aA   | –               | 15.2 ± 5.4 aA    | 15.8 ± 7.1 aA   | 15.4 ± 6.0 aA   | 12.8 ± 5.1 Aab   | 14.2 ± 5.3 A    |                |
| P3             | 1.8 ± 1.9 bC    | –               | 7.6 ± 3.5 bA     | 17.0 ± 4.6 aA   | 15.4 ± 6.5 aA   | 15.0 ± 5.4 aA    | 11.4 ± 7.2 B    |                |
| Averages/month | 7.3 ± 5.3 c     | –               | 10.3 ± 6.0 bc    | 13.0 ± 7.0 ab   | 14.3 ± 5.1 a    | 11.9 ± 5.1 ab    | –               |                |
| Points         | CS              |                 |                  |                 |                 |                  |                 | Averages/month |
|                | Jan             | Mar             | May              | Aug             | Sept            | Nov              |                 |                |
| P1             | 78.6 ± 31.1 dB  | 99.6 ± 9.7 cdB  | 112.6 ± 12.3 cA  | 108.8 ± 13.3 cA | 158.6 ± 31.1 bA | 189.0 ± 16.0 aA  | 111.6 ± 33.3 A  |                |
| P2             | 59.4 ± 38.5 cB  | 121.5 ± 29.7 bB | 96.8 ± 36.7 bcA  | 111.8 ± 28.3 bA | 165.2 ± 30.3 aA | 180.2 ± 17.8 aAB | 122.5 ± 50.0 A  |                |
| P3             | 114.8 ± 1.6 bcA | 170.6 ± 27.3 aA | 113.6 ± 33.6 bcA | 83.6 ± 14.2 cA  | 118.2 ± 47.9 bA | 157.6 ± 39.4 aB  | 126.4 ± 41.26 A |                |
| Averages/month | 84.3 ± 35.6 d   | 130.6 ± 37.9 b  | 107.7 ± 28.5 c   | 101.4 ± 22.5 cd | 147.3 ± 40.7 b  | 168.9 ± 31.2 a   | –               |                |

Averages followed by the same lowercase letter in rows and uppercase letter in columns are statistically equal according to Duncan’s test at a 0.05 probability

to the presence of chemical compounds in the water and that their presence causes changes in the genetic material of the cell. Gutierrez et al. (2015) observed that anthropogenic compounds affected the health of different species of fish, causing damage to the genetic material of the cell and inducing the appearance of micronuclei, mainly in urban and industrial areas. This fact corroborates the results found, as in places (P2 and P3) where the number of MCNs was higher. A higher quantity of metals and oxy-tetracycline, which favors the use of this technique to evaluate the action of chemical agents in the cell, was also found.

In the months of August and September, a greater amount of MCNs and less rainfall were observed (Fig. 4). Such result indicates the influence of rainfall on the genetic damage that may be due to greater concentration of pollutants in water during periods of low rainfall. This result corroborates the studies by Pavlica et al. (2008) which demonstrated the influence of seasonal variations and the interaction with the contaminants with the amount of MCNs in the cells of mussels, especially in winter. In this way, the appearance of MCNs in erythrocytes of *A. lacustris* exposed to the water of the “Curral de Arame” stream may be related to both the presence of chemical compounds and the variation in climatic factors and the interactions between them. There was interaction between the points and time for the micronucleus variable.

#### Comet assay

#### Cell score

The results related to CS were presented in Table 4. The highest damage to the DNA was found in P3 in January

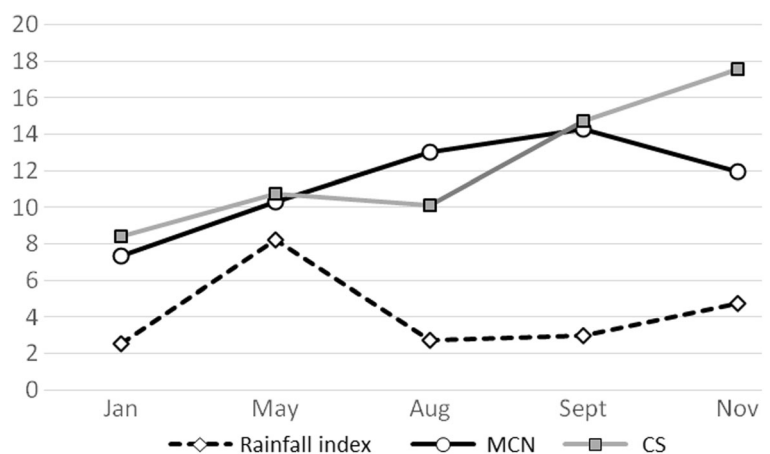
(114.8) and March (170.6) and P2 in August and September (111.8 and 165.2, respectively). In May and November, no difference was observed between those points. The month that showed the highest number of cells with DNA damage was September (147.3), and the lowest number was registered in January (84.3).

P1 and P2 presented a higher number of damaged cells in the month of November (189.0 and 180.2, respectively) and a lower number in January (78.6 and 59.4, respectively). In P3, the highest number of damaged cells was observed in the month of March (170.6) and the lowest in August (83.6).

Changes to DNA observed in erythrocytes of *A. lacustris* may be related to the existence of genotoxic agents found in the water of the “Curral de Arame” stream (metals and antibiotic). According to Carita and Marin-Morales (2008), many of these compounds produce changes in the structure and function of the DNA, which brings on mutations, affects their integrity, and causes breaks and losses of genetic material. Fatima et al. (2014) identified that chronic exposure to metals (Cr, Pb, Ni, and Zn) in different species of fish originated damages to the genetic material, increased DNA degradation, and caused damages to fish health. As such, those results corroborate with the ones achieved in the present study and indicate that metals may cause impacts to aquatic biota due to bioaccumulation in fish and can affect human beings who are the final consumers in the food chain.

Another factor that influences the concentration of chemical compounds in aquatic environments is seasonal variation. Rainfall data, when compared with the Comet Assay, indicated that the months when there was a higher DNA damage in exposed fish (September and November) were those when the amount of rainfall was scarce (Fig. 4).

**Fig. 4** Comparison of rainfall index (mm) with the number of micronuclei (MCN) and cell score (CS) according to months



This fact suggests that this parameter is related to the increase of genotoxicity in fish. Scalón et al. (2010) observed greater genotoxic damage in the cells of *Hypheossobrycon luetkenii* fish during Spring using the Comet Assay technique. Their results indicate the possible influence of seasonality in the concentration of chemical compounds; however, this is not the only influent factor in this respect. As such, these results highlight the influences of organic compounds on the potential induction of DNA damage to the exposed organisms, especially in periods with low rainfall, showing the influence of urban and agricultural activities on the water quality of the “Curral do Arame” stream.

## Conclusion

The tests used for stream monitoring were efficient to establish the anthropic interference that water receives. P3 presented a higher contamination (high levels of metals, higher concentrations of total coliforms and chlorophyll *a*, and major genetic changes). This location is the closest to the mouth of the river and all contamination (agricultural and urban) received in this path was dragged by the water. According to the CONAMA (357/2005), stream water could be directed to some human activities (to supply for human consumption after conventional or advanced treatment; irrigation of trees, cereals, and fodder crops, amateur fishing, secondary contact recreation, and watering of animals). However, the results proved the impossibility of using this water for such activities, through direct contact with the water or by fish consumption. The population is then subjected to contamination by bacteria or genotoxic agents.

**Acknowledgments** The authors thank Dr. Jorge Luiz Raposo Junior for carrying out the analysis of metals, Dr. Yzel Rondon Suárez for lending the equipment Hanna HI 9829 Multiparameter, the financial help of the Foundation for the Support to Development of Education, Science and Technology of the State of Mato Grosso do Sul—FUNDECT, and the National Scientific and Technological Development Center—CNPq for the master’s scholarships.

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