



Fluorinated waste and firefighting activities: biodegradation of hydrocarbons from petrochemical refinery soil co-contaminated with halogenated foams

Renato Nallin Montagnolli¹ · Paulo Renato Matos Lopes² · Ederio Dino Bidoia¹

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Abstract

Perfluorinated compounds, including fluorotelomers, are important constituents of firefighting foams to extinguish fuel fires in the petrochemical industry, airports, and at fire-training sites. In this study, we monitored the biodegradation process in a co-contamination scenario with monoaromatic hydrocarbons commonly found in fuels (benzene, toluene) and fluorotelomers. The CO₂ production rates were evaluated by a factorial design taking into account the effect of seasonality at in situ natural attenuation processes. Headspace analysis by gas chromatography with a thermal conductivity detector (GC-TCD) was applied to detect CO₂ production, whereas monoaromatics were analyzed by gas chromatography coupled to mass spectrometry (GC-MS). According to our results, seasonality had a detectable effect during summer, yielding different CO₂ production rates. Higher temperatures increased CO₂ production rate, while higher concentrations of fluorotelomer inhibited the biodegradation process. On average, benzene and toluene were depleted 17.5 days earlier in control assays without fluorotelomers. Toluene removal efficiency was also notably higher than benzene. The noticeable decrease in degradation rates of monoaromatics was caused by perfluorinated compounds that are possibly linked to metabolic inhibition mechanisms. Fluorotelomer diminished catabolism in all of our batch cultures. In addition to this, an alternative production of by-products could be detected. Thus, we propose that transient components of the benzene and toluene degradation may be differentially formed, causing the benzene, toluene, and perfluorinated co-contaminations to go through switched metabolic stages under the presence of fluoride in a contamination scenario.

Keywords Perfluorinated compounds · Fluorotelomer · Bioremediation · Chromatography · Halogenated waste · Bioremediation · Benzene · Toluene

Introduction

Petroleum derivatives can be hazardous contaminants, as many of its compounds persist in the environment for an extended amount of time. Fires and explosions are another, more severe, way that petroleum can be introduced to ecosystems. When fossil fuels reach the ground, most of their components are split into

three phases: dissolved, liquid, and gaseous. A minor fraction of this mixture is prone to dissolving in groundwater, whereas most of the portion is retained in the porous spaces between soil particles in its pure liquefied form (also known as residual saturation). A third portion evaporates, leading to atmospheric contamination (Nadim et al. 1999; Freije 2015; Tobiszewski and Namieśnik 2015; Mishra et al. 2016; Cholakov 2016; Zhuo et al. 2017). The petroleum contamination is relevant due to the potential to contaminate aquifers and springs that often supply urban and industrial demand (Qin et al. 2009; Choi and Lee 2011; Montagnolli et al. 2014; Postigo et al. 2017).

Many references (NFPA 1991; Chettouh et al. 2016; Zhang et al. 2016; Runefors et al. 2017) state that fires are prone to occur in both confined areas (such as fuel storage tanks at refineries or gas stations) and in wide-open facilities (such as wells and oil spills). Therefore, accidents are potential causes of catastrophic combustion of fuels. In this context, water is

Responsible editor: Robert Duran

✉ Ederio Dino Bidoia
ederio@rc.unesp.br

¹ Department of Biochemistry and Microbiology, Biosciences Institute, Sao Paulo State University (UNESP), Rio Claro, São Paulo, Brazil

² College of Agricultural and Technological Sciences, São Paulo State University (UNESP), Dracena, São Paulo, Brazil

not the most appropriate substance to suppress fires, due to its ineffective extinguishing properties. Firefighting foams are the proper extinguishing resource. They are better flame retardants when hydrocarbons ignite. However, these foams also combine with soil particles on long-term, causing a mixed contamination situation along with the remaining non-ignited hydrocarbon (Nolan 2011; Borg and Njå 2013). Therefore, petroleum pollution can be intensified by such firefighting foams, due to their formulation, which is mostly based on fluorinated substances.

Emergency situations related to firefighting are often encountered in the petroleum industry, as well as airports and even fire training sites at military bases. These possibly disastrous scenarios are all solved using aqueous film-forming foams, which are substances specifically designed to extinguish hydrocarbon fires. The main components of such foams are long carbon chains containing many strong C-F bonds (perfluorinated chains) that act by reducing surface tension (Dauchy et al. 2017). Fluorotelomers are the active substance within firefighting foam formulations that cause improved spreading and fire suppression (Donaldson 2016). Certainly, firefighting occurs in many scenarios, but they are all linked to other petroleum-based substances such as diesel, kerosene, and other fuels, hence causing a much more complex co-contamination in the environment (Vecitis et al. 2009; Hinnant et al. 2017).

The co-contamination of hydrocarbons with firefighting foams has a considerable impact on biodegradation rates. Petroleum products, such as gasoline, also contain toxic monoaromatic compounds (e.g., benzene and toluene) as reported by Wang et al. (2016), Fernández et al. (2016), Meyer et al. (2014), and Oberoi and Philip (2017). Among hydrocarbon found in contaminated sites, monoaromatic hydrocarbons are often found as co-contaminants in petrochemical refinery and storage areas. Gasoline, for example, has benzene and toluene on 30% of its formulation. Opportunely, these compounds are also known for their potential biodegradability using aerobic microbial communities (Mathur and Majumder 2010). Aromatic hydrocarbons can be dangerous if a major exposition to humans in these environments occurs, whether in chronic or accidental releases (Hamed et al. 2003).

Recently, the detection of fluoride-based compounds has also been linked to the susceptibility of petroleum-contaminated sites to toxicological effects (Laitinen et al. 2014; Panneerselvam et al. 2016). In this context, co-contamination poses a threat for any remediation strategy of contaminated groundwater containing both hydrocarbons from refining processes and firefighting foams. Fluorinated compounds have a major role in promoting an even greater toxicity of hydrocarbons in soil. Their surfactant properties also increase the risk of spreadability of contamination plumes over time.

Bioremediation techniques are cost-effective approaches that can be applied to such contamination scenarios, by using biological agents to remove hazardous residue. Microorganism-based

strategies result in the mineralization of the pollutant via microbial biodegradation processes, yielding carbon dioxide and water (Alexander 1975; Kolvenbach et al. 2014; Khatoon et al. 2017; Kapellos 2017). The biomass can either be solely composed of indigenous microorganisms from the site itself or mixed with enhanced cultures. Bioremediation aims to exploit many microbial-specific oxidation-reduction reactions, in which hydrocarbon is optimally oxidized and the most appropriated electron acceptors are selected (Martínková et al. 2009; McGenity 2014; Ghattas et al. 2017). Under the appropriate environmental conditions, benzene and toluene compounds can be co-remediated in the presence of perfluorinated compounds, thus simultaneously outputting maximum defluorination and hydrocarbon mineralization.

It is unknown if perfluorinated-adapted microbial populations are active and fully able to diverge the biotransformation process of monoaromatic hydrocarbons. A complete knowledge of metabolic pathways of hydrocarbon-degrading and fluorotelomer-transforming microorganisms would allow the design of effective remediation systems specifically aimed towards both hydrocarbon mineralization and defluorination at foam-impacted sites.

In this context, our research simulated a co-contamination scenario with perfluorinated compounds, benzene, and toluene. We aimed to measure the biodegradation process in a co-contamination scenario with monoaromatic hydrocarbons commonly found in fuels (benzene, toluene) and fluorotelomers. The CO₂ production rates were evaluated by a factorial design taking into account the effect of seasonality at in situ natural attenuation processes. Furthermore, aromatic hydrocarbon concentration was simultaneously monitored at a set temperature to determine the biodegradation rate of benzene and toluene in such co-contamination environment. After the biodegradation process of both benzene and toluene substrates was over, we searched for individual metabolites from the biodegradation process. Our approach complementarily provided new insights towards the behavior of firefighting foams when released in a soil matrix.

Methods

The biodegradation was monitored using headspace analysis by gas chromatography with a thermal conductivity detector (GC-TCD) to detect CO₂ production, whereas aromatics were analyzed by gas chromatography coupled to mass spectrometry (GC-MS). With such methodologies, we inferred how benzene and toluene biodegradation was influenced by the perfluorinated substances from firefighting foams.

Sample collection

Soil samples were collected in 2017 from Replan Petrobras oil refinery in Paulinia, Brazil (22° 43' 24.2" S 47° 08' 00.3" W), in an area close to fuel tanks. The soil was collected at around

0.9 m subsurface depth. All samples were stored in non-fluoride-based containers at $-80\text{ }^{\circ}\text{C}$ before the assay setup. This soil sample has been chosen due to the local microbiota adaptations to biodegrade hydrocarbons. Besides, this soil has a history of the occasional release of perfluorinated compounds during firefighting training. Over 20 years ago, approximately 36,800 L of firefighting foam was released into this soil, when firefighters struggled to extinguish more than 5,000,000 gal of burning diesel and gasoline. Traces of perfluorinated substances are still detected in this region nowadays (Figueredo and Sabadini 2003).

Serum bottle assays

The experiments were set up using 100-mL serum bottles. Each glass bottle contained a simulated microenvironment with soil microbiota and contaminants. In order to better evaluate environmental parameters towards benzene and toluene biodegradation, the experimental design followed a function of factors. Two factors or parameters such as temperature and perfluorinated compound concentration were chosen as the variables at three defined levels diverging from a central point ($-1, 0, +1$). Degradation rates were determined according to CO_2 production in the respirometric assays. The simulated seasonality was achieved by storing our assays at different temperatures according to Table 1.

All assays were assembled in a sterile environment. The temperature range (20 to $35\text{ }^{\circ}\text{C}$) was chosen due to the temperature

history from winter to summer from where the soil samples were collected. Both fluorotelomer and aromatic concentrations were determined from the average concentration of such contaminants in polluted sites. All these assays were monitored using our respirometric method with GC-TCD. However, the mid-range temperature assays were the only ones selected for a thorough analysis of benzene and toluene consumption and biodegradation metabolite evaluation using GC-MS.

The mid-range assay quantities are specified in Table 2. The proposed soil-free medium controls (C_M) assessed the potential impacts of media components onto fluorinated compounds. Another set of sterile controls (C_N) was prepared by three consecutive soil autoclaving procedures followed by overnight freezing at $-20\text{ }^{\circ}\text{C}$. Full assays (BT+) and fluoride-free assays (BT-) were amended with BT. The assays were not supplemented with antibiotic solution, and our sterile conditions were ensured with the C_N only. These controls were all established to verify any losses of the compounds due to sampling procedures or inadequate sealing.

Each assay from Table 1 incubated at the lowest and highest temperature assays also had their respective controls (C_M , C_F , and C_N), but they were omitted in Table 2 to avoid redundancy. Ethyl-benzene and xylene, which are also monoaromatic hydrocarbons as well as components of BTEX mixtures (benzene, toluene, ethyl-benzene, and xylenes), were not used in our experimental setup. This deliberate design choice was based on pilot data containing all BTEX (benzene, toluene, ethyl-benzene, and xylene) components, where the consumption rates of ethyl-benzene and xylenes were all proportional. Besides, we aimed to simplify our approach, since the combined metabolic pathways involved in the degradation of such compounds can lead to even more variables that could induce erroneous conclusions. Therefore, benzene and toluene were selected as good indicators towards the whole mixture biodegradability. We will further refer to the benzene and toluene mixture as BT in this paper.

Our experiments were designed to isolate the effects of perfluorinated compounds; however, the full formulation of firefighting foams contains many other substances that could act as substrates that deviate the microbial metabolism from the targeted biodegradation process. Such substances include surfactants, diethylene glycol butyl ether, preservatives, and other non-fluorinated compounds. Along with diethylene glycol butyl ether, these substances are most of firefighting foam formulation. It is important to notice that diethylene glycol butyl ether contributes to over 80% of the total organic carbon of foams. Therefore, we used only fluorotelomers in our simulated contamination, as they were provided separately from the mixture by the local firefighting foam manufacturer. The perfluorinated portion from the commercially available 6% Sintex AFFF formulation (S1371/11, manufactured in Vinhedo, SP, Brazil) contained fluorotelomer thioether amido sulfonates (FtTAoS) ranging from 4 to 8 fluoride-bound carbon chains.

Table 1 Experimental design with temperature, benzene, toluene, and fluorotelomer concentration factors

Assay number	Factor A: Temperature ($^{\circ}\text{C}$)	Factor B: Benzene and toluene (μL)	Factor C: Fluorotelomer (μL)
1	35.0	15.0	10.0
2	27.5	5.0	70.0
3	27.5	15.0	30.0
4	20.0	35.0	30.0
5	27.5	35.0	30.0
6	27.5	35.0	10.0
7	20.0	15.0	10.0
8	20.0	5.0	30.0
9	35.0	5.0	30.0
10	35.0	35.0	30.0
11	27.5	15.0	30.0
12	27.5	15.0	30.0
13	20.0	15.0	70.0
14	35.0	15.0	70.0
15	27.5	5.0	10.0

Table 2 Biodegradation assay contents

Assay ID	Description	Media (mL)	Soil (g)	Fluorotelomer (μL)	Benzene (μL)	Toluene (μL)
C _M	Media control	25	2.5	–	–	–
C _F	Fluoride control	25	2.5	30	–	–
C _N	Sterile control	25	2.5*	30	15	15
BT–	No fluoride	25	2.5	–	15	15
BT+	Full assay	25	2.5	30	15	15

*Autoclaved soil

The minimum saline media used in our experiment proposed by Bushnell Haas (BH media) were composed of 0.2 g L⁻¹ MgSO₄, 0.02 g L⁻¹ CaCl₂, 1.0 g L⁻¹ KH₂PO₄, 1.0 g L⁻¹ K₂HPO₄, 1.0 g L⁻¹ NH₄NO₃, and 0.05 g L⁻¹ FeCl₃. The minimum media ensured no other substrate besides the BT and fluorotelomer would interfere with the biodegradation process. The soil itself carried organic matter into our experiments; however, it was part of our experimental design aiming to simulate a soil-microenvironment.

All vials were sealed using fluoride-free rubber stoppers. This issue was addressed to avoid any other halogenated source during experiments. Each cap was checked for their capacity of BT volatilization prevention. The materials were verified by pilot assays and controls, showing no adsorption of volatile compounds at the maximum expected duration of the experiments. Moreover, the BT mixture, with 15.0 mg L⁻¹ of each compound, had its final concentration calculated according to Henry's law. We took into account the volatility of the BT compounds and left them to be stabilized for 24 h before injection into each vial.

The experiment lasted for 50 days, during which all vials had their headspace aliquoted for GC analysis. Simultaneously with each sampling procedure, O₂ was replenished to sustain aerobic conditions. All experiments (Table 1) were run in triplicate and shaken at 120 rpm for each temperature in separate incubators. The headspace was sampled every 5 days with sterile syringes and immediately analyzed. No other co-factors were added to the culture media.

Respirometry using GC-TCD

Respiration activity from the soil microcosm was monitored from headspace samples at an Agilent 6850B model GC-TCD. This equipment was operated with helium as carrier gas set to a 46.7 mL min⁻¹ flow. The injection temperature was 280 °C. The oven temperature was 60 °C (isothermally operated) and the detector temperature was set to 250 °C. A HP-Plot capillary column (30 m × 530 mm × 40 mm) column was used in the system. The volumes of CO₂ produced were normalized to standard temperature and pressure conditions (273 K and 101.3 kPa).

Each measurement removed 250 μL of headspace volume from the flasks. The same amount of compressed sterile air was added to keep pressure equilibrium between the flasks and the atmosphere, as well as sustain between 15 and 25% (v/v) of O₂ throughout the entire experiment. No depressurizing was necessary for the bottles, as sterile needles inserted in the buffer valve would still result in contamination. The CO₂ production rate for every 5 days was then used to produce a surface response according to the proposed factorial design.

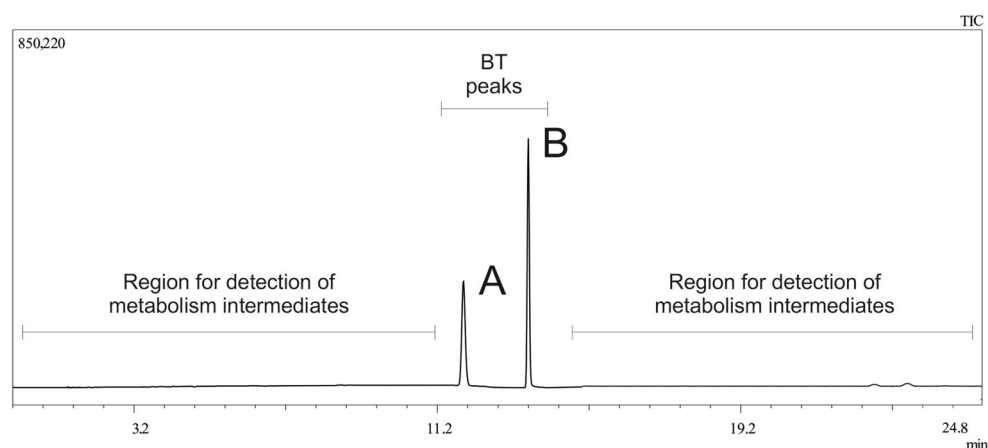
Headspace analysis of aromatic hydrocarbon using GC–MS

We monitored the biodegradation of BT using a Shimadzu QP2010 Ultra GC–MS. The concentration of BT was monitored through a 100 μL headspace aliquot injected into a Rtx-5MS chromatographic column. The column's stationary phase was composed of 5% phenyl and 95% dimethylpolysiloxane (30 m × 0.25 mm × 0.25 mm). Sampling was performed every 5 days, which is also the same interval in which CO₂ analyses were performed.

The parameters were modified from the method 12-154 Shimadzu for general hydrocarbon detection. The carrier gas was helium. For a reliable analysis that did not impair the equipment and the column, the injector and detector temperatures were set to 290 °C, although the original methodology recommends 310 °C. Our results show no noticeable difference in signal intensity when the temperature is lowered. The BT chromatogram (Fig. 1) showed sufficient resolution to quantify the consumption of BT with an initial signal intensity greater than 5000 pA. The following temperature program was applied: 35 to 85 °C, with a 60 °C min⁻¹ ramp and then an increase from 80 to 115 °C at a rate of 35 °C min⁻¹. The total run time for each sample was 5.6 min.

Similarly to the CO₂ production rate measurements, the gas samples did not undergo any type of pre-preparation. Headspace was injected directly after removal from serum bottles. The gas withdrawal was performed with a glass gastight syringe that perforated directly through the rubber caps. Serum bottles were not opened throughout the experiment, to avoid risking any interference with the biodegradation process.

Fig. 1 Chromatogram from a benzene (peak A) and toluene (peak B) mixture



Liquid extraction and analysis of biodegradation products

After 50 days, the serum bottles were opened and mixed with 20 mL of ethyl acetate and 10 mL of dichloromethane. This solvent choice was aimed at increasing the miscibility between the solvent and the sample during the chromatographic stages. The following steps were modified from Gołębiowski et al. (2011) and Yang et al. (2017). Our method aimed towards full optimization of the extraction methodology applied to aromatic hydrocarbons. This method also ensured that most of the metabolites that eventually formed inside the flasks would be extracted as well. The 1:1:2 mixture of solvent and sample was subjected to manual stirring for 5 min and sonicated for 30 min. In addition, the soil particles were separated from the liquid phase by centrifugation at 15,000g for 15 min prior to further processing. The contents were rapidly transferred into 2.5-mL vials and analyzed by GC–MS.

Our method for BT quantification from headspace aliquots was extremely optimized for such peaks. However, such short running time was not adequate for the detection of metabolic by-products which may have been formed during the BT biodegradation process. For this reason, we expanded this method for the detection of compounds over a wider spectrum of chromatographic separation. Figure 1 shows the regions in which the presence of aromatic biodegradation intermediates was expected in this expanded method. The products could be smaller biodegraded compounds or even larger metabolic intermediates throughout the process.

The parameters we used were similar to the headspace analysis, except that temperature program applied was 35 °C kept for 5 min, then a ramp of 10 °C min^{−1} until 220 °C was reached followed by a steeper temperature ramp at 20 °C min^{−1} until 250 °C was reached. The total run time for each sample was 25.0 min.

Mass spectrum (70 eV) was specifically used to identify the new compounds. While the BT biodegradation in the previous section was detected from peak area integration and a standard

solution-based calibration curve, the mass spectrometry allowed the identification of new compounds formed after 50 days of biodegradation. The compounds were analyzed according to their mass spectra. The ion source of the mass-spectrometer was kept at 200 °C. A detector containing an electron ionization source and a quadrupole mass analyzer was operated in scan mode for chemical identification. The detector was at 250 °C and the source of ions at 200 °C. The NIST11 Mass Spectrum Library was used for identification of biodegradation intermediates.

Results and discussion

CO₂ production rate

The interaction effects of temperature, BT, and fluorotelomer concentration on the degradation rates have been illustrated with the contour plots shown in Fig. 2.

The contour plots in Fig. 2a depict that the degradation rate increases at higher temperatures. The observed temperature effect on CO₂ production rate was expected since higher temperatures are related to an increase in microbial activity. Seasonality has a detectable effect during summer (35 °C) and winter (20 °C) conditions, yielding different CO₂ production rates. In other words, GC-TCD respirometry was a very sensitive technique to detect variations in CO₂ concentrations within mild weather conditions. Such conditions, however, were observable in a controlled laboratory environment. Extreme temperatures would be beyond the resolution limit of our method, as column saturation would be too high in summer or compound degradation would be too low to be detected in winter.

No isolated peaks were observed at the response surface for both temperature and fluorotelomer concentration. The contour plots show that there is a semi-linear approach to CO₂ production response when all three factors are considered. Even though there is no peak for specific concentrations or

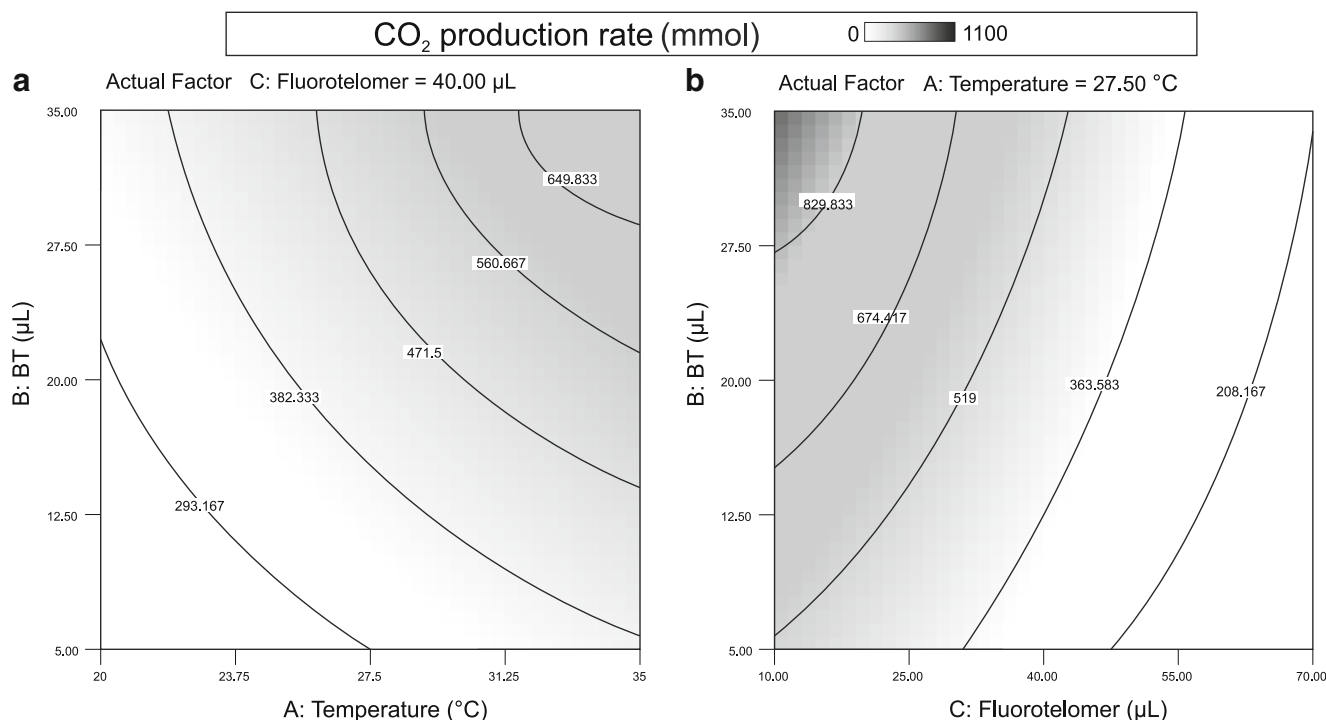


Fig. 2 Contour plot of temperature, benzene, toluene, and fluorotelomer concentrations factors in CO₂ production rate response at a set fluorotelomer concentration (a) and temperature (b)

temperatures, the tested parameters (temperature, BT, and fluorotelomer) were all related to CO₂ variations.

Higher temperatures increased CO₂ production rate, while higher concentrations of fluorotelomer inhibited the biodegradation process. The BT factor, however, caused a directly proportional increase in CO₂ production at higher concentrations. Increased benzene and toluene concentration turned them into a readily available substrate to the indigenous microorganisms in our soil samples, as a result of increased degradation rates. It is worth mentioning, however, that the BT concentrations in this study were carefully planned to be below toxicity levels, since beyond 200 mg L⁻¹ the CO₂ response was expected to decrease, thus indicating inhibition due to the toxicity of benzene. The inhibition effect of toluene was also reported at a concentration as low as 158 mg L⁻¹, according to Monero et al. (2003).

The response surface contours (Fig. 2) can also properly identify the interactions between variables and predict the maximum response of a given measured parameter. According to Zhao et al. (2016), a circular contour pattern is likely to have better interaction between the variables, whereas an elliptical pattern indicates non-significant interaction. The contour plots observed in our co-contaminations containing fluorotelomer and BT were all circular for the fluorotelomer factor (Fig. 2b), indicating a relevant interaction between the factors.

The respirometric analysis pointed towards the inhibition of CO₂ production when higher concentrations of fluorotelomer are present. There is a balance between an increase in CO₂ production due to the presence of higher concentrations of

BT compounds and an inhibition of microbial activity at higher firefighting foam concentrations. Contrary to BT, the high concentration of fluorotelomers in our assays led to toxicological effects that may have halted several cellular functions and therefore lowered CO₂ production to a minimum.

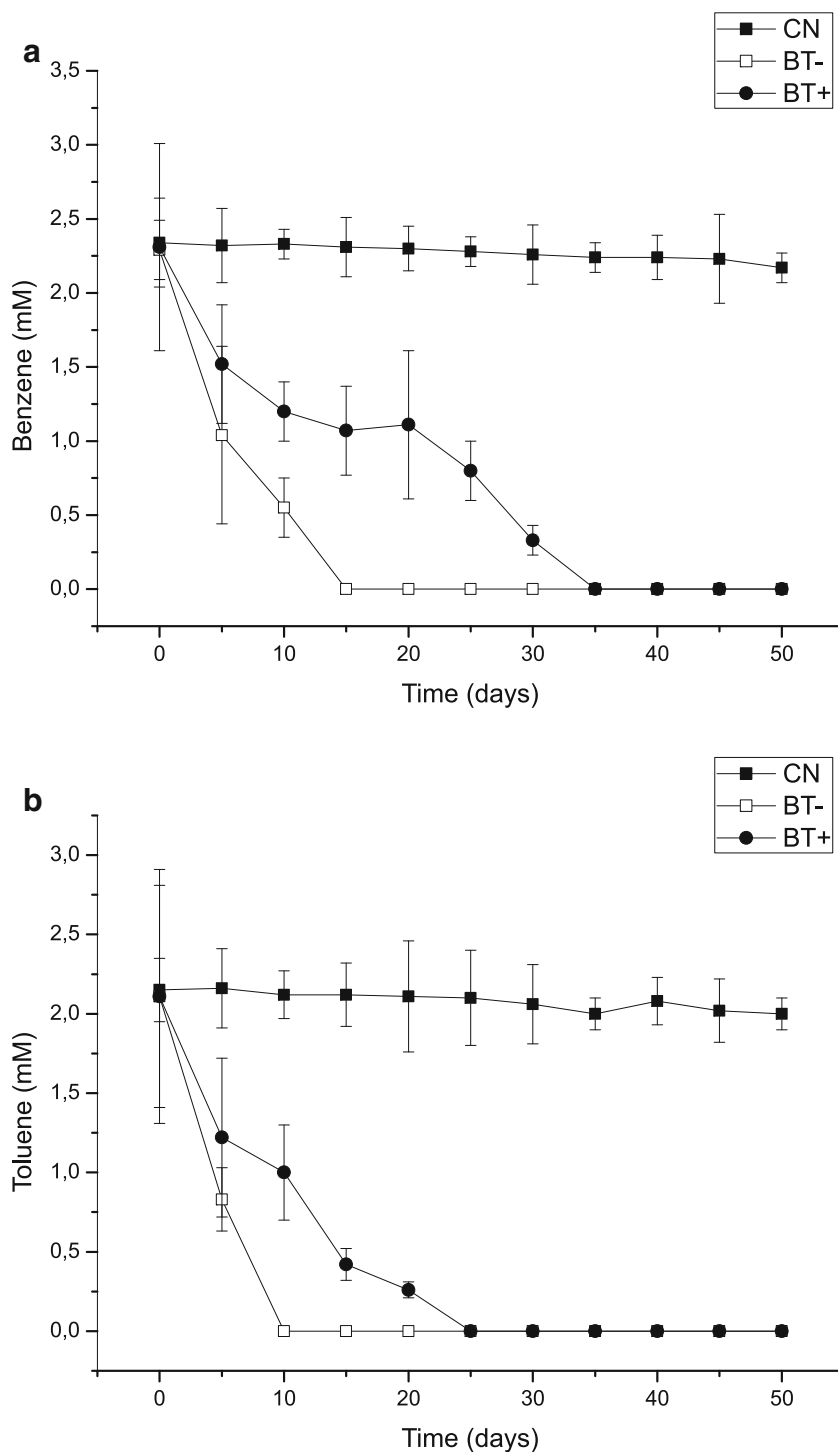
Our mixed-substrate assays simulated the main bioremediation challenge when creating new cost-effective and viable technologies for environmental remediation applications (Vinuseli et al. 2012; Aparicio et al. 2018). We encourage further studies that promote an in-depth investigation of the effects of perfluorinated compounds in diminishing removal efficiency of aromatics. This is particularly urgent since environmental contamination containing both compounds is fairly common in the petrochemical industry. Understanding the BT biodegradation profile may also provide additional information, compared to CO₂ monitoring. Therefore, BT consumption profile can elucidate the key role of fluorotelomer in the biodegradation processes.

Benzene and toluene biodegradation

The difference between datasets from BT assays was considered statistically significant ($p > 0.05$) compared to controls (C_M, C_N, and C_S). In fact, control assays did not statistically vary when compared to each other, with threshold variations in soil controls (C_S). The biodegradation of the individual BT components throughout 50 days is shown in Fig. 3.

The soilless assays (C_M) yielded null microbial activity (as expected from non-inoculated media). No aromatic

Fig. 3 Benzene (a) and toluene (b) concentration in the biodegradation assays



biodegradation was detected within those assays. This is important since the minimum output confirmed that no leaks or adsorption of volatiles occurred during the course of our entire experiment. Even though benzene and toluene substrates were depleted at less than 40 days for each assay, we kept the assays incubated for 50 days, since the biodegradation process of aromatics could still be occurring.

Figure 3 shows the effect of mid-range fluorotelomer concentration on BT consumption. Based on our results, benzene and toluene did not benefit from the addition of perfluorinated substances in culture media. Without fluorotelomer, the biodegradation of each BT component was seen to occur at a much higher rate than that of other assays. The absence of an adaption phase to the BT substrate during consumption form BT- assays means that no sudden inhibition to culture

growth was caused by benzene or toluene. This might be due to the proper enrichment at a tolerable BT concentration (Kim et al. 2005; Khodaei et al. 2017).

On average, benzene and toluene were depleted 17.5 days earlier in BT[−] assays than those in BT⁺ assays. We believe that this is not due to a preferential consumption of fluorotelomers over BT by the indigenous microbes, whose metabolic pathways will less likely prioritize perfluorinated compounds as substrate. There was a noticeable decrease in degradation rates of BT when perfluorinated compounds were present. This decrease signifies that the fluorotelomer repressed BT catabolism in batch cultures. Thus, fluorotelomers were a dominant repressing substrate for our indigenous soil microbiota from a petrochemical plant. We hypothesize that BT consumption was found to be reduced in the presence of fluorotelomers due to their toxicity to the indigenous soil microbiota, since monoaromatic consumption rate decreased only when fluorotelomers were present in the batches, hence indicating inhibition.

From Fig. 3, it can also be concluded that monoaromatic uptake by the soil inoculum had undergone some preference mechanism. In other words, one of the aromatics was preferably consumed as a carbon source among the two available hydrocarbons. Toluene removal efficiency, in particular, was notably higher than benzene. The results on toluene consumption were much higher compared to the ones found in a broad literature review by Yeom et al. (1997), as well as recent results by Su et al. (2014) and Rajamanickam et al. (2017). Therefore, the conditions adopted in this study demonstrated that an unusual approach (i.e., testing fluorocarbon effect in soil microbiota) found conditions that may favor an improved biodegradation process of toluene, despite the inhibition effects of perfluorinated compounds at a higher concentration.

Benzene took a longer time to be consumed than toluene. Still, both degradation pathways were activated when BT was available. On the other hand, fluorotelomer negatively affected the biodegradation rates. It is uncertain whether this also means competitive inhibition, as BT degradation pathways are too complex. Even though the analysis of metabolic pathways is beyond the scope of this study, we further investigated the formation of intermediate compounds after the proposed 50 days of biodegradation monitoring. This was the reason for incubating our assays 20 days after benzene and toluene were no longer detected at headspace sample analyses.

Biodegradation intermediates formation

After the 50-day biodegradation period, we searched for metabolites that could have been formed in the headspace or the soil matrix. The differential biodegradation profile between assays with and without fluorinated compounds prompted us to perform a detailed analysis of the chromatogram to search for intermediates of aromatic hydrocarbon biodegradation

(Fig. 4) by cross-referencing peaks with other common known BT degradation intermediates at the NIST11 Mass Spectrum Library.

In order to determine the metabolic intermediates during BT biodegradation, samples were collected from the optimized mid-range fluorotelomer assays after 50 days of incubation for analysis in GC–MS. The result of GC–MS showed the presence of various compounds. However, the compound formation was different when perfluorinated compounds were present. Many commonly reported biodegradation intermediates of aromatic hydrocarbons (Singh and Fulekar 2010) such as catechol (Fig. 4, indicated as 9), maleic acid (Fig. 4, indicated as 4), 2-hydroxymuconate semialdehyde (Fig. 4a, indicated as 7), and p-cresol (Fig. 4, indicated as 8) were found in our chromatograms.

The presence of metabolites confirms the biodegradation of benzene and toluene by the indigenous microbial culture from refinery soil. Most of the compounds were associated with the aerobic activity of the dioxygenase enzyme. The mono- and dioxygenases are often involved in the oxidation of aromatic ring of benzene and toluene by incorporating two oxygen atoms to form 2-hydroxy-substituted radicals (Littlejohns and Daugulis 2008; Kim et al. 2009). The metabolic intermediates produced during the BT biodegradation in this study have been supported by many established biodegradation pathways in literature. The microbial adaptation to aromatics (especially benzene and toluene) was broadly studied, as wisely compiled by El-Naas et al. (2014) in their review. Recent studies on aromatics suggest that the adaptation of microbial communities affects the usually reported biodegradation pathways, since co-metabolites may result in higher efficiency of aromatic removal from a contaminated site.

Even though BT⁺ (Fig. 4b) and BT[−] (Fig. 4a) showed differences in metabolites, they shared similarities to many of the expected compounds. It is worth mentioning that the first peak (Fig. 4, indicated as 1) corresponds to the elution solvent (ethyl acetate) and it is not a biodegradation metabolite. The main similarity between intermediates in BT⁺ and BT[−] assays, which is also the source of many subsequent oxidation steps, is catechol (Fig. 4, indicated as 9). Catechol was found in both of our assays (Fig. 4, indicated as 9), in which its further mineralization is associated with the support of biomass growth, despite the presence of fluorotelomers. Under certain circumstances, this compound is usually converted to 2-hydroxymuconate semialdehyde (Fig. 4a, indicated as 7), among many others, and is ultimately converted to substrates of the citric acid cycle that ultimately leads to the final biodegradation step that produces CO₂ and water (Meckenstock et al. 2004; Kadri et al. 2017). The 2-hydroxymuconate semialdehyde, however, was not found in BT⁺ assays (Fig. 4b), thus indicating a metabolic shift due to fluorotelomer influence.

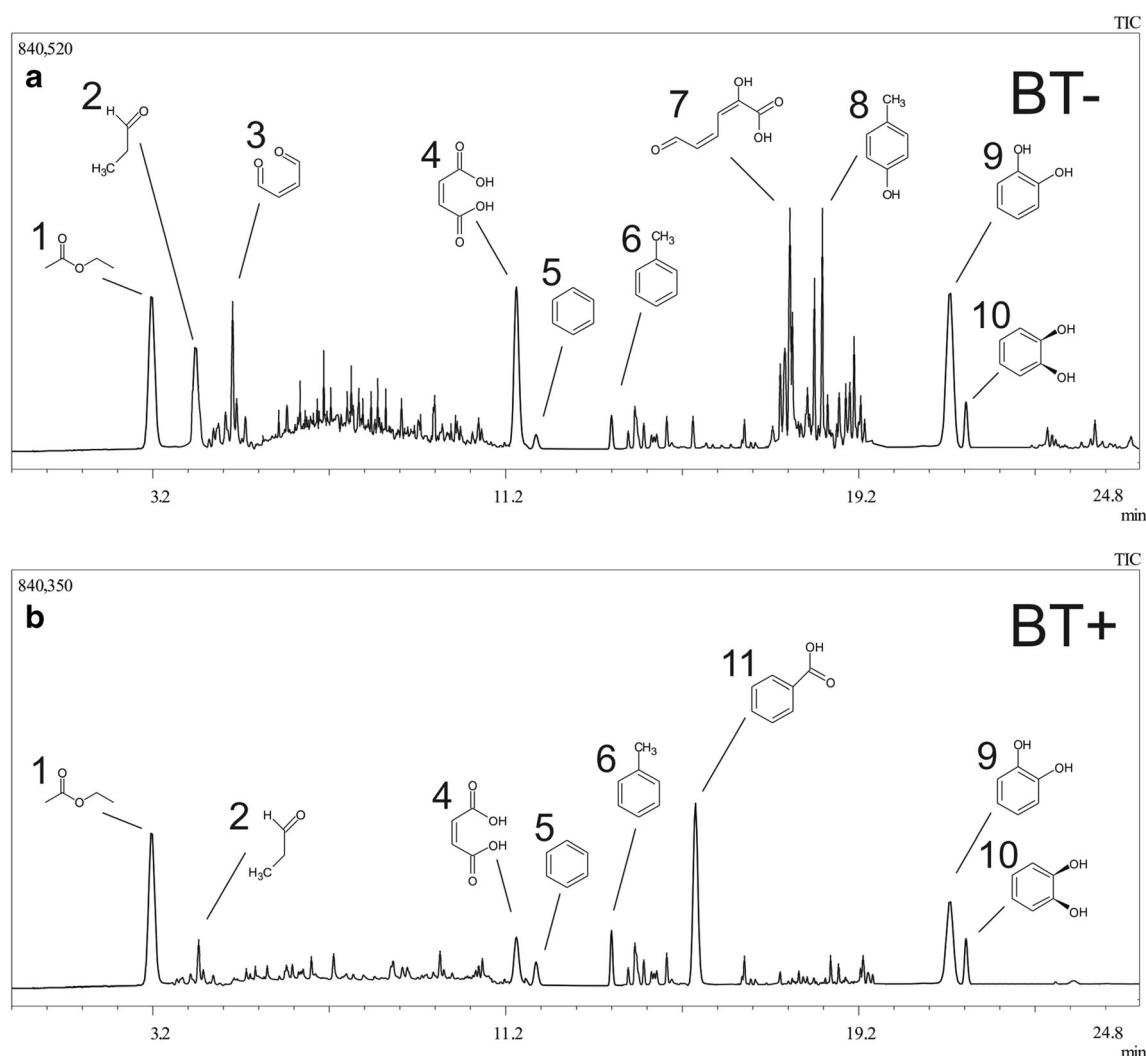


Fig. 4 Post-biodegradation chromatogram without BT (**a**) and with BT (**b**) showing peaks identified by their mass spectra indicated as 1, ethyl acetate; 2, propionaldehyde; 3, butanediol; 4, maleic acid; 5, benzene; 6,

toluene; 7, 2-hydroxymuconate semialdehyde; 8, p-cresol; 9, catechol; 10, cis-1,2-dihydrobenzene-1,2-diol; 11, benzoic acid

As shown in Fig. 4, GC–MS analysis of BT and its by-products (after microbial consumption) had possibly identified two exclusive metabolites via similarity search on the NIST11 library. The benzene and toluene molecules were shown as very small peaks after the biodegradation process (Fig. 4, indicated as 5 and 6, respectively) when compared to the control chromatogram (Fig. 1). The lengthy run time with a total retention time of 24.88 min allowed a thorough capture of many compounds. An alternative intermediate peak eluted at 15.49 min exclusively at BT+ assays (Fig. 3b, indicated as 11) compared to the BT– chromatogram. It was identified as benzoic acid and is associated with alternate toluene biodegradation routes. Considering this molecule is a late compound in toluene H-abstraction pathway (Ramos et al. 1997), we may conclude that biodegradation occurred up to more advanced steps. Despite the lower diversity of compounds in BT+ assays, an alternate pathway was triggered by the presence of perfluorinated

compounds. Peak intensity was much lower in BT+ than that in BT– though, which confirms the minor biodegradation rate found in both headspace and respirometric approaches.

Our previous studies on aromatics and fluorinated compounds (Montagnolli et al. 2017) suggested that the adaptation of microbial communities to the fluorotelomers had an impact on the biotransformation of the carbon sources. The fluorinated compounds changed the BTEX biodegradation products, as the halogenated substances could have contributed to a shift in styrene concentrations in co-contamination scenarios. Catechol was also rapidly metabolized to further metabolic stages. At that time, we were unsure whether any carbon source or specific aromatic biodegradation enzymes could further expand biotransformation. These results led to the follow-up experiments proposed in this paper and the search for other compounds (such as benzoic acid and p-cresol) that could be related to alternative pathways triggered by fluorotelomers.

While we cannot assure that our proposed method had scanned all compounds present in our post-biodegradation samples due to factors such as the volatility, melting temperature, molecular structures, and various other conditions that are important for a successful final elution, our approach was able to detect benzene and toluene metabolites. All biodegradation intermediates found in this study share properties such as high volatility, low molecular weight, and low boiling temperature. The GC–MS-based metabolite analysis had a high sensitivity and successfully confirmed BT in many established biodegradation pathways available in current literature (Bielefeldt and Stensel 1999; Gieg et al. 2014; Varjani 2017). The differences between metabolites in BT+ and BT– assays, however, were the most relevant result. The discovery of an unusual environmental behavior of co-contamination scenarios can help unraveling the effect of enzyme expression on metabolism and organism performance (Jonsson et al. 2004; Schauer et al. 2005; Vogt et al. 2016) in corresponding bioremediation strategies.

Other apparent metabolites that eluted at 5.6 and 17.9 min were butanediol and 2-hydroxymuconate semialdehyde, respectively (according to the NIST library). These compounds were only present in assays without perfluorinated substances. Both intermediates detected by GC–MS could be the result of specific enzymatic mechanisms that are inactivated in the presence of fluoride or enzymes in the classical pathways that are inhibited, leading to alternate pathways. It has been long demonstrated (over 20 years ago) by Yeom et al. (1997) that even microbes adapted to only one of aromatic compounds may change their enzymatic mechanisms, thus attaining the ability to degrade toluene and other monoaromatics faster than control assays through—at that time—unidentified pathways. However, there are few studies that link perfluorinated compound co-contamination to hydrocarbon biodegradation. Therefore, the differential biodegradation mechanisms observed in our co-contamination scenarios through GC–MS analysis might lead to the development of different bioremediation strategies that can predict differential accumulation of compounds in the environment.

Conclusions

Perfluorinated substances caused the biodegradation process of monoaromatic hydrocarbons to significantly differ. When seasonal effects were considered, a reduced biodegradation rate (in terms of CO₂ production rate) indicated inhibitory effects, despite different temperatures. In fact, the sole presence of such organo-halogenated substances was capable of reducing the microbial uptake of benzene and toluene. Moreover, the intermediate compound analysis confirmed that a specific compound could be detected exclusively to soil microbiota contaminated with perfluorinated compounds. In summary, we found that fluorotelomers can trigger a new

metabolic pathway which may lead to previously unknown biodegradation profiles and microbial behavior.

Thus, we propose that transient components of the benzene and toluene degradation may be differentially formed, causing the BT and perfluorinated co-contaminations to go through switched metabolic stages under the presence of fluoride in a co-contamination scenario. The fluorotelomers contributed to the development of different concentrations of by-products, even though no production/consumption spikes occurred when concentration and temperature factors were evaluated. The synergic effect of firefighting foams on benzene and toluene biodegradation might provide a better perspective for more efficient bioremediation techniques and fully optimized environmental cleaning-up strategies. The better understanding of microbial behavior and biodegradation patterns at firefighting sites can narrow current efforts for the improvement of ongoing development on environmental biotechnology. Therefore, we encourage further studies about novel benzene and toluene metabolites that expand the current knowledge of the effects of organo-halogenated substances during the biodegradation of monoaromatics.

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