## UNIVERSIDADE ESTADUAL PAULISTA – UNESP CENTRO DE AQUICULTURA DA UNESP

## CONTABILIDADE AMBIENTAL DE SISTEMAS SEMI-INTENSIVOS DE AQUICULTURA: ESTUDO DE CASO DA LAMBARICULTURA

Tamara Fonseca de Almeida

Jaboticabal, São Paulo 2021

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Orientador: Prof. Dr. Wagner C. Valenti

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### DEDICATÓRIA

Aos pequenos produtores rurais do Brasil.

"Fomos, durante muito tempo, embalados com a história de que somos a humanidade. Enquanto isso, fomos nos alienando desse organismo de que somos parte, a Terra, e passamos a pensar que ele é uma coisa e nós, outra: a Terra e a humanidade. Eu não percebo onde tem alguma coisa que não seja natureza. Tudo é natureza [...]. Enquanto isso, a humanidade vai sendo descolada de uma maneira tão absoluta desse organismo que é a Terra. Os únicos núcleos que ainda consideram que precisam ficar agarrados nessa Terra são aqueles que ficaram meio esquecidos pelas bordas do planeta, nas margens dos rios, nas beiras dos oceanos, na África, na Ásia ou na América Latina. "

(Ailton Krenak, Ideias para adiar o fim do mundo)

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#### SUMÁRIO

RESUMO	9
ABSTRACT	10
INTRODUÇÃO GERAL	11
CAPÍTULO 1: Environmental accounting of the yellow-tail lambari farming: small cha do make a difference	
Abstract	16
1. Introduction	17
2. Materials and Methods	
3. Results	25
4. Discussion	
5. Conclusion	35
CAPÍTULO 2: Ecosystem functions in pond aquaculture and their roles in providing ecosystem services and disservices	44
Ecosystem functions in pond aquaculture and their roles in providing ecosystem ser and disservices	
1. Introduction	46
2. Materials and methods	
3 Results	54
4 Discussion	55
5. Conclusion	59
CAPÍTULO 3: Multi-Criteria Sustainability Assessment of Lambari Aquaculture in Bra	<b>zil</b> 65
Abstract	66
1. Introduction	67
2. Materials and Methods	69
3. Results	73
4. Discussion	76
5. Conclusion	78
CONCLUSÕES GERAIS	
REFERÊNCIAS COMPLEMENTARES	

#### RESUMO

A aquicultura pode ser uma ferramenta para conciliar o desenvolvimento socioeconômico à conservação dos recursos naturais, contribuindo para o desenvolvimento sustentável de comunidades rurais. A atividade vem crescendo de forma acelerada no Brasil e é realizada majoritariamente por pequenos produtores rurais, com o uso de sistemas semi-intensivos de água doce. Os lambaris são um grupo de peixes nativos com alto potencial para a aquicultura sustentável. Dessa forma, o presente estudo tem como objetivo avaliar a sustentabilidade dos sistemas de produção de lambari-do-rabo-amarelo por três diferentes abordagens: síntese em emergia, funções ecossistêmicas e avaliação multicritério dos cinco setores. Os resultados indicam que os níveis de controle do produtor (baixo, moderado e alto) sobre as práticas de manejo afetam a eficiência na utilização de recursos naturais. Além disso, os viveiros de aquicultura são ecossistemas antrópicos que podem ser manejados para a maximização de externalidades positivas e minimização de externalidades negativas, aumentando a resiliência dos sistemas produtivos por meio da restauração dos recursos naturais, dos quais ele depende. Finalmente, os sistemas de baixo controle são socialmente mais sustentáveis e contribuem mais para o desenvolvimento local, enquanto os sistemas de moderado e alto controle são economicamente mais lucrativos e utilizam os recursos naturais de forma mais eficiente.

**PALAVRAS CHAVE:** sustentabilidade; economia ambiental; desenvolvimento rural; peixes tropicais; aquicultura de água doce; pequenos produtores; uso de recursos naturais.

### ABSTRACT

Aquaculture can be a tool to reconcile socioeconomic development with the conservation of natural resources, contributing to the sustainable development of rural communities. The activity has been growing fast in Brazil, and is performed mainly by small rural producers, in semi-intensive freshwater systems. Lambari is a group of native fish with high prospects for sustainable aquaculture. Therefore, this study aims to access the sustainability of the lambari aquaculture production in Brazil by three approaches: emergy synthesis, ecosystem functions, and multicriteria assessment of the five-sectors sustainability model. The results indicate that the levels of control (low, moderate and high) over the management practices adopted by farmers affect the efficiency of natural resources consumption. In addition, aquaculture ponds are man-made ecosystems that can be managed to maximize positive externalities and minimize negative externalities, increasing the resilience of the productive systems by restoring the natural resources in which they depend. Finally, the low-control systems are more sustainable socially, and contribute more to local development, while moderate and high-control systems are higher economically feasible and use natural resources more efficiently.

**KEYWORDS:** sustainability; freshwater aquaculture; smallholder farms; natural resources consumption.

### INTRODUÇÃO GERAL

Conciliar o desenvolvimento socioeconômico à conservação dos recursos naturais é um dos maiores desafios globais para o desenvolvimento de comunidades rurais (Goswami et al., 2017; Jung et al., 2017). A aquicultura pode ser uma ferramenta para a solução desse problema. Ela tem o potencial de ser mais sustentável quando comparada às monoculturas agrícolas e à produção de animais terrestres (Costa-Pierce and Page, 2010). Além disso, a aquicultura pode contribuir para a geração de renda e empregos diretos, indiretos e auto-empregos, e produzir alimentos de alto valor nutricional (Béné et al., 2016).

Os principais desafios atuais para o progresso da aquicultura de água doce em países em desenvolvimento, como o Brasil, estão relacionados à regulamentação governamental, a falta de uma cadeia produtiva bem estruturada, aos impactos e poluição ambiental, fuga de espécies exóticas e híbridas interespecíficas, bem-estar e saúde animal, nutrição adequada e suporte técnico (Boyd et al., 2020; Brugère et al., 2019; Henares et al., 2019). A comunidade científica tem desenvolvido diversas tecnologias para enfrentar alguns desses desafios, tais como: tecnologia para tratamento de efluentes, pacotes tecnológicos para o cultivo de espécies nativas, medicamentos e probióticos, melhoramento genético, diminuição da taxa de conversão alimentar, substituição da farinha de peixe na ração por proteína vegetal e sistemas de recirculação (Antonucci and Costa, 2020; Dawood et al., 2019; Humphries et al., 2019; Lulijwa et al., 2019; Tacon, 2020). Apesar das expressivas melhoras alcançadas, a maioria destas soluções são caras e, por vezes, inatingíveis por pequenos produtores rurais. Além disso, a tecnologia atual pode se revelar ineficaz na mitigação das ameaças complexas causadas pela crise da COVID19 e pelas mudanças climáticas em um futuro próximo. Reconhecendo que o modelo de "business as usual" não tem contemplado essas questões, tecnologias inovadoras na aquicultura, que possam se ajustar a esses desafios, serão vitais para sua sustentabilidade a longo prazo (United Nations, 2020).

A aquicultura é o setor de produção de alimentos que cresceu mais rápido nas últimas décadas, e tem previsão de crescer 37% até 2030 (Garlock et al., 2020). A produção aquícola mundial atingiu 82.1 milhões de toneladas de pescados e 32.4 milhões de toneladas de plantas aquáticas em 2018, e já ultrapassou a pesca em 35 países. O Brasil ocupa a 8ª posição no ranking mundial de maiores produtores de peixes pela aquicultura (FAO, 2020). A produção aquícola em 2019 foi de aproximadamente 800.000 toneladas, o que representa uma receita bruta de ~US\$ 1 bilhão. Atualmente, mais de 200 mil fazendas de piscicultura de água doce estão em atividade no Brasil (Valenti et al., 2021). As espécies de água doce são as mais produzidas, sendo que a tilápia (*Oreochromis niloticus*) e o tambaqui (*Colossoma macropomum*) são predominantes (IBGE, 2020). No entanto, outras espécies nativas, como o lambari (*Astyanax lacustris*), possuem alta relevância socioeconômica regional. Além disso, a maior parte da produção aquícola brasileira vem de pequenas propriedades rurais (<2 ha), onde o cultivo é realizado em viveiros escavados de água doce (Valenti et al., 2021).

Entre as espécies de peixes nativos com grande potencial para a aquicultura, destaca-se os lambaris (Fonseca et al., 2017). O cultivo surgiu como uma fonte de renda alternativa para pequenos produtores rurais no sudeste brasileiro. A produção cresceu nos últimos anos baseada no mercado de iscas vivas para pesca recreativa. No entanto, além do uso como iscas, o lambari também é consumido como aperitivo em bares e restaurantes. Algumas espécies apresentam grande potencial para o mercado de peixes ornamentais. Ainda, o uso como um substituto mais sustentável das sardinhas utilizadas na pesca industrial de atum tem sido investigado. A maioria dos lambaricultores é formada por pequenos produtores familiares, que adotam sistemas de produção semi-intensivos em viveiros de fundo natural e a principal espécie cultivada é o lambari-do-rabo-amarelo (*Astyanax lacustris*) (Fonseca et al., 2017). Porém, uma recente expansão no mercado tem atraído investidores com mais capital, que operam em fazendas maiores (> 20 ha) e demandam infraestrutura mais complexa.

Fonseca et al. (2017) revisaram as informações disponíveis sobre a produção de lambaris e identificaram que os sistemas de cultivo e estratégias de produção variam entre os produtores. Cada produtor estabeleceu a sua estratégia empiricamente ou baseado em protocolos para outras espécies (Silva et al., 2011). Práticas de manejo menos eficientes são frequentemente adotadas e as informações científicas são insuficientes para gerar tecnologias adequadas às necessidades dos produtores (Fonseca et al., 2017). Além disso, pelo fato de não

existir dieta comercial específica para o lambari, os produtores escolhem a ração com base no tamanho do pélete que o animal é capaz de ingerir (Silva et al., 2011); essa situação persiste ainda em 2020. Comumente são usadas dietas desenvolvidas para juvenis de outras espécies, com alta concentração de proteína bruta e de alto custo financeiro. Esses fatores implicam em baixa produtividade e uso inadequado dos recursos naturais (Fonseca et al., 2017).

Como espécie nativa e de baixo nível trófico, o lambari tem potencial para produzido de forma sustentável, promovendo o desenvolvimento ser socioeconômico e conservando os recursos naturais. A diversidade de mercados permite a atuação de produtores de diferentes tamanhos e níveis de tecnificação. Além disso, o cultivo de lambaris pode ser implantado em pequenas propriedades como atividade complementar, aumentando a renda familiar. Porém, para que essas potencialidades sejam atingidas, é essencial conhecer os sistemas de produção usados e a conjuntura nas quais eles se inserem para identificar seus pontos fracos e sugerir alternativas. Assim, medir a sustentabilidade dos sistemas permite identificar os principais gargalos, bem como conhecer as forças condicionantes da produção sustentável. Com base nessas informações, pode-se tornar a produção mais eficiente e adequada à realidade de cada produtor, bem como aumentar a sustentabilidade ambiental, econômica e social.

No capítulo 1, a sustentabilidade dos sistemas de produção de lambari foi avaliada sob a ótica da Síntese em Emergia. Emergia, com "m", compreende toda a energia utilizada direta ou indiretamente para a produção de um bem ou serviço (Odum, 1996). Utilizando esse método, foi possível quantificar o investimento da natureza no sistema produtivo sob uma abordagem eco-cêntrica. Ainda, foi possível avaliar a dependência dos sistemas sobre os recursos naturais renováveis e nãorenováveis, e compará-los identificando quais estratégias de manejo da produção são mais ou menos eficientes na utilização dos recursos.

No capítulo 2, foram avaliadas as principais funções ecossistêmicas que ocorrem nos viveiros escavados utilizados para a lambaricultura, e de quais formas elas influenciam na prestação de serviços e desserviços ecossistêmicos. As funções ecossistêmicas são os processos ecológicos que controlam os fluxos de energia, matéria orgânica e nutrientes nos ecossistemas naturais (Lee and Brown, 2021). Os viveiros de aquicultura podem ser vistos como ecossistemas antrópicos,

manejados para maximizar o serviço de provisão de biomassa: peixes. No entanto, outros serviços como, regulação de microclima, e desserviços como a geração de efluentes, ocorrem simultaneamente. Dessa forma, as estratégias de manejo que maximizam os serviços e minimizam os desserviços foram investigadas.

Finalmente, no capítulo 3, os sistemas de lambaricultura foram avaliados pela ferramenta de análise multicritério de sustentabilidade dos cinco setores (5SEnSU). Ela se baseia na premissa de que os sistemas humanos são considerados termodinamicamente abertos, os quais demandam energia e materiais advindos da natureza, que serão transformados em bens e serviços por meio do trabalho humano (Giannetti et al., 2019). De acordo com modelo conceitual 5SEnSU, o meio ambiente atua como fornecedor de recursos naturais (setor 1) e recebedor de subprodutos e resíduos (setor 2) do setor produtivo (setor 3). Por outro lado, a sociedade atua como fornecedora de mão-de-obra e tecnologia (setor 4), e recebedora dos bens e serviços (setor 5) produzidos pelo setor 3 (Giannetti et al., 2019). Este modelo permite uma visão holística do sistema produtivo, que ocorre por meio da quantificação das trocas físicas que ocorrem entre os setores ambiental, econômico e social, o que confere uma mensuração cientificamente robusta da sustentabilidade dos sistemas produtivos.

## CAPÍTULO 1: Environmental accounting of the yellow-tail lambari farming: small changes do make a difference

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# Environmental accounting of the yellow-tail lambari aquaculture: small changes do make a difference.

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#### 14 Abstract

Freshwater pond aquaculture is the prevailing fish culture system worldwide, especially in 15 developing countries. Climate change outcomes and inadequate environmental practices 16 challenge its sustainability. This study applies emergy synthesis to assess the environmental 17 18 performance of freshwater pond aquaculture in Brazil, aiming to identify and propose management practices towards sustainability. As a study model, nine semi-intensive lambari 19 farms operating at three levels of management were evaluated: low (LC), moderate (MC) and 20 high (HC) control. Results showed that the main inputs for LC are services (27-46%), feed 21 (7-39%) and water (15-21%), while for the MC farms, the main inputs are feed (35-49%) and 22 services (33-39%), and for HC farms, the main flows are feed (17-48%) and services (26-23 36%). All farms required more than 60% of their emergy from purchased inputs, resulting in 24 low emergy sustainability index (ESI). By replacing fish meal and fish oil by vegetal protein 25 and oil on feed, using superficial water instead of spring water, increasing juvenile 26 27 productivity, and controlling pond fertilization improved the emergy performance, leading all systems to higher efficiency and resilience. Therefore, simple changes in culture practices 28 do make a difference. 29

30 **Keywords**: Best management practices; Brazil; Emergy; Fish production.

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#### 31 **1. Introduction**

The sustainable development was stated as a fundamental goal in the ecosystem 32 approach for aquaculture (EAA) document published by FAO in 2008 (Soto et al., 2008), and 33 it still remains a major concern (Boyd et al., 2020; FAO, 2020). Discussions on the 34 sustainability of aquaculture have focused on the assessment of different production systems 35 or levels of intensification. Different methods in quantifying sustainability has been used, such 36 as life cycle assessment, sets of sustainability indicators, and emergy (with an 'm') synthesis. 37 Additionally, innovative systems, such as integrated multi-trophic aquaculture (IMTA), 38 aquaponics and bioflocs have been developed to achieve higher productivity and 39 40 sustainability (Henares et al., 2019; Mungkung et al., 2013; Valenti et al., 2018; Wilfart et al., 2013). The majority of the aquaculture systems are small-scale and inland located, despite 41 that fact, they have received less attention in strategic planning and management within EAA 42 concepts than coastal and marine systems (Brugère et al., 2019). Possibly, the EAA 43 framework lacks of a systemic approach to understand how small-scale aquaculture works, 44 i.e. the ways they are connected with the surrounding social, economic and environmental 45 systems. 46

The small-scale inland aquaculture should shift to more sustainable production systems 47 in order to achieve the goals established by EAA guidelines, and by the 2030 agenda for the 48 sustainable development. Strategies towards sustainability includes the use of native 49 species, efficient use of feed and locally available resources, the level of control and 50 monitoring over the production variables, technical qualification and infrastructure, and 51 residue treatment (Boyd et al., 2020). Nevertheless, the current strategies are costly and 52 sometimes unattainable by rural small farmers. Moreover, SARS-CoV-2, climate change and 53 economic crises may increase the vulnerability of small farms, which demands innovative 54 technologies in aquaculture that can adjust to these challenges and promote sustainability in 55 the long term. 56

Brazilian aquaculture sector achieved economic relevance in the early 1980's, and currently holds the 8<sup>th</sup> position in the ranking of major fish aquaculture producers, with >600 thousand tonnes produced in 2018 (FAO, 2020; Valenti et al., 2020). Lambari (*Astyanax spp*) is an indigenous fish group commercialized as live-bait, which culture is growing very fast in Brazil. Production attained ~1000 t, and lambari farms ranked at 5<sup>th</sup> in number of aquaculture properties in 2019 (Valenti et al., 2020). Lambari culture is comparable to most kinds of smallscale land-based fish culture in Brazil. Thus, its technical advantages and disadvantages may

be an archetype of similar fish culture farms. Lambari is a group of native low-trophic level 64 freshwater fish species widely distributed in Brazil, and its production was initially considered 65 as a secondary product for additional income for small farmers in the southeast region. Due 66 to the expansion of the live bait market for sport fishing in the region, lambari production has 67 grown during the past decade. Its production occurs primarily in small aguaculture farms, 68 operating in semi-intensive earthen pond systems (Fonseca et al., 2017), but the market 69 expansion has attracted investors to implement larger farms (>20 ha) that operate under 70 higher demand for infrastructure and energy. 71

Several different management practices are used in the farming of lambari (Fonseca et 72 al., 2017). Producers settled their management choices empirically, or based on other 73 species culture protocols. Inadequate management practices are often observed, since the 74 amount of available scientific information and its access are insufficient to address the needs 75 of most producers (Fonseca et al., 2017). As a result, some producers face low productivity, 76 inefficient use of natural resources and generation of waste. Currently, lambari culture in 77 Brazil ranges from farms with no technical support and low control of feeding regime, survival 78 rate, and water flow to farms with gualified employees, monitoring equipment and indoor 79 hatcheries. Nevertheless, the technology applied in all farms rarely relies on scientific-based 80 information. 81

The absence of scientific-based protocols allied to the high variation in lambari culture 82 currently practiced make it a good model to study the sustainability of different practices in 83 freshwater fish culture. Furthermore, this situation claims for efforts in establishing more 84 efficient and sustainable lambari farm systems. The assessment of the main aspects that 85 drives sustainability would support the development of a scientific-based management 86 towards more sustainable lambari production. Emergy synthesis is a useful tool for assessing 87 bio-economic systems such as aquaculture (Agostinho et al., 2019; Bonilla et al., 2010; B.F. 88 Giannetti et al., 2011). This method allows the evaluation of the work done by nature, society 89 and economy on a common basis, identifying the main issues in a holistic way (Odum, 1996). 90 Therefore, this study applies emergy synthesis to assess the sustainability of lambari 91 aquaculture, providing a systemic perspective on the shortcomings of EAA related to the 92 sustainability of small freshwater aquaculture. Additionally, management practices that 93 negatively affect sustainability in semi-intensive pond freshwater systems were identified and 94 more sustainable alternatives were proposed. 95

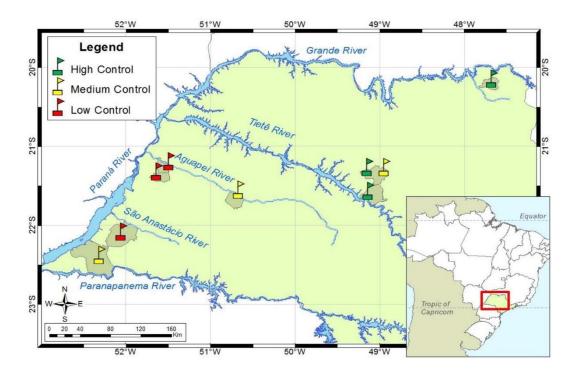
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#### 97 2. Materials and Methods

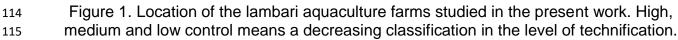
### 98 2.1. Data source and farms description

Nine lambari production farms were studied. They are located in the São Paulo State, 99 Brazil (Figure 1), a subtropical region. Farms were selected with the assistance of the São 100 Paulo State Rural Technical Assistance Agency. All farms produce the yellowtail lambari 101 (Astyanax lacustris) in semi-intensive earthen ponds, intensively fed with commercial feed. 102 Lambari farms differ for land and pond sizes, management strategies, and capital for 103 economic investments in infrastructure and equipment. These dissimilarities resulted in the 104 following three categories, or levels of control, according to systems technification degree: 105 low control, moderate control, and high control (Table 1). The farms were grouped 106 considering the breeding techniques used (natural, semi-natural or controlled), infrastructure 107 and equipment available, control and monitoring of water quality and supplied feed, and 108 survival rates. The process occurred according to the characteristics of production systems 109 and was validated by regional experts in lambari production that work in the national rural 110 offices. 111

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Data on natural and economic inputs, management practices and landscape features were obtained in situ, for each farm, through collecting samples of water, sediment, diet, and organisms that occurred in two visits at the beginning and end of one production cycle. Additional information was obtained through a semi-structured survey applied to the nine farmers in the beginning of the production cycle. The questionnaire focused on accounting for the total amount of materials, equipment, and infrastructure purchased, as well as labor, taxes, and depreciation. All inflows of materials, energy and money were accounted in unities/hectare, and they correspond to one year (i.e. 3 production cycles) of farm operation. Farmers validated the data collected at the end of the survey.

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#### 126 **2.2. Emergy synthesis procedure**

Data obtained from each farm were subjected to an emergy synthesis. Emergy is all the 127 energy directly and indirectly used to generate a product or a service (Odum, 1996). This 128 method is a biophysical approach based on a donor side perspective in establishing value 129 for natural resources. Thus, it considers a holistic view, which recognizes all the effort done 130 by nature in making available a resource. Moreover, as a donor side approach, emergy 131 synthesis avoids the inherent subjectivities of the receiver side analysis, such as the life cycle 132 assessment. The emergy synthesis procedure consists in three main steps: (i) elaborating 133 the energy diagram by defining system's boundaries, input and output flows, and their 134 relationship in internal processes (Figure 2); (ii) quantifying the main flows in the emergy 135 accounting table (i.e. inventory), choosing suitable unit emergy values (UEVs), and 136 calculating the emergy flows; (iii) calculating the emergy indicators to support comparisons 137 and discussions. In the present study, the system boundaries were the same as the farm 138 boundaries, including all local and external resources that sustain lambari aquaculture and 139 their interaction within the production system. All input resources were classified as natural 140 renewable resources (R), natural local non-renewable resources (N), or purchased resources 141 from the economy (F). Input resources were accounted in mass (g), energy (J) or money 142 (US\$) units, and correspond to one year of farm operation, at one hectare farm basis, allowing 143 comparisons between farms of different sizes. 144

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Table 1. Characteristics of the evaluated lambari aquaculture systems. Low control (LC), moderate control (MC) and high control (HC) management levels. N/A = not available.

Production factors	LC	MC	HC
Breeding/spawning	Natural without control	Hormone-induced inside the pond	Hormone-induced controlled hatchery
Production cycle (months)	4	4	4
Crops/year	3	3	3
Pond area (ha)	<1.5	1.5 - 6.2	>6.0 ha
Fertilization regime	Poultry manure	Poultry manure	Poultry manure and/or chemical fertilizer
Stocking seed	larvae	larvae	juvenile
Stocking density - Nursery phase	N/A	N/A	250
Stocking density - Grow-out phase	~9	17-25	~30
Pond water exchange (%/day)	3.7	7.0	5.8
Water source	Spring water	Spring water	Superficial water
Diet protein content (%)	28	32-56	32-56
Survival (%)	N/A	N/A	56
Final fish length (mm)	80.0	93.3	96.6
Final fish weight (g)	10	16	18
Productivity (t/ha)	1.8	6.1	6.9

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After quantified, the input resource flows were multiplied by their respective unit emergy 149 values (UEVs), resulting in flows of the same unity: solar emjoules (sej). All UEV's used in 150 this work (see Appendix A) were obtained from the scientific literature and the Emergy 151 Evaluation Folios published by the Center of Environmental Policy from the University of 152 Florida. The UEVs were updated to the global baseline of 1.20E+25 sej/yr (Brown and Ulgiati, 153 2016), and do not include labor and services, that were accounted separately as suggested 154 by Ulgiati and Brown (2014). Additionally, the partial renewability values for each resource 155 input were considered when available, as proposed by Agostinho et al. (2018). The sum of 156 the emergy flows in solar emjoules (sej) results on the total emergy demanded (Y). 157 Transformity is the UEV measured in sej/J, which is calculated by dividing the total emergy 158 demanded in sej (input) by the total yield measured in joules (output). The emergy indices 159 calculated in this work are presented in Table 2. 160

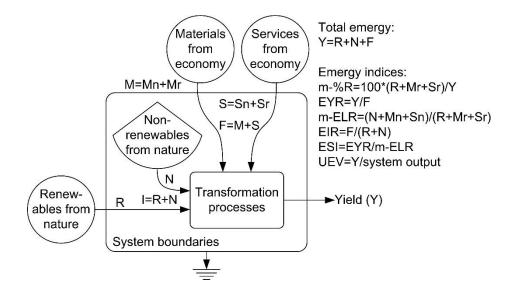


Figure 2. Generic energy diagram with symbols, acronyms and indicators used in emergy synthesis as presented in Table 2. Source: Agostinho et al. (2019).

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166	Table 2. Emergy indicators used in the present study.
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Emergy indicator	Algebra	Description
Unit emergy value	UEV = Y / E	Ratio of the total emergy demanded by the unity output. Example of units are sej/J, sej/kg and sej/\$.
Renewability <sup>a</sup>	m-%R = 100 (R+Mr+Sr) / Y	Ratio of the nature and economy's renewable fraction by the total emergy demanded to produce lambari.
Environmental loading ratio <sup>a</sup>	m-ELR = (N+Mn+Sn) / (R+Mr+Sr)	Ratio of the total non-renewable resources by the total renewable resources.
Emergy yield ratio	EYR = Y / F	Ratio of the total emergy demanded to produce lambari by the resources from economy.
Emergy investment ratio	EIR = F / (R+N)	Ratio of the resources from economy by the nature's renewable and non-renewable resources.
Emergy sustainability index	ESI = EYR / m-ELR	Ratio between the emergy yield ratio by the environmental loading ratio.

167 Source: Odum (1996)

<sup>168</sup> <sup>a</sup> Indicator modified according to Agostinho et al. (2018).

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A resource is defined as renewable when its natural replenishment rate is higher than its extraction rate. In this study, the spring water withdraw rate for LC and MC farm was compared with the natural recharge rate of the regional aquifer, where the farm is located. The natural recharge rate for the regional aquifer is about 25-27% of the yearly rainfall per hectare (CPRM, 2012), which is approximately ten times lower than the farms' withdraw rate.
 Therefore, spring water input was assumed as a non-renewable resource demanded by

aquaculture, as similarly considered by Cavalett et al. (2006), Wilfart et al. (2013), and Zhang
et al. (2012, 2011). The UEV of fish feed was estimated (Appendix B) based on a diet
formulated for lambari by Sussel et al. (2014).

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#### 180 **2.3. Better scenario**

181 . The variables seed production efficiency, water source, fertilization regime and protein 182 and oil source in feed formulation are key inflows for fish productivity. Therefore, they were 183 select to establish a better scenario for each lambari farm, wherein its effect on the emergy 184 performance were further assessed. The practices that comprise the scenario are following 185 described.

Practice 1. Improved seed productivity. This practice considers the introduction of 186 substrates inside the ponds used for reproduction to protected newly hatched larvae for LC 187 and MC farms. Lambari reproduction may be performed by hormone-induced spawning 188 inside indoor tanks, which allows higher larvae productivity, fish size homogeneity, stocking 189 density control, and survival rate (Felizardo et al., 2012) and should be maintained by HC 190 farms. Nevertheless, the economic investment on infrastructure, equipment and qualified 191 labor is often not affordable by the LC and MC farmers, even considering the higher profits. 192 As an alternative, the introduction of substrates inside ponds for natural reproduction is a low-193 cost technique that reduces larvae losses (Rezende et al., 2005). Since LC and MC farms 194 have lower control and monitoring of production, an increase in seed survival may result in a 195 higher final productivity. Therefore, a 25% increase in productivity for the LC and MC systems 196 197 is assumed as a plausible scenario as consequence of adopting this practice.

Practice 2. Changing water source. This practice includes the replacement of spring water by superficial water in the LC and MC systems, while the total water volume used remains the same. Therefore, the Unity Emergy Value (UEV) of the water source was replaced for LC and MC farms in the emergy accounting table (see Appendix A). The HC farms consumes superficial water, thus, this practice should be conserved.

203 *Practice 3. Controlling pond fertilization.* Fish nutrition is improved by the intake of 204 organisms from the natural biota existing inside ponds. Chemical or organic fertilizers are 205 inputs commonly applied in fish farms, but usually under improperly techniques that leads to 206 inefficiency. To support natural food production, the empirical practices as currently

performed by farmers is replaced by a controlled fertilization protocol. This practice establishes the use of 90 kg/1000m<sup>2</sup>/yr of lime, 56 kg/1000m<sup>2</sup>/yr of manure, 6.3 kg/1000m<sup>2</sup>/yr of urea and 2.3 kg/1000m<sup>2</sup>/year of phosphorus, as suggested by Pucher et al. (2014), and is accounted for all lambari farms.

Practice 4. Replace fish meal protein and fish oil by vegetable protein and oil sources. 211 This practice considers the total replacement of animal protein by vegetable sources in 212 commercial feed, following a diet formulated by Sussel et al. (2014) for lambari. Currently, 213 commercial feed used in lambari aquaculture relies on high protein contents that derives from 214 animal sources, such as marine fish meal and oil, and livestock byproducts such as viscera, 215 feather, and bones. Other components of commercial feed are mainly soy and corn. Fish oil 216 and fishmeal are environmentally costly as their consumption causes a pressure on the 217 natural stocks of marine fish. In addition, a diet that do not match the target fish requirements 218 will increase nutrient wastes in the effluents and consequently cause eutrophication in 219 receiving body of waters (Boyd et al., 2007; Flickinger et al., 2019b). Since lambari is a low-220 trophic level fish, the use of vegetable protein sources rather than animal ones is a feasible 221 alternative that does not affects productivity (Sussel et al., 2014). 222

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#### 224 **2.4. Approaches for results analysis**

Results analysis followed three approaches: (i) emergy index-by-index comparison 225 among the assessed nine-lambari farms considering the current and the simulated scenario 226 practices; (ii) the use of emergy ternary diagram; (iii) Emergy sustainability index and global 227 efficiency graph (ESI-UEV). The second and third methods are explained as follows. The 228 ternary diagram is an equilateral triangle, which the three corners represent each emergy 229 230 sources (R, N and F). Thus, any system plotted in the diagram is represented by a point, in which R, N and F can be determined by reading from zero along the basal line (axis) at the 231 bottom of the diagram to 100% at the vertex of the triangle (see Bonilla et al., 2010 and 232 Giannetti et al., 2011 for details). The emergy ternary diagram allows a visual comparison 233 between systems in terms of proportion for R, N and F emergy flows, and spatial 234 representation of system emergy performance (Almeida et al., 2007; Giannetti et al., 2006). 235 Besides lambari real and scenario data, nine different aquaculture systems assessed under 236 the emergy synthesis were obtained from the literature for further comparison. In the ESI-237 UEV graph, emergy sustainability indicator (ESI) and efficiency (the inverse of UEV) data for 238 each lambari system were plotted on a two-axis graph, in which larger ESI x UEV area 239

represents higher performance. Sustainability can be defined as an optimum balance between resilience and efficiency (Byrne, 2016). The emergy sustainability index (ESI, accounts for the total environmental pressure of the system over the biosphere capacity (a viewpoint of environmental resilience), and global efficiency (or inverse UEV) measures how efficient a system is for converting the emergy inflow into a product. Therefore, this graph aims to represent which lambari system have the best balance of both.

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#### 247 **3. Results**

#### 248 3.1. Lambari production under current practices

The energy system diagram (Figure 3) shows the lambari production features under the 249 systemic view of emergy synthesis. Most of the energy flows comes from outside the farms 250 boundaries, such as sun, rain, commercial feed, equipment, materials and labor. All the 251 lambari aquaculture systems evaluated in this study rely on similar external inputs and 252 internal processes, in which the differences are related to the amount and proportions for R, 253 N and F input resources demanded by each farm. Besides, the high (HC) and moderate (MC) 254 control level systems rely on external labor either permanent or eventual, while the low control 255 (LC) system relies on local family labor. Energy flows interact within system boundaries with 256 internal stocks of natural capital, hatchery (in the HC farms), and the pond, allowing the 257 production of lambari fish as the main output. Environmental services are co-products and 258 effluents are sub-products produced at different rates among farms. Overall, farms with lower 259 control over the production – divergent practices from the established by scientific-technical 260 protocols – and lower productivity demand lower emergy per hectare compared to the farms 261 with higher control (Table 3). The main inputs for the LC systems are services (27-46%), feed 262 (7-39%) and water (15-21%), while for the MC farms, the main inputs are feed (35-49%) and 263 services (33-39%), and for HC farms, the main flows are feed (17-48%) and services (26-264 36%). Unrelated to the management control level, all systems depend mostly on non-265 renewable sources, as purchased inputs F are responsible for more than 60% of the total 266 emergy required (Table 3). 267

The emergy indicators showed a random pattern among the evaluated farms (Tables 4, 5 and 6).. The HC1 farm shows the worst performance for UEV, achieving a value approximately 5 times higher than the farm with the best performance (HC2) (Table 6). The HC2 shows the best overall emergy performance among the studied farms, including the

- highest renewability (m-%R) and sustainability (ESI) and the lowest environmental loading
  (ELR) and emergy investment ratios (EIR). As well, EIR seems slightly lower in the LC farms
  (Table 4) compared to MCs and HC1 and HC3. Anyhow, all the lambari farms studied are
  strongly dependent on F resources, which means a low contribution to the larger economy
  (EIR > 1), and show an emergy sustainability index (ESI) bellow 1, which is an indicative of
  unsustainable systems.
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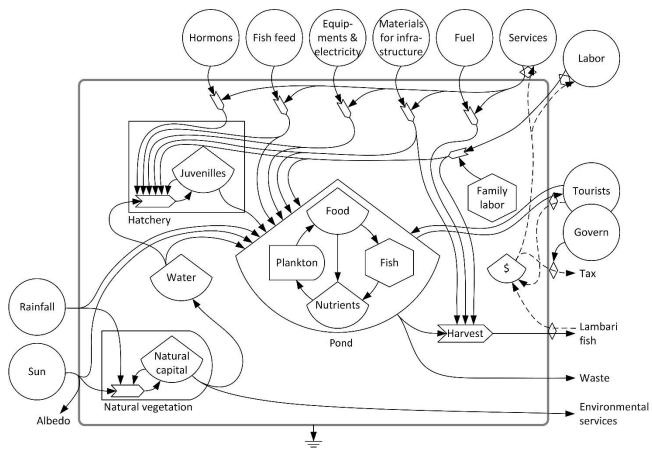


Figure 3. Energy diagram of lambari aquaculture production systems. Hatchery "box" is present only in high control farms (HC). Arrows represent energy flows, circles represents the outside sources, stocks are represented by tanks, and energy transformation processes represented by the interaction symbol; dashed arrows represent monetary flows; outputs are the harvested lambari, water effluent and environmental services. Symbol details in Odum (1996).

Table 3. Emergy accounting results in sej/ha/yr for the nine evaluated lambari aquaculture systems in Brazil. Low control (LC), moderate control (MC) and high control (HC) management levels. Numbers (1, 2 and 3) are the identification of different farms within a same control level. R, renewable resources from nature. N, non-renewable resources from nature. F, resources from the larger economy. Emergy columns presents the emergy flow from each item for each farm. Percentage columns (%) presents the emergy fraction of an item relative to the total emergy (Y) for each farm.

ltare	LC1		LC2		LC3		MC1		MC2		MC3		HC1		HC2		HC3	
Item	Emergy	%	Emergy	%	Emergy	%	Emergy	%	Emergy	%								
Sun (R)	4.67E+13	<1	4.67E+13	<1	4.67E+13	<1	4.67E+13	<1	4.67E+13	<1								
Rainfall (R)	2.12E+15	3	2.12E+15	2	2.12E+15	3	2.12E+15	1	2.12E+15	1	2.12E+15	1	1.36E+15	<1	1.99E+15	2	1.36E+15	1
Superficial water (R)	0.00E+00	0	0.00E+00	0	1.17E+16	3	1.42E+16	17	1.08E+16	4								
Soil occupation (N)	1.22E+15	2	8.01E+14	1	3.87E+14	1	3.33E+15	1	1.43E+15	1	2.10E+15	1	2.10E+16	5	3.30E+15	4	1.18E+15	<1
Groundwater (N)	1.41E+16	21	2.44E+16	19	1.03E+16	15	1.41E+16	5	1.48E+16	10	1.46E+16	9	0.00E+00	0	0.00E+00	0	0.00E+00	0
Feed (F)	4.77E+15	7	4.95E+16	39	1.87E+16	27	1.26E+17	49	5.43E+16	35	5.89E+16	37	2.19E+17	48	2.63E+16	31	4.02E+16	17
Equipment (F)	1.12E+13	<1	7.27E+14	1	2.13E+13	<1	1.20E+15	<1	1.38E+14	<1	5.25E+14	<1	2.69E+14	<1	1.80E+13	<1	5.89E+15	2
Electricity (F)	9.82E+12	<1	7.24E+14	1	1.92E+13	<1	1.20E+15	<1	1.38E+14	<1	5.21E+14	<1	2.69E+14	<1	1.80E+13	<1	5.89E+15	2
Infra-structure (F)	3.67E+12	<1	5.25E+12	<1	4.09E+12	<1	6.08E+12	<1	2.36E+12	<1	8.52E+12	<1	8.57E+12	<1	1.50E+12	<1	8.73E+12	<1
Lime (F)	4.66E+15	7	4.66E+15	4	4.66E+15	7	4.66E+15	2	4.66E+15	3	4.66E+15	3	4.66E+15	1	4.66E+15	6	4.66E+15	2
Organic fertilizer (F)	9.21E+15	13	9.21E+15	7	9.21E+15	13	9.21E+15	4	9.21E+15	6	9.21E+15	6	9.21E+15	2	9.21E+15	11	9.21E+15	4
Fuel (diesel) (F)	4.98E+14	1	8.62E+14	1	4.88E+14	1	7.32E+15	3	2.36E+15	2	8.79E+15	6	6.10E+15	1	1.46E+15	2	5.58E+16	23
Labor (F)	0.00E+00	0	4.76E+14	<1	0.00E+00	0	6.48E+15	3	5.22E+15	3	6.27E+15	4	1.70E+16	4	1.16E+15	1	2.93E+16	12
Services (F)	3.16E+16	46	3.40E+16	27	2.43E+16	35	8.41E+16	33	6.09E+16	39	5.16E+16	33	1.66E+17	36	2.15E+16	26	8.13E+16	34
Total emergy (Y)	6.83E+16		1.27E+17		7.02E+16		2.59E+17		1.55E+17		1.59E+17		4.57E+17		8.38E+16		2.40E+17	
Total (R) <sup>a</sup>	8.25E+15	13	1.18E+16	9	8.23E+15	12	2.44E+16	9	164E+16	11	1.56E+16	10	5.32E+16	12	2.24E+16	27	3.62E+16	15
Total (N)	1.54E+16	23	2.52E+16	20	1.07E+16	15	1.74E+16	7	1.63E+16	10	1.63E+16	11	2.10E+16	5	3.30E+15	4	1.18E+15	<1
Total (F)	4.43E+16	65	8.98E+16	71	1.07E+16	73	2.17E+17	84	1.23E+17	79	1.627E+17	80	3.82E+17	84	5.81E+16	69	2.02E+17	84

<sup>a</sup> Includes the flows of Sun, Rainfall, Superficial water and the renewable fraction from N and F flows.

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Table 4 Emergy indicators for the current management and the better scenario of lambari aquaculture low control farms. Current low control (LC) and better scenario for low control (LC'). Numbers (1, 2 and 3) are the identification of each different farm within a same control level. UEV = Unity emergy value; m-%R = renewable fraction; m-ELR = Environmental loading ratio; EYR = Emergy yield ratio; EIR = Emergy investment ratio; ESI = Emergy sustainability ratio.

Indicator	LC1	LC1'	LC2	LC2'	LC3	LC3'
UEV (E6 sej/J)	2.84	1.86	1.89	1.21	3.07	1.90
UEV (E10 sej/g)	4.88	3.19	4.23	2.72	7.02	4.34
UEV (E6 sej/J)*	1.53	0.79	1.38	0.80	2.01	1.04
UEV (E10 sej/g)*	2.62	1.37	3.09	1.79	4.59	2.37
m-%R (%)	13	37	9	33	12	31
m-ELR	6.9	1.7	9.8	2.0	7.5	2.2
EYR	1.3	1.4	1.3	1.3	1.2	1.3
EIR	3.9	3.4	4.6	4.0	5.5	4.5
ESI	0.2	0.8	0.1	0.7	0.2	0.6

\* without services.

Table 5. Emergy indicators for the current management and the better scenario of lambari aquaculture moderate control farms. Current moderate control (MC) and better scenario for moderate control (MC'). Numbers (1, 2 and 3) are the identification of each different farm within a same control level. UEV = Unity emergy value; m-%R = renewable fraction; m-ELR = Environmental loading ratio; EYR = Emergy yield ratio; EIR = Emergy investment ratio; ESI = Emergy sustainability ratio.

Indicator	MC1	MC1'	MC2	MC2'	MC3	MC3'
UEV (E6 sej/J)	1.55	1.02	0.91	0.61	2.17	1.44
UEV (E10 sej/g)	3.38	2.23	2.07	1.38	5.09	3.38
UEV (E6 sej/J)*	1.05	0.62	0.55	0.32	1.46	0.32
UEV (E10 sej/g)*	2.28	1.35	1.26	0.18	3.44	0.18
m-%R (%)	9	17	11	23	10	22
m-ELR	9.6	4.8	8.5	3.4	9.2	3.6
EYR	1.1	1.1	1.1	1.2	1.1	1.2
EIR	13.2	11.5	8.4	7.4	8.4	7.4
ESI	0.1	0.2	0.1	0.3	0.1	0.3

\* without services.

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3	т	Τ.

312	Table 6. Emergy indicators for the current management and the better
313	scenario of lambari aquaculture high control farms. Current high control
314	(HC) and better scenario for high control (HC'). Numbers (1, 2 and 3)
315	are the identification of each different farm within a same control level.
316	UEV = Unity emergy value; m-%R = renewable fraction; m-ELR =
317	Environmental loading ratio; EYR = Emergy yield ratio; EIR = Emergy
318	investment ratio; ESI = Emergy sustainability ratio.

Indicator	HC1	HC1'	HC2	HC2'	HC3	HC3'
UEV (E6 sej/J)	4.68	2.54	0.86	0.55	2.47	1.81
UEV (E10 sej/g)	10.3	5.62	2.09	1.33	4.97	3.63
UEV (E6 sej/J)*	2.97	1.45	0.64	0.37	1.63	1.13
UEV (E10 sej/g)*	6.58	3.20	1.56	0.89	3.28	2.27
m-%R (%)	12	14	27	32	15	16
m-ELR	7.6	6.3	2.7	2.1	5.6	5.2
EYR	1.1	1.1	1.3	1.4	1.1	1.1
EIR	13.4	11.4	4.3	3.4	18.0	16.4
ESI	0.1	0.2	0.5	0.7	0.2	0.2

\* without services

The ternary diagram (Figure 4a) shows the emergy performance of the nine evaluated lambari farms, compared with nine other aquaculture systems data obtained from literature. All systems are located very close to each other and to the F vertex, indicating a dependence on purchased resources (> 63%), which leads to an overall unsustainable performance (ESI<1).

<sup>319</sup> 

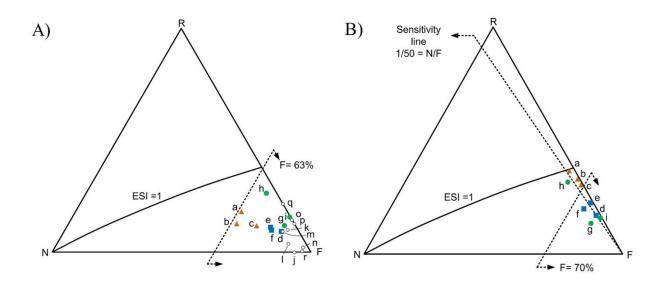


Figure 4. A) Ternary emergy diagram representing the proportions of renewable resources 328 (R), non-renewable resources (N) and resources from economy (F). Evaluated lambari 329 aquaculture systems in the present study were represented by ▲, ●, and ∎; data from 330 literature were represented by  $\circ$ . ESI = emergy sustainability index. a = LC1; b = LC2; c = 331 LC2; d = MC1; e = MC2; f = MC3; g = HC1; h = HC2; i = HC3; j = recirculating aguaculture332 system, k = extensive pond system, and l = semi-extensive system from Wilfart et al. (2013);333 m = integrated pig-grains-fish culture and n = semi-intensive fish pond system from Cavalett 334 et al. (2006); o = semi-intensive fish pond from Cheng et al. (2017); p = net-cage intensive 335 system and q = net-cage intensive system + bamboo substrate from David et al. (2018); r = 336 intensive fish pond from Zhang et al. (2011). B) Ternary diagram representing the proportions 337 of renewable resources (R), non-renewable resources (N) and resources from economy (F) 338 for lambari aquaculture systems after the simulated scenarios for better management 339 practices. Legend: LC systems ( $\blacktriangle$ ); MC systems ( $\blacksquare$ ); HC systems ( $\bullet$ ); ESI = emergy 340 sustainability index. A = LC1'; b = LC2'; c = LC3'; d = MC1'; e = MC2'; f = MC3'; g = HC1'; h 341 = HC2'; i = HC3'. 342

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#### 344 **3.2 Better scenario**

The simulated BMPs lead to an improvement of emergy performance for all evaluated lambari farms, including higher renewability and efficiency, while reducing the environmental loading ratio. The LC systems achieved the greatest improvements for renewability (between 164 and 255% increase), while reducing the ELRs (between 71 and 80% decrease) and

transformities (between 35 and 38% decrease) (Table 4). The MC (Table 5) and HC (Table 349 6) systems obtained an increase for renewability (in a range of 81-124% and 6-20% of 350 increase, respectively), and reduced their ELRs (in a range of 50-61% and 7-23% of 351 reduction) and transformities (in a range of 27-46% of reduction). Likewise, the ESI for all 352 farms were increased (LC increase of >269% from 0.2 to 0.7; MC between 78-142% from 0.1 353 to 0.3, and HC between 6-47% from 0.1 to 0.2 for HC1 and HC3, and from 0.5 to 0.7 for 354 HC2). Although the BMPs resulted in better emergy performance, all the evaluated farms still 355 remain below the ESI=1 line and, therefore, they are considered unsustainable. 356

As the proportions of R, N and F emergy flows changed after the simulated BMPs, the 357 spatial position of all farms also changed in the ternary diagram (Figure 4). Farms LC1', LC2', 358 LC3' and HC2' moved closer to the ESI=1 line compared to their relative position before 359 applying BMPs (Figure 4), as the proportion of renewable resources was increased. Although 360 increasing their renewability ratios (m-%R), the farms MC1', MC2', MC3', HC1' and HC3' 361 position remain distant from the ESI=1 and close to F vertex, resulted from the high 362 dependence (>70%) on F resources. A sensitivity line indicates that emergy sustainability for 363 the lambari production systems is improved by going in direction of R vertex, but the 364 proportion of 1/50 between N and F resources keeps approximately the same. When more 365 data becomes available, further studies can be developed to verify this tendency. 366

The Figure 5 indicates the systems with better performance for ESIxUEV balance. Three aspects deserves attention: (i) the simulated BMPs resulted in higher performance for all the evaluated systems; (ii) the LC systems achieved the highest improvement, as a result of the management improvements scenario; (iii) the system with the best balance of higher efficiency and environmental sustainability was the HC2', followed by LC2', MC2', HC2, LC1', LC3', MC1', MC3', MC2, HC3', HC3, HC1', MC1, LC2, LC1, MC3, LC3, HC1 in an hierarchical order.

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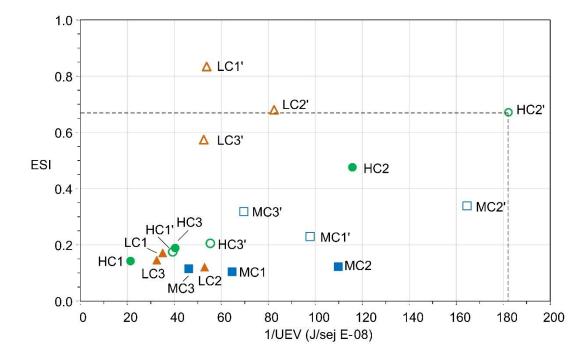


Figure 5. Emergy sustainability index (ESI) and global efficiency (inverse of unit emergy value) for the current management and proposed scenarios for the lambari production farms evaluated. Higher area means higher performance. The dashed line represents the area of the system with higher performance.  $LC = \blacktriangle$ ;  $LC' = \triangle$ ;  $MC = \blacksquare$ ;  $MC' = \Box$ ;  $HC = \bullet$ ;  $HC' = \circ$ . Different colors represent different farms within the same management level: orange = low control; blue = moderate control; green = high control.

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#### 383 **4. Discussion**

The lambari aquaculture systems evaluated in this study are dependent on similar 384 resources. Despite the existing similarities, farms show different emergy performances for 385 efficiency, renewability, environmental pressure and nature's investment, regardless of their 386 level of management control. The farms with higher control level (HC) achieved higher 387 productivity. Nevertheless, they are the most dependent on resources from the larger 388 economy (F). The HC2 farm had the best performance for all emergy indicators, surpassing 389 HC1 and HC3 although they have the same control level. The HC2 consumes lower amount 390 of commercial feed emergy (sej/ha/year) than HC1, more organic fertilizers than HC1 and 391 HC3 and reaches similar productivity than both. Therefore, HC2 represent a system with 392 more effective use of natural food. Conversely, the LC farms have the lowest productivity and 393 consume higher volumes of spring water per hectare, which is a local non-renewable 394 resource (N), leading to lower emergy performance. These findings imply that, when higher 395 the level of management control, higher the farm productivity but not necessarily higher 396

emergy performance or sustainability. For the evaluated farms under the current practices,
 feeding regimes and water management appears as key aspects for the emergy sustainability
 of lambari production. However, none farm achieved ESI higher than 1, indicating that all
 remain unsustainable production systems under the emergy view.

Commercial feed has been reported as the main emergy flow for fish production 401 systems (David et al., 2018; Garcia et al., 2014; Odum, 2000; Zhang et al., 2011). Intensive 402 feeding regimes may lead to a higher productivity, but large amounts of the commercial feed 403 consumed flows out with the water effluent, causing eutrophication in the surrounding water 404 405 streams (Boyd et al., 2020), or remain accumulated in pond bottom (David et al., 2017a, 2017b; Flickinger et al., 2020, 2019b, 2019a) causing issues to pond management. The 406 inefficient use commercial feed also cause the depletion of natural stocks of fish shoals and 407 other resources (soybean, corn, etc.) that sustain commercial feed production (Ahmed et al., 408 2019; Boyd et al., 2007). In the extreme opposite situation, those fish culture systems that 409 demands lower amount (< 20% of total emergy) or even do not use commercial feed are 410 recognized as unproductive and economically unfeasible (Wilfart et al., 2013; Zhang et al., 411 2011). These unfed fish farms require larger land areas and rely on chemical and organic 412 fertilizers for natural food production in the water ponds, which are their highest emergy input 413 flow (Zhang et al., 2012). Our results reinforces that fish feeding practice is environmental 414 costly and deserves attention for improvements. As an alternative, the replacement of fish 415 meal and oil by vegetal protein and oil sources is suggested for the culture of omnivorous 416 species such as lambari (Sussel et al., 2014), which would lead to a reduction of 27% in the 417 feed emergy flow. The results obtained in the present study suggest that replacing animal 418 raw materials by vegetal ones in the feed industry leads towards higher emergy sustainability 419 for fish production. 420

Water is another expressive emergy input flow for the evaluated LC and MC systems. 421 This high emergy value results from the large volumes demanded added to the high UEV 422 value for spring water. Climate change may affect water availability in near future, increasing 423 the risk of local conflicts triggered by water demand (Ahmed and Thompson, 2019; 424 Jamalimoghaddam et al., 2019). Extreme weather events, such as droughts, are expected to 425 occur in higher frequency and intensity. These events threatens groundwater stocks level 426 and total fresh water availability, causing production loss (Ahmed et al., 2019). Most of 427 lambari producers are considered small farmers, with low access to capital or technology to 428 deal with the environmental threats and social conflicts. Considering this scenario, the 429 replacement of spring water by superficial water sources improves the systems resilience 430

(Figure 5). Additionally, under a systemic perspective of the farms, producers could make
use of the high nutrient water stored within the ponds as fertilizers for the production of fruits
and vegetables within farms (Bosma and Verdegem, 2011), since many small farms diversify
their crops for subsistence and additional income.

The urgency for sustainable production systems is a well-established global concern 435 (United Nations, 2015), in which the challenge is how to combine the constant increasing 436 demand for highly productive systems within Earth's biocapacity. Systems traditionally known 437 as highly productive, such as the intensively fed monocultures, are strongly dependent on 438 fossil energy, a non-renewable resource. Conversely, the so-called extensive systems use 439 more space and do not take full advantage of the available free natural resources, failing in 440 competing with those intensive systems that are more efficient. Therefore, the idea of 441 sustainability as a contingent balance of efficiency and resilience seems to be more effective 442 than a linear advance towards a static state (Byrne, 2016). Thus, combining emergy 443 sustainability (ESI) with efficiency (inverse UEV), as shown in Figure 5, provides a better 444 image of the sustainability path of the assessed lambari systems. 445

Although sustainability is context-dependent and requires constant adaptation, the HC2 446 farm showed most successful balance of emergy sustainability and efficiency among all the 447 evaluated fish production systems under current practices. Moreover, the simulated scenario 448 for BMPs showed improvements for all systems. This indicates that aquaculture technologies 449 designed with an ecological approach are likely to succeed on the long term, as they can 450 achieve higher performance through simple and feasible changes in their management 451 procedures. As one example, integrated multi-trophic aquaculture (IMTA) is a framework that 452 allows the culture of aquatic species from different trophic levels and with complementary 453 ecosystem functions (Ridler et al., 2007). IMTAs goal is to increase the productivity by using 454 the uneaten feed, wastes, and considering by-products as fertilizers and feed as energy 455 sources for other crops, taking advantage of the synergistic interactions among species 456 (Flickinger et al., 2019b; Franchini et al., 2020). For a more sustainable fish production, 457 systems dependence on fossil energy must be reduced at the same time their productivity is 458 increased, and this can be achieved by more ecologically driven and integrated production 459 systems. 460

All lambari production systems evaluated in this work uses similar kinds of external resources; nevertheless, lack of efficiency and dependence on F resources affects their emergy performance. The access of high quality information from technology-transfer and research centers is important to increase productivity under lower consumption of non-

renewable resources. Accurate staff training, technical assistance, personal relationship 465 between the aquaculture producers and research institutes, organization of producers in 466 cooperatives or associations, and effective government regulation are key factors towards a 467 more sustainable aquaculture systems (Boyd et al., 2020; Henares et al., 2019). When the 468 'information input' is neglected or inappropriately used, the system operates inefficiently. 469 Although hardly accounted, information is a costly emergy input for many kinds of natural and 470 human systems, but its use guarantees systems self-organization capacity towards higher 471 emergy efficiency (Odum, 1996). In the Brazilian aquaculture sector, the information sharing 472 process has been underrated, which has led to inefficient production systems, resulting in 473 high environmental pressures and low economic returns (Henares et al., 2019). Therefore, 474 technology transfer is a strategic step, perhaps the most important one, towards more 475 sustainable fish production systems. 476

477

#### 478 **5.** Conclusion

All lambari production systems studied rely mostly on non-renewable resources, mainly 479 on commercial feed and water, regardless of the control level (low, moderate or high). The 480 emergy performance of all farms were similar, with slight advantage for the high control (HC) 481 farm #2. This variation may be a result of accessibility to high-quality information on 482 production management. Excepted by HC farm #2, the low renewability (m-%R<15%), high 483 environmental load (ELR>5.6) and low emergy yield ratio (EYR<1.3) indicates that the 484 systems are unsustainable (ESI<0.2). Nevertheless, a scenario of simple and feasible 485 management practices including water-source change, control of pond fertilization, increase 486 of productivity by breeding management, and the exchange of animal protein and oil sources 487 by vegetal ones, results in higher emergy performance for all farms. Although the emergy 488 sustainability of the proposed scenario is still low (ESI<0.8) due to the high demand for 489 purchased resources (EYR<1.4), their renewability increased (m-%R<33%) along with a 490 reduction of the environmental loading ratio (ELR>1.7), indicating that the proposed practices 491 provide benefits under an emergy perspective. Additionally, the scenario increased systems 492 resilience , expressed by the emergy sustainability index x global efficiency (ESIxUEV) 493 relationship. Efforts are still needed towards better management practices on the lambari 494 aquaculture production, but the findings of this work highlight that simple changes make a 495 difference on the sustainability of small-scale inland aquaculture. This conclusion is 496

- 497 strengthened by the use of the emergy synthesis, which is a holistic approach in assessing
- <sup>498</sup> sustainability that recognizes the effort of nature in providing resources.

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## 504 **References**

- Agostinho, F., Oliveira, M.W., Pulselli, F.M., Almeida, C.M.V.B., Giannetti, B.F., 2019.
   Emergy accounting as a support for a strategic planning towards a regional sustainable
   milk production. Agric. Syst. 176, 102647. https://doi.org/10.1016/j.agsy.2019.102647
   Agostinho, F., Sevegnani, F., Almeida, C.M.V.B., Giannetti, B.F., 2018. Exploring the
- potentialities of emergy accounting in studying the limits to growth of urban systems.
   Ecol. Indic. 94, 4–12. https://doi.org/10.1016/j.ecolind.2016.05.007
- Ahmed, N., Thompson, S., 2019. The blue dimensions of aquaculture: A global synthesis.
   Sci. Total Environ. 652, 851–861. https://doi.org/10.1016/j.scitotenv.2018.10.163
- Ahmed, N., Thompson, S., Glaser, M., 2019. Global Aquaculture Productivity,
   Environmental Sustainability, and Climate Change Adaptability. Environ. Manage. 63,
   159–172. https://doi.org/10.1007/s00267-018-1117-3
- Ahmed, N., Ward, J.D., Thompson, S., Saint, C.P., Diana, J.S., 2018. Blue-Green Water
   Nexus in Aquaculture for Resilience to Climate Change. Rev. Fish. Sci. Aquac. 26,
   139–154. https://doi.org/10.1080/23308249.2017.1373743
- Almeida, C.M.V.B., Barrella, F.A., Giannetti, B.F., 2007. Emergetic ternary diagrams: five
   examples for application in environmental accounting for decision-making. J. Clean.
   Prod. 15, 63–74. https://doi.org/10.1016/j.jclepro.2005.07.002
- Antonucci, F., Costa, C., 2020. Precision aquaculture: a short review on engineering innovations. Aquac. Int. 28, 41–57. https://doi.org/10.1007/s10499-019-00443-w
- Béné, C., Arthur, R., Norbury, H., Allison, E.H., Beveridge, M., Bush, S., Campling, L.,
  Leschen, W., Little, D., Squires, D., Thilsted, S.H., Troell, M., Williams, M., 2016.
  Contribution of Fisheries and Aquaculture to Food Security and Poverty Reduction:
  Assessing the Current Evidence. World Dev. 79, 177–196.
- 528 https://doi.org/10.1016/j.worlddev.2015.11.007
- Bonilla, S.H., Guarnetti, R.L., Almeida, C.M.V.B., Giannetti, B.F., 2010. Sustainability
  assessment of a giant bamboo plantation in Brazil: exploring the influence of labour,
  time and space. J. Clean. Prod. 18, 83–91.
- 532 https://doi.org/10.1016/j.jclepro.2009.07.012
- Bosma, R.H., Verdegem, M.C.J., 2011. Sustainable aquaculture in ponds: Principles,
   practices and limits. Livest. Sci. 139, 58–68. https://doi.org/10.1016/j.livsci.2011.03.017
- Boyd, C.E., D'Abramo, L.R., Glencross, B.D., Huyben, D.C., Juarez, L.M., Lockwood, G.S.,
   McNevin, A.A., Tacon, A.G.J., Teletchea, F., Tomasso, J.R., Tucker, C.S., Valenti,
- 537 W.C., 2020. Achieving sustainable aquaculture: Historical and current perspectives and 538 future needs and challenges. J. World Aquac. Soc. 51, 578–633.

https://doi.org/10.1111/jwas.12714 539 Boyd, C.E., Tucker, C., Mcnevin, A., Bostick, K., Clay, J., 2007. Indicators of Resource Use 540 Efficiency and Environmental Performance in Fish and Crustacean Aquaculture. Rev. 541 Fish. Sci. 15, 327-360. https://doi.org/10.1080/10641260701624177 542 Brandt-Williams, S., 2002. Handbook of Emergy Evaluation Folio 4: Emergy of Florida 543 Agriculture. Center for Environmental Policy, University of Florida, Gainesville. 544 Brown, M.T., Bardi, E. (Eds.), 2001. Handbook of Emergy Evaluation A Compendium of 545 Data for Emergy Computation Issued in a Series of Folios Folio # 3 Emergy of 546 Ecosystems. Environmental Engineering Sciences Center for Environmental Policy 547 University of Florida Gainesville, Gainesville. 548 Brown, M.T., McClanahan, T.R., 1996. EMergy analysis perspectives of Thailand and 549 Mekong River dam proposals. Ecol. Modell. 91, 105-130. https://doi.org/10.1016/0304-550 3800(95)00183-2 551 Brown, M.T., Protano, G., Ulgiati, S., 2011. Assessing geobiosphere work of generating 552 global reserves of coal, crude oil, and natural gas. Ecol. Modell. 222, 879-887. 553 https://doi.org/https://doi.org/10.1016/j.ecolmodel.2010.11.006 554 555 Brown, M.T., Ulgiati, S., 2016. Assessing the global environmental sources driving the geobiosphere: A revised emergy baseline. Ecol. Modell. 339, 126-132. 556 https://doi.org/10.1016/j.ecolmodel.2016.03.017 557 Brown, M.T., Ulgiati, S., 2004. Emergy Analysis and Environmental Accounting, in: 558 Cleveland, C.J. (Ed.), Encyclopedia of Energy. Boston, MA, USA, pp. 329–354. 559 https://doi.org/https://doi.org/10.1016/B0-12-176480-X/00242-4 560 Brugère, C., Aguilar-Manjarrez, J., Beveridge, M.C.M., Soto, D., 2019. The ecosystem 561 approach to aquaculture 10 years on – a critical review and consideration of its future 562 role in blue growth. Rev. Aquac. 11, 493-514. https://doi.org/10.1111/rag.12242 563 Buenfil, A.A., 2001. Emergy evaluation of water. University of Florida. 564 Buranakarn, V., 1998. Evaluation of recycling and reuse of building materials using the 565 emergy analysis method. University of Florida, Gainesville, USA. 566 Byrne, E., 2016. Transdisciplinary Perspectives on Transitions to Sustainability, 1st ed. 567 Transdisciplinary Perspectives on Transitions to Sustainability, Routledge, Farnham, 568 Surrey, UK; Burlington, VT: Ashgate, 2016. |. https://doi.org/10.4324/9781315550206 569 Castellini, C., Bastianoni, S., Granai, C., Bosco, A.D., Brunetti, M., 2006. Sustainability of 570 poultry production using the emergy approach: Comparison of conventional and 571 organic rearing systems. Agric. Ecosyst. Environ. 114, 343-350. 572 https://doi.org/https://doi.org/10.1016/j.agee.2005.11.014 573 574 Cavalett, O., Queiroz, J.F. De, Ortega, E., 2006. Emergy assessment of integrated production systems of grains, pig and fish in small farms in the South Brazil. Ecol. 575 Modell. 193, 205-224. https://doi.org/10.1016/j.ecolmodel.2005.07.023 576 Cheng, H., Chen, C., Wu, S., Mirza, Z.A., Liu, Z., 2017. Emergy evaluation of cropping, 577 poultry rearing, and fish raising systems in the drawdown zone of Three Gorges 578 Reservoir of China. J. Clean. Prod. 144, 559-571. 579 https://doi.org/10.1016/j.jclepro.2016.12.053 580 Cohen, M.J., Sweeney, S., Brown, M.T., 2007. Computing the unit emergy value of crustal 581 elements., in: Brown, M., Bardi, E., Campbell, D.E., Comar, V., Huang, S., Rydberg, 582 T., Tilley, D. and, Ulgiat, S. (Eds.), Emergy Synthesis 4: Theory and Applications of the 583 Emergy Methodology. Proceedings of the 4th Biennial Emergy Conference. Center for 584 Environmental Policy, University of Florida, Gainesville, p. 483. 585 Comar, V., 2001. Emergy Evaluation of Organic and Conventional Horticultural Production 586 in Botucatu, Sao Paulo State, Brazil, in: Brown, M.T. (Ed.), Emergy Synthesis 1: 587 Theory and Applications of the Emergy Methodology. Proceedings of the 1st Biennial 588

589	Emergy Conference. Center for Environmental Policy, University of Florida,
590	Gainesville, pp. 181–195.
591	David, F.S., Proença, D.C., Valenti, W.C., 2017a. Phosphorus Budget in Integrated
592	Multitrophic Aquaculture Systems with Nile Tilapia, Oreochromis niloticus, and Amazon
593	River Prawn, Macrobrachium amazonicum. J. World Aquac. Soc.
594	https://doi.org/10.1111/jwas.12404
595	David, F.S., Proença, D.C., Valenti, W.C., 2017b. Nitrogen budget in integrated aquaculture
596	systems with Nile tilapia and Amazon River prawn. Aquac. Int.
597	https://doi.org/10.1007/s10499-017-0145-y
598	David, L.H.C., Pinho, S.M., Garcia, F., 2018. Improving the sustainability of tilapia cage
599	farming in Brazil: An emergy approach. J. Clean. Prod. 201, 1012–1018.
600	https://doi.org/10.1016/j.jclepro.2018.08.124
601	Dawood, M.A.O., Koshio, S., Abdel-Daim, M.M., Van Doan, H., 2019. Probiotic application
602	for sustainable aquaculture. Rev. Aquac. 11, 907–924.
603	https://doi.org/10.1111/raq.12272
604	Dong, X., Ulgiati, S., Yan, M., Zhang, X., Gao, W., 2008. Energy and eMergy evaluation of bioethanol production from wheat in Henan Province, China. Energy Policy 36, 3882–
605	3892. https://doi.org/10.1016/j.enpol.2008.04.027
606	FAO, 2020. The State of World Fisheries and Aquaculture 2020. Sustainability in action.
607 608	FAO, Rome. https://doi.org/10.4060/ca9229en
608	Felizardo, V.O., Murgas, L.D.S., Andrade, E.S., López, P.A., Freitas, R.T.F., Ferreira, M.R.,
610	2012. Effect of timing of hormonal induction on reproductive activity in lambari
611	(Astyanax bimaculatus). Theriogenology 77, 1570–1574.
612	https://doi.org/10.1016/j.theriogenology.2011.11.025
613	Flickinger, D.L., Costa, G.A., Dantas, D.P., Proença, D.C., David, F.S., Durborow, R.M.,
614	Moraes-Valenti, P., Valenti, W.C., 2020. The budget of carbon in the farming of the
615	Amazon river prawn and tambaqui fish in earthen pond monoculture and integrated
616	multitrophic systems. Aquac. Reports. https://doi.org/10.1016/j.aqrep.2020.100340
617	Flickinger, D.L., Costa, G.A., P. Dantas, D., Moraes-Valenti, P., Valenti, W.C., 2019a. The
618	budget of nitrogen in the grow-out of the Amazon river prawn (Macrobrachium
619	amazonicum Heller) and tambaqui (Colossoma macropomum Cuvier) farmed in
620	monoculture and in integrated multitrophic aquaculture systems. Aquac. Res.
621	https://doi.org/10.1111/are.14304
622	Flickinger, D.L., Dantas, D.P., Proença, D.C., David, F.S., Valenti, W.C., 2019b.
623	Phosphorus in the culture of the Amazon river prawn (Macrobrachium amazonicum)
624	and tambaqui (Colossoma macropomum) farmed in monoculture and in integrated
625	multitrophic systems. J. World Aquac. Soc. jwas.12655.
626	https://doi.org/10.1111/jwas.12655
627	Fonseca, T., Costa-Pierce, B.A., Valenti, W.C., 2017. Lambari Aquaculture as a Means for
628	the Sustainable Development of Rural Communities in Brazil. Rev. Fish. Sci. Aquac.
629	25, 316–330. https://doi.org/10.1080/23308249.2017.1320647
630	Franchini, A.C., Costa, G.A., Pereira, S.A., Valenti, W.C., Moraes-Valenti, P., 2020.
631	Improving production and diet assimilation in fish-prawn integrated aquaculture, using
632	iliophagus species. Aquaculture 521, 735048.
633	https://doi.org/10.1016/j.aquaculture.2020.735048
634	Garcia, F., Kimpara, J.M., Valenti, W.C., Ambrosio, L.A., 2014. Emergy assessment of
635	tilapia cage farming in a hydroelectric reservoir. Ecol. Eng. 68, 72–79.
636	https://doi.org/10.1016/j.ecoleng.2014.03.076
637	Garlock, T., Asche, F., Anderson, J., Bjørndal, T., Kumar, G., Lorenzen, K., Ropicki, A.,
638	Smith, M.D., Tveterås, R., 2020. A Global Blue Revolution: Aquaculture Growth Across

639Regions, Species, and Countries. Rev. Fish. Sci. Aquac. 28, 107–116.640https://doi.org/10.1080/23308249.2019.1678111

- Giannetti, B.F., Agostinho, F., Moraes, L.C., Almeida, C.M.V.B., Ulgiati, S., 2015.
   Multicriteria cost-benefit assessment of tannery production: The need for breakthrough
   process alternatives beyond conventional technology optimization. Environ. Impact
   Assess. Rev. 54, 22–38. https://doi.org/10.1016/j.eiar.2015.04.006
- Giannetti, B.F., Barrella, F.A., Almeida, C.M.V.B., 2006. A combined tool for environmental
   scientists and decision makers: ternary diagrams and emergy accounting. J. Clean.
   Prod. 14, 201–210. https://doi.org/10.1016/j.jclepro.2004.09.002
- Giannetti, B.F., Faria, L., Almeida, C.M.V.B., Agostinho, F., Coscieme, L., Liu, G., 2018.
   Human-nature nexuses in Brazil: Monitoring production of economic and ecosystem services in historical series. Ecosyst. Serv. 30, 248–256.
- 651 https://doi.org/https://doi.org/10.1016/j.ecoser.2017.10.008
- Giannetti, B.F., Ogura, Y., Bonilla, S.H., Almeida, C.M.V.B., 2011. Accounting emergy flows
   to determine the best production model of a coffee plantation. Energy Policy 39, 7399–
   7407. https://doi.org/10.1016/j.enpol.2011.09.005
- Giannetti, B. F., Ogura, Y., Bonilla, S.H., Almeida, C.M.V.B., 2011. Emergy assessment of
  a coffee farm in Brazilian Cerrado considering in a broad form the environmental
  services, negative externalities and fair price. Agric. Syst. 104, 679–688.
  https://doi.org/10.1016/j.agsy.2011.08.001
- Goddard, S., Delghandi, M., 2020. Importance of the Conservation and Management of
   Freshwater to Aquaculture, in: Encyclopedia of the World's Biomes. Elsevier, pp. 35–
   44. https://doi.org/10.1016/B978-0-12-409548-9.11954-2
- Gonçalves Lima, J.S., Rivera, E.C., Focken, U., 2012. Emergy evaluation of organic and
   conventional marine shrimp farms in Guaraíra Lagoon, Brazil. J. Clean. Prod. 35, 194–
   202. https://doi.org/10.1016/j.jclepro.2012.05.009
- Goswami, R., Saha, S., Dasgupta, P., 2017. Sustainability assessment of smallholder farms
   in developing countries. Agroecol. Sustain. Food Syst. 41, 546–569.
   https://doi.org/10.1080/21683565.2017.1290730
- Henares, M.N.P., Medeiros, M. V., Camargo, A.F.M., 2019. Overview of strategies that
   contribute to the environmental sustainability of pond aquaculture: rearing systems,
   residue treatment, and environmental assessment tools. Rev. Aquac. 1–18.
   https://doi.org/10.1111/raq.12327
- Humphries, F., Benzie, J.A.H., Morrison, C., 2019. A systematic quantitative literature
   review of aquaculture genetic resource access and benefit sharing. Rev. Aquac. 11,
   1133–1147. https://doi.org/10.1111/raq.12283
- Jamalimoghaddam, E., Yazdani, S., Salami, H., Peykani, G., 2019. The impact of water
   supply on farming systems : A sustainability assessment. Sustain. Prod. Consum. 17,
   269–281. https://doi.org/10.1016/j.spc.2018.11.001
- Jung, S., Rasmussen, L.V., Watkins, C., Newton, P., Agrawal, A., 2017. Brazil's National
   Environmental Registry of Rural Properties: Implications for Livelihoods. Ecol. Econ.
   136, 53–61. https://doi.org/10.1016/j.ecolecon.2017.02.004
- Lulijwa, R., Rupia, E.J., Alfaro, A.C., 2019. Antibiotic use in aquaculture, policies and
   regulation, health and environmental risks: a review of the top 15 major producers.
   Rev. Aquac. 2000, 1–24. https://doi.org/10.1111/raq.12344
- Mungkung, R., Aubin, J., Prihadi, T.H., Slembrouck, J., Van Der Werf, H.M.G., Legendre,
   M., 2013. Life cycle assessment for environmentally sustainable aquaculture
   management: A case study of combined aquaculture systems for carp and tilapia. J.
- 687 Clean. Prod. 57, 249–256. https://doi.org/10.1016/j.jclepro.2013.05.029
- Nobile, A.B., Cunico, A.M., Vitule, J.R.S., Queiroz, J., Vidotto-Magnoni, A.P., Garcia,

D.A.Z., Orsi, M.L., Lima, F.P., Acosta, A.A., da Silva, R.J., do Prado, F.D., Porto-689 Foresti, F., Brandão, H., Foresti, F., Oliveira, C., Ramos, I.P., 2019. Status and 690 recommendations for sustainable freshwater aquaculture in Brazil. Rev. Aquac. 1-23. 691 https://doi.org/10.1111/rag.12393 692 Odum, H.T., 2000. Emergy evaluation of salmon pen culture. Int. Inst. Fish. Econ. Trade 693 2000 Proc. 1-8. 694 Odum, H.T. (Ed.), 1996. Environmental Accounting: Emergy and Environmental Decision 695 Making, 1st ed. John Wiley & Sons, New York. 696 Pucher, J., Mayrhofer, R., El-Matbouli, M., Focken, U., 2014. Pond management strategies 697 for small-scale aquaculture in northern Vietnam: fish production and economic 698 performance. Aquac. Int. 23, 297-314. https://doi.org/10.1007/s10499-014-9816-0 699 Rezende, F.P., Filho, O.P.R., Pereira, M.M., Takabatake, É.Y., Navarro, R.D., Santos, L.C., 700 Silva, R.F., Filho, C.B.C., 2005. EFICIÊNCIA DE DIFERENTES SUBSTRATOS NA 701 DESOVA DE LAMBARI TAMBIÚ (astvanax bimaculatus LINNAEUS, 1758). Rev. 702 Ceres. 703 Ridler, N., Wowchuk, M., Robinson, B., Barrington, K., Chopin, T., Robinson, S., Page, F., 704 705 Reid, G., Szemerda, M., Sewuster, J., Boyne-Travis, S., 2007. Integrated multi-trophic aquaculture (IMTA): A potential strategic choice for farmers. Aquac. Econ. Manag. 11, 706 99-110. https://doi.org/10.1080/13657300701202767 707 Shi, H., Zheng, W., Zhang, X., Zhu, M., Ding, D., 2013. Ecological-economic assessment of 708 monoculture and integrated multi-trophic aquaculture in Sanggou Bay of China. 709 Aquaculture 410-411, 172-178. https://doi.org/10.1016/j.aquaculture.2013.06.033 710 Soto, D., Aguilar-Manjarrez, J., Hishamunda, N. (Eds.), 2008. Building an ecosystem 711 approach to aquaculture., in: FAO Fisheries and Aquaculture Proceedings. FAO, 712 Rome, p. 221. 713 Sussel, F.R., Viegas, E.M.M., Evangelista, M.M., Gonçalves, G.S., Salles, F.A., Gonçalves, 714 L.U., 2014. Substituição de proteína animal por proteína vegetal em dietas para 715 lambari-do-rabo- amarelo astyanax altiparanae. Acta Sci. - Anim. Sci. 36, 343-348. 716 https://doi.org/10.4025/actascianimsci.v36i4.23836 717 Tacon, A.G.J., 2020, Trends in Global Aquaculture and Aquafeed Production: 2000–2017. 718 Rev. Fish. Sci. Aquac. 28, 43-56. https://doi.org/10.1080/23308249.2019.1649634 719 Takahashi, F., Ortega, E., 2010. Assessing the sustainability of Brazilian oleaginous crops 720 - possible raw material to produce biodiesel. Energy Policy 38, 2446-2454. 721 https://doi.org/doi:10.1016/j.enpol.2009.12.038 722 Ulgiati, S., Brown, M.T., 2014. Labor and services as information carriers in emergy-LCA 723 724 accounting. J. Environ. Account. Manag. 2, 163–170. https://doi.org/10.5890/JEAM.2014.06.006 725 United Nations, 2015. Transforming our World: the 2030 Agenda for Sustainable 726 Development. 727 Valenti, W.C., Barros, H.P., Moraes-Valenti, P., Cavalli, R.O., 2020. Aquaculture in Brazil: 728 past, present and future. Aquac. Reports in press. 729 Valenti, W.C., Kimpara, J.M., Preto, B. de L., Moraes-Valenti, P., 2018. Indicators of 730 sustainability to assess aquaculture systems. Ecol. Indic. 88, 402-413. 731 https://doi.org/10.1016/j.ecolind.2017.12.068 732 Vassallo, P., Bastianoni, S., Beiso, I., Ridolfi, R., Fabiano, M., 2007. Emergy analysis for 733 the environmental sustainability of an inshore fish farming system. Ecol. Indic. 7, 290-734 298. https://doi.org/10.1016/j.ecolind.2006.02.003 735 Verdegem, M.C.J.J., Bosma, R.H., Verreth, J.A.J.J., 2006. Reducing water use for animal 736 production through aquaculture. Int. J. Water Resour. Dev. 22, 101–113. 737 https://doi.org/10.1080/07900620500405544 738

- Wilfart, A., Prudhomme, J., Blancheton, J.P., Aubin, J., 2013. LCA and emergy accounting
   of aquaculture systems: Towards ecological intensification. J. Environ. Manage. 121,
   96–109. https://doi.org/10.1016/j.jenvman.2013.01.031
- Zhang, L.X., Song, B., Chen, B., 2012. Emergy-based analysis of four farming systems:
- Insight into agricultural diversification in rural China. J. Clean. Prod. 28, 33–44.
   https://doi.org/10.1016/j.jclepro.2011.10.042
- Zhang, L.X., Ulgiati, S., Yang, Z.F., Chen, B., 2011. Emergy evaluation and economic
   analysis of three wetland fish farming systems in Nansi Lake area, China. J. Environ.
   Manage. 92, 683–694. https://doi.org/10.1016/j.jenvman.2010.10.005
- Zhao, S., Song, K., Gui, F., Cai, H., Jin, W., Wu, C., 2013. The emergy ecological footprint
   for small fish farm in China. Ecol. Indic. 29, 62–67.
- 750 https://doi.org/10.1016/j.ecolind.2012.12.009

Appendix A. Inventory and unit emergy values (UEV) for the nine evaluated lambari production systems. Legend: Low control (LC), moderate control (MC) and high control (HC) management levels. Numbers (1, 2 and 3) represent different farms within the same control level. R, renewable resources from nature; N, non-renewable resources from nature; F, resources from the larger economy; %R, renewability fraction in %. Calculation details in the Supplementary Material.

Item and its	Unit	UEV <sup>a</sup>	Deference for UEV	0/ D	Amount in Unit/ha/yr								
classification	Unit	(sej/Unit)	Reference for UEV	%R	LC1	LC2	LC3	MC1	MC2	MC3	HC1	HC2	HC3
1. Sun (R)	J	1.00E+00	Odum, 1996	100	4.67E+13	4.67E+13	4.67E+13	4.67E+13	4.67E+13	4.67E+13	4.67E+13	4.67E+13	4.67E+13
2. Rainfall (R)	J	2.31E+04	Odum, 1996	100	9.16E+10	9.16E+10	9.16E+10	9.16E+10	9.16E+10	9.16E+10	9.16E+10	9.16E+10	9.16E+10
<ol> <li>Superficial water (R)</li> </ol>	J	5.23E+04	Comar, 2001	100	0.00E+00	0.00E+00	0.00E+00	0.00E+00	0.00E+00	0.00E+00	0.00E+00	0.00E+00	0.00E+00
<ol> <li>Soil occupation (N)</li> </ol>	J	9.42E+04	Brandt-Williams, 2002	0	1.30E+10	8.50E+09	8.50E+09	8.50E+09	1.52E+10	1.52E+10	1.52E+10	1.52E+10	1.52E+10
5. Springwater (N)	J	5.63E+04	Buenfil, 2001	0	2.51E+11	4.34E+11	4.34E+11	4.34E+11	2.63E+11	2.63E+11	2.63E+11	2.63E+11	2.63E+11
<ol> <li>Feed (F)</li> <li>Equipment (F)</li> </ol>	g	7.01E+09	Appendix B	5	6.80E+05	7.06E+06	7.06E+06	7.06E+06	7.74E+06	7.74E+06	7.74E+06	7.74E+06	7.74E+06
7.1 Iron	g	7.63E+10	Buranakarn, 1998	0	3.40E+00	5.88E+00	3.33E+00	2.00E+01	1.61E+00	4.96E+01	8.33E+00	3.75E-01	2.00E+01
7.2 Plastic	g	3.90E+09	Buranakarn, 1998	0	3.82E+00	1.36E+01	6.69E+00	2.21E+01	4.69E+00	5.39E+01	1.73E+01	1.25E+00	1.93E+01
7.3 Steel	g	5.92E+09	Brown and Ulgiati, 2004	0	1.79E+02	4.95E+02	3.02E+02	1.08E+01	5.81E-01	2.67E+01	1.08E+01	1.33E+00	1.08E+01
7.4 Aluminum	g	1.62E+10	Buranakarn, 1998	0	6.80E-03	1.00E-02	6.67E-03	3.30E-01	1.61E-03	8.18E-01	3.30E-01	3.95E-03	3.30E-01
7.5 Glass fiber	g	1.00E+10	Buranakarn, 1998	0	0.00E+00	0.00E+00	0.00E+00	7.00E-01	0.00E+00	1.74E+01	2.33E+00	3.85E-02	7.00E+00
8. Electricity (F)	J	1.11E+05	Giannetti et al., 2015	68	8.82E+07	6.50E+09	1.72E+08	1.08E+10	1.24E+09	4.67E+09	2.41E+09	1.62E+08	5.28E+10
9. Infrastructure (F)													
9.1 Copper	g	7.43E+10	Cohen et al., 2007	0	3.13E-01	5.41E-01	3.07E-01	0.00E+00	1.47E-01	2.72E-01	1.01E+00	7.77E-02	1.06E+00
9.2 Bricks	g	2.79E+09	Buranakarn, 1998	0	1.31E+03	1.87E+03	1.46E+03	2.18E+03	8.41E+02	3.05E+03	3.05E+03	5.37E+02	3.10E+03
10. Lime (F)	g	1.24E+09	Odum, 1996	0	3.75E+06	3.75E+06	3.75E+06	3.75E+06	3.75E+06	3.75E+06	3.75E+06	3.75E+06	3.75E+06
11. Organic fertilizer (F)	g	3.07E+09	Castellini et al., 2006	16	3.00E+06	3.00E+06	3.00E+06	3.00E+06	3.00E+06	3.00E+06	3.00E+06	3.00E+06	3.00E+06
12. Fuel (diesel) (F)	J	1.37E+05	Brown et al., 2011	0	3.63E+09	6.28E+09	6.28E+09	6.28E+09	1.72E+10	1.72E+10	1.72E+10	1.72E+10	1.72E+10
13. Labor (F)	\$	3.23E+12	Giannetti et al., 2018	15	0.00E+00	1.48E+02	1.48E+02	1.48E+02	1.62E+03	1.62E+03	1.62E+03	1.62E+03	1.62E+03
14. Services (F)	\$	3.23E+12	Giannetti et al., 2018	15	3.40E+03	1.24E+03	1.24E+03	1.24E+03	9.79E+03	9.79E+03	9.79E+03	9.79E+03	9.79E+03

<sup>a</sup> UEVs updated to the 1.20E25 sej/yr emergy baseline without accounting for labor and services.

Appendix B. Unit emergy value (UEV) estimation for lambari commercial feed. The amount of ingredients relate to 1g of commercial feed and based on Sussel et al. (2014). %R = renewability fraction in %.

Item	Unit	UEV <sup>a</sup> sej/Unit	Amount (Unit)	%R	Renewable emergy flow (sej)	Non-renewable emergy flow (sej)	Total emergy (sej)	Reference for UEV
Rice bran	g	9.70E+08	0.09	0	0.00E+00	8.73E+07	8.73E+07	Brown and McClanahan, 1996
Corn bran	g	1.45E+10	0.26	0	0.00E+00	3.77E+09	3.77E+09	Brandt-Williams, 2002
Soybean meal	g	3.35E+09	0.2	30	1.99E+08	4.71E+08	6.70E+08	Takahashi and Ortega, 2010
Cottonseed meal	g	4.01E+09	0.09	17	6.12E+07	3.00E+08	3.61E+08	Takahashi and Ortega, 2010
Wheat bran	g	1.09E+09	0.2	22	4.88E+07	1.69E+08	2.18E+08	Dong et al., 2008
Poultry viscera meal	g	4.05E+09	0.0325	16	2.11E+07	1.11E+08	1.32E+08	Castellini et al., 2006
Meat and bone meal	g	4.64E+10	0.027	0	0.00E+00	1.25E+09	1.25E+09	Brandt-Williams, 2002
Fishmeal	g	3.13E+09	0.0175	50 <sup>b</sup>	2.73E+07	2.73E+07	5.47E+07	Brown and Bardi, 2001
Blood meal	g	4.64E+10	0.01	0	0.00E+00	4.64E+08	4.64E+08	Brandt-Williams, 2002
Total			-		3.57E+08	6.65E+09	7.01E+09	

<sup>a</sup> UEVs updated to the 1.20E25 sej/yr emergy baseline without accounting for labor and services.

<sup>b</sup> The Brazilian sardinella (*Sardinella brasiliensis*) is one of the sardine species used as a protein source ingredient in animal feed composition. FAO suggests that excessive fishing pressure could exacerbate biomass declines and delay or compromise potential natural recoveries (available at: http://firms.fao.org/firms/resource/13329/en). Therefore, we assumed a 50% renewability for the fishmeal flow due to its current overexploitation.

# CAPÍTULO 2: Ecosystem functions in pond aquaculture and their roles in providing ecosystem services and disservices

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## 1 Ecosystem functions in pond aquaculture and their roles in providing

### 2 ecosystem services and disservices

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7

#### 8 Abstract

Managing aquaculture systems for maximizing not only fish productivity, but also other 9 potential benefits that the "aqua-ecosystems" can provide is a pathway towards more 10 sustainable systems. This study aims to provide a donor-side approach for accounting 11 the ecosystem functions and its roles on providing services and dis-services in 12 freshwater aqua-ecosystems. We applied emergy synthesis on nine semi-intensive 13 lambari aquaculture farms as a model to identify the connections between aquaculture 14 practices and the ways they may increase or diminish aqua-ecosystem services. We 15 accounted for seven water ecosystem functions that are linked to four ecosystem 16 17 services and three disservices provided by lambari culture in freshwater ponds. Water 18 regulation and microclimate regulation are services inherent to the systems features. Water provision, fish provision, and global climate regulation are influenced by the 19 management practices adopted and can have a positive or a negative impact. In this 20 study, the eutrophic effluent causes the larger disservice by demanding energy for water 21 dilution (range from 4.30E+12 to 1.69E+13 sej/ha.year<sup>-1</sup>) or treatment (range from 22 1.45E+15 to 5.65E+15 sej/ha.year<sup>-1</sup>). Greenhouse absorption and emissions vary 23 according to the management practices adopted, and in this study has a neutral effect 24 25 on global climate change. The trade-offs between positive and negative externalities in lambari aquaculture indicates that the services surpasses the disservices as long as the 26 system operates under nature's carrying capacity. 27

28

Keywords: Ecosystem approach for aquaculture; freshwater resources; ecosystem
 services; emergy; sustainability.

31

#### 32 **1. Introduction**

Aquaculture systems are widely recognized as an alternative for the future of animal 33 protein production, with prospects to deliver high quality food for the increasing world 34 population (FAO, 2020). Nevertheless, the fast expansion of the activity has amplified 35 the concern over its sustainability on the long term. While the negative impacts of 36 aquaculture production on the natural environment and local communities had 37 38 decreased over the past years, efforts remain crucial for facing the imminent threats of climate change and the post-pandemic crisis (Boyd et al., 2020). The harms caused by 39 40 climate change comprise the extreme weather events, such as floods and droughts that 41 will become more frequent and intense, consequently affecting the resilience of the 42 aquaculture systems (Ahmed et al., 2018).

These forthcoming challenges will affect especially the small-scale farms, since they 43 frequently lack of adequate technical and financial support (Ahmed et al., 2019). 44 Besides, widespread unsustainable practices in these systems, related to water use 45 efficiency and fish productivity, are major issues in climate change adaptation 46 (Verdegem and Bosma, 2009). Freshwater earthen ponds are the most adopted system 47 by small farms, and are likely to remain the prevailing system in developing areas 48 (Edwards, 2015). Ponds produce over 90% of the total production in freshwater 49 aquaculture, and contribute to sustain the livelihoods and food security of rural 50 communities (Boyd and Davis, 2020). Therefore, strategies for maximizing human 51 opportunities wile safeguarding the natural environment are urgent (United Nations, 52 2020). 53

Aquaculture ponds are engineered aquatic ecosystems, managed for improving the provision of food and income (Willot et al., 2019). As an interface between natural and human-made capital, it relies on both natural and economy domains for the delivery of inputs, management of production processes, and wastes recycling. This definition was previously applied in agriculture ecosystems, named as "agro-ecosystems", but it also suits for aquaculture, named as "aqua-ecosystems" (Shah et al., 2019; Willot et al., 2019). Human interventions has intended to maximize the provisioning service of such

systems, by constantly improving intensification technologies, which resulted on high productivity and efficiency (Boyd and Davis, 2020). On the other hand, these interventions have led systems to an overload on natures capacity to absorb pollution and restore natural stocks that sustain the activity itself (Ahmed et al., 2019; Bosma and Verdegem, 2011; Boyd et al., 2020).

66 Managing aquaculture systems for maximizing not only fish productivity, but also other potential benefits that the "aqua-ecosystems" can provide, is a pathway towards more 67 sustainable systems (Willot et al., 2019). Ecosystem services (ES) concept defines the 68 vast contributions that human beings obtain from the natural ecosystems for sustaining 69 human life and wellbeing (Costanza et al., 2017). Currently, it was identified more than 70 40 potential ES provided by aqua-ecosystems, including provisioning, regulating and 71 cultural services (Alleway et al., 2019; Beveridge et al., 1997; Custódio et al., 2020; 72 Weitzman, 2019; Willot et al., 2019). While a proper management can improve ES 73 delivery, unsustainable practices also generates dis-services, defined as the ecosystems 74 functions and aspects that results in perceived or actual damage for human wellbeing 75 (Costanza et al., 2017; Shah et al., 2019). 76

The trade-offs between positive and negative externalities are hard to evaluate, and 77 78 therefore, they are rarely presented on ES valuation studies (Yang et al., 2018). Difficulties increase as one particular practice can deliver services and dis-services 79 concurrently. For example, a shellfish farm operation at a certain area can improve 80 water quality by filtering organic particles suspended in the water column, while at the 81 same time causing damage to the benthonic environment by shellfish feces (Suplicy, 82 2020). Frameworks for evaluating ES in aquaculture systems has been a hot topic on 83 the recent scientific literature (Alleway et al., 2019; Custódio et al., 2020; Needles et al., 84 2015; Weitzman, 2019; Willot et al., 2019). Nevertheless, the receiving-side approaches 85 considered by them can be highly variable and subjective, as they rely on humans` 86 perceptions and willingness to conserve a service or product, and do not account on the 87 ecosystems structural and functional components (Giannetti et al., 2011). These 88 procedures leads to over or underestimations of the material world, failing to 89 comprehend the real wealth produced by nature (Odum and Odum, 2000; Yang et al., 90 2018). 91

92 Ecosystem services (ES) derive from ecosystems functions (EF), which are the ecological processes that control the fluxes of energy, nutrients and organic matter, through 93 ecological systems (Lee and Brown, 2021). Besides, the ability of an ecosystem to 94 provide services depends on the ability of the ecosystem to capture the natural 95 resources available (Lu et al., 2017). Therefore, a donor-side evaluation is the most 96 adequate approach to separate EF from ES, and to determine the efficiency of 97 ecosystems to produce services (Lee and Brown, 2021; Lu et al., 2017). Emergy synthesis 98 is a biophysical approach, based on a donor-side perspective valuation that recognize all 99 the work done by nature in making available a resource (Odum, 1996). The 100 understanding of the emergy driving flows and its configuration allows the design of 101 higher effective policies for the global sustainability objectives. 102

Assessing ecosystem functions flows and its services trade-offs is fundamental 103 for managing the aqua-ecosystems so they can sustainably provide the services humans 104 need with the minimum damage possible. Since water is a core element for freshwater 105 aquaculture, understanding the roles of aquatic EF in pond aquaculture is a strategic 106 107 step towards sustainability (Boltz et al., 2019; Goddard and Delghandi, 2020). Moreover, 108 it can increase aquaculture resilience for facing climate change disturbances that are water related (Ahmed et al., 2018). Therefore, this study aims to provide an emergy-109 based method for accounting the ecosystem functions and its roles on providing services 110 and dis-services in aqua-ecosystems. We identified the main connections between 111 aquaculture practices and the ways they may increase or diminish aqua-ecosystem 112 services. Finally, we applied emergy synthesis on semi-intensive lambari aquaculture 113 farms as a case study. 114

115

#### 116 **2. Materials and methods**

117

118 2.1 Ecosystem services survey and emergy synthesis of ecosystem functions

119 Several studies have highlighted the need of a non-monetary approach for the 120 assessment of nature's work as a complementarian valuation (Costanza et al., 2017). 121 Emergy synthesis has shown to be the most appropriate method for assessing

ecosystem functions, which are the ecological processes that sustain the delivery of ecosystem goods and services (Lee & Brown, 2021). The emergy synthesis is an ecocentric approach that accounts for all the work done by nature, society and economy on a common basis: solar-equivalent joules, named solar emjoules (sej). In other words, emergy with m, is all the available energy directly and indirectly embodied for the production of a good or service (Odum, 1996).

The ecosystem services provided by freshwater pond aquaculture were identified 128 based on the available literature (Custódio et al., 2020; Fu et al., 2018; Willot et al., 129 130 2019). They were summarized as water regulation, water provision, global climate regulation, microclimate regulation and fish provision. A system diagram was designed 131 for the lambari aquaculture system, aiming to identify the ecosystem functions related 132 to the services provided. Then, the emergy flows driving each function was highlighted 133 and further accounted. The calculation procedure was modified from Shah et al. (2019) 134 and Yang et al. (2019), and the equations are described below. 135

136

## 137 2.2. Ecosystems functions accounting procedure

138

*2.2.1. Groundwater recharge*: The emergy driving groundwater recharge was calculatedby:

$$Em_{gr} = K_i * p * G * UEV_{wi}$$

142 where:

143 Em<sub>gr</sub> = Emergy used to replenish the groundwater stock (sej/ha.yr<sup>-1</sup>);

144  $K_i$  = Seepage rate of soil *i* of the aquaculture pond (m<sup>3</sup>/ha.yr<sup>-1</sup>). For the studied system,

the soil is predominantly clay type with a seepage range from 1.25 to 10 mm/day (Dias

de Oliveira and Moraes, 2017; FAO, 2021);

147  $p = Water density (g/m^3);$ 

148 G = Gibbs free energy (4.49 J/g);

149 UEV<sub>wi</sub> = Unity Emergy Value of water infiltration (2.04E+04 sej/J, Brown and Bardi, 2001).

150

2.2.2 *Greenhouse gases absorption/emission*: The emergy driving greenhouse gases
absorption and emission was calculated by:

153

154

$$Em_{gg} = CO_{2-eq}(a-e) * G * UEV_{phyto}$$

155 Where:

Emgg= is the emergy used for greenhouse gases absorption/emission (sej/ha.yr<sup>-1</sup>). When emissions surpass absorption, the system provides a disservice. When absorptions surpass emissions, the system provides a service;

 $CO_{2-eq} =$  is the total greenhouse gases accounted in  $CO_2$  equivalents. For the present study, the amount of greenhouse gases emitted and absorbed by the ponds was calculated according to Valenti et al., (2018);

162 a = is the total mass of  $CO_2$  equivalents absorbed by the system (g/ha.yr<sup>-1</sup>);

163 e = is the total mass of  $CO_2$  equivalents emitted (g/ha.yr<sup>-1</sup>);

164 G = Gibbs free energy (8.96E+03 J/g);

UEV*phyto* = is the Unity Emergy Value of phytoplankton productivity (1.94E+04 sej/J,
 from Giannetti et al., 2019).

167

168 *2.2.3 Water purification*: The emergy driving the water purification function in 169 aquaculture systems was calculated under two perspectives:

170

A) Water dilution method: is the emergy driving the effluent dilution, which accounts for the nature's investment on recovering the resource quality by diluting the nutrients in a higher volume of water so the ecosystem can assimilate it and avoid eutrophication. The accounting procedure was:

175

176 
$$Em_{wd} = \left(\frac{N_{out}}{N_{st}} * EV\right) * p * G * UEVsw$$

177 Where:

- Em<sub>wd</sub> = is the emergy driving the dilution process (sej/ha.yr<sup>-1</sup>);
- N<sub>out</sub> = is the nutrient (phosphorous or nitrogen) concentration in the effluent water
   (mg/l);

N<sub>st</sub> = is the standard or targeted concentration of the same nutrient on a high quality water, <u>or</u> the nutrient concentration in the inlet water used by the farm (mg/l). In the present study, we adopted the standard concentration of phosphorous for class 1 freshwater, defined by the Brazilian Environmental Council, which is 0.02 mg/l (CONAMA, 2005);

186 EV = Is the volume of the effluent discharged by the aquaculture ponds  $(I/ha.yr^{-1})$ 

187 
$$p$$
 = Water density (g/l);

188 G = Gibbs free energy (4.49 J/g);

- UEV<sub>sw</sub> = is the Unity Emergy Value of superficial water (5.23E+04 sej/J, from Comar,
  2001).
- 191

B) Water treatment method: the emergy driving the construction and maintenance
of a water treatment plant, which accounts for the nature's investment in a system that
replaces the natural purification service. It was calculated by:

195

196 
$$Em_{wt} = \sum (R * UEV_R) + (N * UEV_N) + (F * UEV_F)$$

197 Where:

198 Em<sub>wt</sub> = is the emergy driving a wetland system for water treatment (sej/ha.yr<sup>-1</sup>);

R = is the total amount of renewable resources used, such as sun, wind and rain (joules,
 g or US\$ /ha.yr<sup>-1</sup>);

201 UEV<sub>R</sub> = is the Unity Emergy Value for each renewable input (sej/unity);

N = is the total amount of local non-renewable resources used, for example soil (joules,
 g or US\$ / ha.yr<sup>-1</sup>);

204	$UEV_N$ = is the Unity Emergy Value for each non-renewable input (sej/unity);
205	F = is the total amount of consumed resources from economy, such as materials, fuels,
206	electricity and money (joules, g or US\$ / ha.yr <sup>-1</sup> );
207	$UEV_F$ = is the Unity Emergy Value for each "F" resource (sej/unity).
208	
209	2.2.4 Evaporation: The emergy of water evaporation was calculated by:
210	
211	$Em_{we} = E_i * p * G * UEV_{we}$
212	Where:
213	Emwe = is the emergy driving water evaporation (sej/ha.yr <sup>-1</sup> )
214	Ei = is the annual evaporation ( $m^3$ / ha.yr <sup>-1</sup> ). In the present study, we adopted the average
215	evaporation rate for freshwater ponds (range from 1.0 to 2.3 $m^3$ /ha.yr <sup>-1</sup> for tropical
216	ponds, from Verdegem et al., 2006);
217	p = Water density (g/m <sup>3</sup> );
218	G = Gibbs free energy (4.49 J/g);
219	UEV <sub>we</sub> = is the Unity Emergy Value of water evaporation (7.23E+03 sej/J, from Giannetti
220	et al., 2019).
221	

221

226

222 2.2.5 Biomass increase: The emergy driving biomass increase in aquaculture systems 223 is the same of the emery driving the aquaculture farm. It includes all the resources that 224 are needed for increasing productivity, such as electricity, fuel, labor and others. It was 225 calculated by:

$$Em_{bm} = \sum (R * UEV_R) + (N * UEV_N) + (F * UEV_F)$$

227 Where:

228 Em<sub>bm</sub> = is the emergy driving biomass increase in an aquaculture system (sej/ha.yr<sup>-1</sup>);

R = is the total amount of renewable resources used, such as sun, wind and rain (joules, kg or US\$ /ha.yr<sup>-1</sup>); UEV<sub>R</sub> = is the Unity Emergy Value for each renewable input (sej/unity)

N = is the total amount of local non-renewable resources used, for example groundwater

233 (joules, kg or US\$ / ha.yr<sup>-1</sup>);

- $UEV_N = is$  the Unity Emergy Value for each non-renewable input (sej/unity)
- F = is the total amount of consumed resources from economy, such as materials, fuels,
- electricity and money (joules, kg or US\$ / ha.yr<sup>-1</sup>);
- UEV<sub>F</sub> = is the Unity Emergy Value for each "F" resource (sej/unity).
- Details on the emergy accounting of lambari systems can be found in chapter 1.
- 239

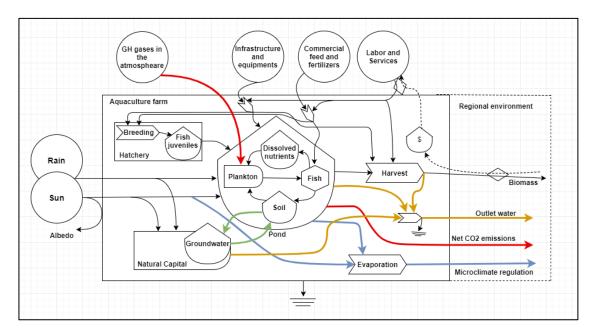
240 2.3 Case study: data source and results analysis

We applied the previous described procedure on nine semi-intensive lambari 241 aquaculture farms. The farms produce the yellowtail lambari (Astyanax lacustris) in 242 243 semi-intensive earthen ponds, intensively fed with commercial feed. Lambari farms differ for land and pond sizes, management strategies, and capital for economic 244 investments in infrastructure and equipment. These dissimilarities resulted in the three 245 categories, or levels of control, according to systems technification degree: low control, 246 moderate control, and high control. The details on the practices of the control levels can 247 be found at chapter 1. Data on natural and economic inputs, management practices and 248 landscape features were obtained in situ. Additional information were obtained through 249 250 a semi-structured survey applied to the nine farmers in the beginning of the production 251 cycle. The questionnaire focused on accounting for the total amount of materials, equipment, and infrastructure purchased, as well as labor, taxes, and depreciation. All 252 253 inflows of materials, energy and money were accounted in unities/hectare, and they correspond to one year (i.e. 3 production cycles) of farm operation. The equations for 254 the ecosystem functions evaluation were applied to each ecosystem function, in each 255 256 lambari farm, separately.

257

#### 258 **3 Results**

The ecosystem services and disservices provided by the freshwater pond aquaculture 259 rely on the ecological functions that are linked with the production system (Figure 1). 260 The functions of groundwater recharge and evaporation are inherent to the system, and 261 occur independently of human interventions. The greenhouse gases balance depends 262 mainly on the primary productivity potential and bottom soil quality, which are 263 influenced by the management practices of the system. The water purification process 264 is linked to the water source input, the amount of allochthone nutrients inputs, and the 265 management practices, as they determine the quality of the outlet water. The 266 interaction of these flows may cause an energy depreciation, by the natural effort on 267 deviating energy for reestablishing the water quality. All ecosystem functions identified 268 have a regional influence, except for greenhouse gases that have a global influence. 269



#### 270

Figure 1 - Energy system diagram of a lambari freshwater pond aquaculture system. Circles represent emergy natural and economy sources, arrows represent flows of emergy that interact with the internal stocks for producing the goods and services. Red arrows are the emergy flows driving greenhouse gases absorption and emissions. Green arrows are the emergy flows driving groundwater recharge. Blue arrows are the emegy flows driving evaporation. Yellow arrows are the emergy flows driving water purification. Black arrows are the emergy flows driving biomass increase. For details on the symbols of energy systems diagrams see Odum (1996).

278

The main services provided by the lambari aquaculture ponds are water regulation, microclimate regulation and fish provision (Table 1). These services depend on the functions of groundwater recharge, evaporation and biomass increase. Biomass 282 increase is the main goal of this aqua-ecosystem, and therefore, demands a high emergy 283 amount that varies according to the systems efficiency. Water purification, which in this case is water pollution, is the largest ecosystem disservice of lambari pond systems, in 284 both ecosystem functions assessed, but the emergy driving water treatment is larger 285 than the emergy driving effluent dilution. Global climate regulation service and 286 disservice are highly variable between the farms assessed, and can be considered as 287 neutral since the emergy driving emissions are similar to the emergy driving absorption. 288 The trade-offs concerning the flows of emergy driving services and disservices have a 289 positive result even when the biomass increase is not accounted. 290

291

Table 1 - Emergy flows driving the water ecosystem functions that support the delivery of ecosystem services and disservices in lambari freshwater pond aquaculture systems. Flows are accounted in sej (solar emjoules) per hectare per year. Positive values are services and negative values are disservices. Net emergy = emergy flow of services – emergy flow of disservices.

Factoria comica	Fooductor function	Turne	Emergy flow (sej/ha.yr <sup>-1</sup> )			
Ecosystem service	Ecosystem function	Туре –	Min	Max		
Water regulation	Groundwater recharge	Service	4.18E+14	3.34E+15		
Water purification	Water dilution	Disservice	- 4.30E+12	- 1.69E+13		
Water purification	Water treatment	Disservice	- 1.45E+15	- 5.65E+15		
Global climate regulation	Greenhouse gas uptake	Service	7.39E+15	1.43E+17		
Global climate regulation	Greenhouse gas release	Disservice	- 5.36E+15	- 1.29E+17		
Microclimate regulation	Evaporation	Service	3.25E+10	7.47E+10		
Fish provision	Biomass accumulation	Service	6.23E+16*	1.07E+18*		
Total emergy flow dri accumulation	7.81E+15	1.46E+17				
Total emergy flow dri dilution	ving disservices, conside	5.37E+15	1.29E+17			
Total emergy flow dri treatment	ving disservices, conside	6.81E+15	1.34E+17			

296

\* Biomass accumulation in sej/ha.yr<sup>-1</sup>, excluding services (i. e. the monetary flow).

297

## 298 **4 Discussion**

Aquaculture ponds are aqua-ecosystems that provides services and disservices. In the present study, we identified and assessed seven water ecosystem functions that are 301 linked to four ecosystem services, and three disservices provided by lambari culture in freshwater ponds. Water regulation and microclimate regulation services are inherent 302 to the systems features, while water purification, fish provision, and global climate 303 regulation, are influenced by the management practices adopted, and can have positive 304 or negative impact. In this study, the eutrophic effluent causes the larger disservice by 305 demanding energy for water dilution or treatment. Greenhouse absorption and 306 emissions vary according to the management practices, and in this study has a neutral 307 effect. Overall, the services provided by the aqua-ecosystem surpass the disservices 308 309 under an eco-centric point of view, as long as the system operates under the carrying capacity of the watershed. 310

Water seepage is an inherent feature of aquaculture earthen ponds, and often 311 represent a technical and environmental problem as it increases the water inflow 312 needed for maintaining the pond level (Bosma and Verdegem, 2011; Boyd and Davis, 313 2020). Nevertheless, the water that infiltrates through the lateral and bottom soil of 314 ponds, reaches the water table. The soil and underlying geological formations act as a 315 316 filter for organic pollutants and harmful bacteria that may be present, restoring water 317 quality (Boyd et al., 2002). The ponds temporally retains water from rain and streams, but a fraction of that water potentially recharge the groundwater stocks. By doing so, 318 aquifers may stock high quality water that are essential for maintaining water cycle and 319 securing water availability. This result indicates that pond aquaculture could be an 320 economic activity introduced in degraded areas, such as eroded lands, as a strategy for 321 recovering the natural capital of the area while benefiting people. 322

Water evaporation occurs directly from pond surfaces, and is controlled by water and 323 324 air temperature, humidity, wind velocity, and atmospheric pressure (Boyd and Davis, 325 2020). Water evaporation from aquatic environments plays an important role on the air quality and thermal comfort, especially in areas suffering with intense and long drought 326 season (Yang et al., 2019). Besides, the service of microclimate regulation is becoming 327 more relevant as climate change intensifies the drought periods. The lambari 328 aquaculture farms studied are located in the southwest region of São Paulo state in 329 Brazil. That landscape area is characterized by the predominance of sugarcane 330 monocultures, which has a high negative impact on the regional air quality due to 331

emissions caused by inadequate harvest practices (Bordonal et al., 2018). Therefore, the Integration of sugarcane and livestock sectors such as aquaculture is recommended to improve land use efficiency in Brazil, since it would increase productivity while mitigating the negative impacts of highly intensive monocultures (Bordonal et al., 2018).

Water purification in aquatic ecosystems is their capacity to remove contaminants by 336 dilution, sedimentation, aeration, absorption, flotation, chemical and biological 337 reactions (Yang et al., 2019). The biophysical mechanisms of aquaculture ponds are 338 similar to any lotic environment (Boyd and Davis, 2020). Nevertheless, human 339 340 interventions on aquaculture ponds aims to overpass its natural productivity by adding alloctonous sources of energy and nutrients. Normally, only 20-40% of the nitrogen and 341 phosphorous added by fertilization and feed to a pond is recovered on fish biomass 342 (Bosma and Verdegem, 2011). From what remains, it becomes effluents to be 343 discharged mainly in the local streams (Boyd and Davis, 2020; Verdegem and Bosma, 344 2009). Because of the eutrophic nature of aquaculture effluents, as in the case of 345 lambari farms, it can have a negative additive effect on the receiving waterbody, spread 346 disease, and decrease the quality of water downstream (Verdegem and Bosma, 2009). 347

348 When the farms externalize these negative impacts, natures deviates energy from other 349 ecological processes for reestablishing the resource quality. The environmental cost of recovering water quality is lower when the dilution of the effluent by the receiving water 350 body is possible. Nevertheless, the carrying capacity of the watershed has to be 351 considered. In the present study, each m<sup>3</sup> of effluent costs in average 17 m<sup>3</sup> of high 352 quality water invested by nature. On the other hand, the water treatment by a simple 353 354 wetland plant consumes 200 times more emergy per hectare of farm. As an alternative, 355 aquaculture systems could make use of hypereutrophic water sources for maximizing 356 the use of nutrients, which results on a higher quality outlet water compared to its source (David et al., 2017; Flickinger et al., 2019). Moreover, the Integrated multi-357 trophic aquaculture (IMTA), as well as non-fed species culture has ecosystem functions 358 that purify the water used(Alleway et al., 2019; Custódio et al., 2020; Franchini et al., 359 2020). These innovative approaches may be an alternative for internalizing this 360 disservice caused by lambari aquaculture effluent. 361

362 Aquaculture emits and absorbs carbon concomitantly (Ahmed and Thompson, 2019; 363 Boyd and Davis, 2020). Carbon emission occurs via bubbles released from the sediment during the process of organic matter decomposition (Valenti et al., 2018). Additionally, 364 exchanges of gas between the water surface and the atmosphere releases the CO2 365 produced by then respiration process of water animals, plants and microbiota (Valenti 366 et al., 2018). On the other hand, the photosynthesis mechanism relies on carbon 367 sequestration during the daytime (Bosma and Verdegem, 2011). Therefore, the balance 368 of emitted versus absorbed carbon of an aquaculture system demonstrates the 369 370 contribution of the systems as a service or di-service provider. The emergy valuation suggested that lambari aquaculture has a neutral effect on global climate regulation; 371 nevertheless, the results are highly variable among farms, indicating that location, pond 372 age, feed management and species cultured may affect the carbon balance (Flickinger 373 374 et al., 2020).

The trade-offs between positive and negative externalities in lambari aquaculture 375 indicates that the services surpasses the disservices as long as the system operates 376 377 under nature's carrying capacity. Moreover, systems must be designed considering the 378 internalization of the externalities, aiming to maximize the positive and minimize the negative. Nevertheless, the internalization of the environmental costs by remediating 379 the damage does not necessarily leads to more sustainable systems. By doing so, more 380 energy and resources are needed, which diminishes the natural capital in detriment of 381 recovering the damage caused by the system itself. An ecological approach that 382 considers the integration of species and cultures, the upcycling of resources, and the 383 surrounding biophysical aspects, makes systems higher efficient at a lower 384 385 environmental cost, which seems much more effective on the long term.

The ecosystem features and their ability for converting the available resources into services are determinant for the quantity and quality of services delivered. Lee & Brown (2021) stated that the value of an ecosystem function is the total emergy driving the system, and since the functions are co-products of an ecosystem driven by the same sources, they should not be added. Moreover, a static evaluation does not account for the feedback and cycling of emergy over time scales greater than one year (Lee & Brown 2021). Nevertheless, when assessing EFs in a production system that is artificially

organized for maximizing productivity regardless of other possible environmental services and disservices, the valuation of the emergy flows driving the specific EFs can be strategic for sustainability. An environmental management of human-made ecosystems that maximizes the emergy production and use is more likely to enhance the delivery of goods and services (Odum e Odum, 2000).

398

## 399 **5. Conclusion**

Besides fish provision, aquaculture freshwater pond systems provides ecosystem 400 services and disservices that rely on the ecological functions of the system. The services 401 402 of water regulation and microclimate regulation are positive inherent features of the 403 system, and rely on groundwater recharge and evaporation functions. Greenhouse absorption and emissions vary according to the management practices applied, and in 404 this study has shown a neutral effect on global climate regulation. The eutrophic quality 405 of the farms effluent causes the higher disservice, by wasting resources for recovering 406 the water quality. The trade-offs between positive and negative externalities in lambari 407 aquaculture indicates that the services surpasses the disservices as long as the system 408 409 operates under nature's carrying capacity. An ecological approach that considers the 410 integration of species and cultures, the upcycling of resources, is more effective in the internalization of the positive and negative externalities, improving systems 411 sustainability and resilience by restoring the natural capital. 412

413

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415

#### 416 **References**

- 417 Ahmed, N., Thompson, S., 2019. The blue dimensions of aquaculture: A global
- 418 synthesis. Sci. Total Environ. 652, 851–861.
- 419 https://doi.org/10.1016/j.scitotenv.2018.10.163
- 420 Ahmed, N., Thompson, S., Glaser, M., 2019. Global Aquaculture Productivity,
- 421 Environmental Sustainability, and Climate Change Adaptability. Environ. Manage. 63,
- 422 159–172. https://doi.org/10.1007/s00267-018-1117-3

- 423 Ahmed, N., Ward, J.D., Thompson, S., Saint, C.P., Diana, J.S., 2018. Blue-Green Water
- 424 Nexus in Aquaculture for Resilience to Climate Change. Rev. Fish. Sci. Aquac. 26, 139–
- 425 154. https://doi.org/10.1080/23308249.2017.1373743
- 426 Alleway, H.K., Gillies, C.L., Bishop, M.J., Gentry, R.R., Theuerkauf, S.J., Jones, R., 2019.
- 427 The Ecosystem Services of Marine Aquaculture: Valuing Benefits to People and Nature.
- 428 Bioscience 69, 59–68. https://doi.org/10.1093/biosci/biy137
- 429 Beveridge, M.C.M., Phillips, M.J., Macintosh, D.J., 1997. Aquaculture and the
- 430 environment: The supply of and demand for environmental goods and services by
- 431 Asian aquaculture and the implications for sustainability. Aquac. Res. 28, 797–807.
- 432 https://doi.org/10.1111/j.1365-2109.1997.tb01004.x
- 433 Boltz, F., LeRoy Poff, N., Folke, C., Kete, N., Brown, C.M., St. George Freeman, S.,
- 434 Matthews, J.H., Martinez, A., Rockström, J., 2019. Water is a master variable: Solving
- for resilience in the modern era. Water Secur. 8, 100048.
- 436 https://doi.org/10.1016/j.wasec.2019.100048
- 437 Bordonal, R. de O., Carvalho, J.L.N., Lal, R., de Figueiredo, E.B., de Oliveira, B.G., La
- 438 Scala, N., 2018. Sustainability of sugarcane production in Brazil. A review. Agron.
- 439 Sustain. Dev. 38, 13. https://doi.org/10.1007/s13593-018-0490-x
- 440 Bosma, R.H., Verdegem, M.C.J., 2011. Sustainable aquaculture in ponds: Principles,
- 441 practices and limits. Livest. Sci. 139, 58–68.
- 442 https://doi.org/10.1016/j.livsci.2011.03.017
- Boyd, C.E., D'Abramo, L.R., Glencross, B.D., Huyben, D.C., Juarez, L.M., Lockwood, G.S.,
- 444 McNevin, A.A., Tacon, A.G.J., Teletchea, F., Tomasso, J.R., Tucker, C.S., Valenti, W.C.,
- 445 2020. Achieving sustainable aquaculture: Historical and current perspectives and
- 446 future needs and challenges. J. World Aquac. Soc. 51, 578–633.
- 447 https://doi.org/10.1111/jwas.12714
- Boyd, C.E., Davis, R.P., 2020. Lentic Freshwater: Ponds—Aquaculture Ponds, in:
- 449 Encyclopedia of the World's Biomes. Elsevier, pp. 316–324.
- 450 https://doi.org/10.1016/B978-0-12-409548-9.11956-6
- 451 Boyd, C.E., Wood, C., Thunjai, T., 2002. Aquaculture Pond Bottom Soil Quality

- 452 Management 1–48.
- 453 Comar, V., 2001. Emergy Evaluation of Organic and Conventional Horticultural
- 454 Production in Botucatu, Sao Paulo State, Brazil, in: Brown, M.T. (Ed.), Emergy Synthesis
- 1: Theory and Applications of the Emergy Methodology. Proceedings of the 1st
- 456 Biennial Emergy Conference. Center for Environmental Policy, University of Florida,
- 457 Gainesville, pp. 181–195.
- 458 CONAMA, 2005. RESOLUÇÃO CONAMA Nº 357, DE 17 DE MARÇO DE 2005. Brazil.
- 459 Costanza, R., de Groot, R., Braat, L., Kubiszewski, I., Fioramonti, L., Sutton, P., Farber,
- S., Grasso, M., 2017. Twenty years of ecosystem services: How far have we come and
- how far do we still need to go? Ecosyst. Serv. 28, 1–16.
- 462 https://doi.org/10.1016/j.ecoser.2017.09.008
- 463 Custódio, M., Villasante, S., Calado, R., Lillebø, A.I., 2020. Valuation of Ecosystem
- 464 Services to promote sustainable aquaculture practices. Rev. Aquac. 12, 392–405.
- 465 https://doi.org/10.1111/raq.12324
- 466 David, F.S., Proença, D.C., Valenti, W.C., 2017. Phosphorus Budget in Integrated
- 467 Multitrophic Aquaculture Systems with Nile Tilapia, Oreochromis niloticus, and
- 468 Amazon River Prawn, Macrobrachium amazonicum. J. World Aquac. Soc.
- 469 https://doi.org/10.1111/jwas.12404
- 470 Dias de Oliveira, M.E., Moraes, S.O., 2017. Modeling approaches for agricultural N2O
- 471 fluxes from large scale areas: A case for sugarcane crops in the state of São Paulo -
- 472 Brazil. Agric. Syst. 150, 1–11. https://doi.org/10.1016/j.agsy.2016.09.015
- 473 Edwards, P., 2015. Aquaculture environment interactions: Past, present and likely
- 474 future trends. Aquaculture 447, 2–14.
- 475 https://doi.org/10.1016/j.aquaculture.2015.02.001
- 476 FAO, 2020. The State of World Fisheries and Aquaculture 2020. Sustainability in action.
- 477 FAO, Rome. https://doi.org/10.4060/ca9229en
- 478 Flickinger, D.L., Costa, G.A., Dantas, D.P., Proença, D.C., David, F.S., Durborow, R.M.,
- 479 Moraes-Valenti, P., Valenti, W.C., 2020. The budget of carbon in the farming of the

- 480 Amazon river prawn and tambaqui fish in earthen pond monoculture and integrated
- 481 multitrophic systems. Aquac. Reports. https://doi.org/10.1016/j.aqrep.2020.100340
- 482 Flickinger, D.L., Costa, G.A., P. Dantas, D., Moraes-Valenti, P., Valenti, W.C., 2019. The
- 483 budget of nitrogen in the grow-out of the Amazon river prawn (Macrobrachium
- 484 amazonicum Heller) and tambaqui (Colossoma macropomum Cuvier) farmed in
- 485 monoculture and in integrated multitrophic aquaculture systems. Aquac. Res.
- 486 https://doi.org/10.1111/are.14304
- 487 Franchini, A.C., Costa, G.A., Pereira, S.A., Valenti, W.C., Moraes-Valenti, P., 2020.
- 488 Improving production and diet assimilation in fish-prawn integrated aquaculture, using
- iliophagus species. Aquaculture 521, 735048.
- 490 https://doi.org/10.1016/j.aquaculture.2020.735048
- 491 Fu, B., Xu, P., Wang, Y., Yan, K., Chaudhary, S., 2018. Assessment of the ecosystem
- services provided by ponds in hilly areas. Sci. Total Environ. 642, 979–987.
- 493 https://doi.org/10.1016/j.scitotenv.2018.06.138
- 494 Giannetti, B.F., Marcilio, M.D.F.D.F.B., Coscieme, L., Agostinho, F., Liu, G., Almeida,
- 495 C.M.V.B., 2019. Howard Odum's "Self-organization, transformity and information":
- Three decades of empirical evidence. Ecol. Modell. 407, 108717.
- 497 https://doi.org/10.1016/j.ecolmodel.2019.06.005
- 498 Goddard, S., Delghandi, M., 2020. Importance of the Conservation and Management of
- 499 Freshwater to Aquaculture, in: Encyclopedia of the World's Biomes. Elsevier, pp. 35–
- 500 44. https://doi.org/10.1016/B978-0-12-409548-9.11954-2
- Lee, D.J., Brown, M.T., 2021. Estimating the Value of Global Ecosystem Structure and
- 502 Productivity: A Geographic Information System and Emergy Based Approach. Ecol.
- 503 Modell. 439, 109307. https://doi.org/10.1016/j.ecolmodel.2020.109307
- Lu, H., Campbell, E.T., Campbell, D.E., Wang, C., Ren, H., 2017. Dynamics of ecosystem
- services provided by subtropical forests in Southeast China during succession as
- measured by donor and receiver value. Ecosyst. Serv. 23, 248–258.
- 507 https://doi.org/10.1016/j.ecoser.2016.11.012
- 508 Needles, L.A., Lester, S.E., Ambrose, R., Andren, A., Beyeler, M., Connor, M.S., Eckman,

- J.E., Costa-Pierce, B.A., Gaines, S.D., Lafferty, K.D., Lenihan, H.S., Parrish, J., Peterson,
- 510 M.S., Scaroni, A.E., Weis, J.S., Wendt, D.E., 2015. Managing Bay and Estuarine
- 511 Ecosystems for Multiple Services. Estuaries and Coasts 38, 35–48.
- 512 https://doi.org/10.1007/s12237-013-9602-7
- 513 Odum, H.T. (Ed.), 1996. Environmental Accounting: Emergy and Environmental
- 514 Decision Making, 1st ed. John Wiley & Sons, New York.
- 515 Odum, H.T., Odum, E.P., 2000. The Energetic Basis for Valuation of Ecosystem Services.
- 516 Ecosystems 3, 21–23. https://doi.org/10.1007/s100210000005
- 517 Shah, S.M., Liu, G., Yang, Q., Wang, X., Casazza, M., Agostinho, F., Lombardi, G.V.,
- Giannetti, B.F., 2019. Emergy-based valuation of agriculture ecosystem services and
- 519 dis-services. J. Clean. Prod. 239, 118019.
- 520 https://doi.org/10.1016/j.jclepro.2019.118019
- 521 Suplicy, F.M., 2020. A review of the multiple benefits of mussel farming. Rev. Aquac.
- 522 12, 204–223. https://doi.org/10.1111/raq.12313
- 523 United Nations, 2020. UN Research Roadmap for the COVID-19 Recovery. New York.
- 524 Valenti, W.C., Kimpara, J.M., Preto, B. de L., Moraes-Valenti, P., 2018. Indicators of
- sustainability to assess aquaculture systems. Ecol. Indic. 88, 402–413.
- 526 https://doi.org/10.1016/j.ecolind.2017.12.068
- 527 Verdegem, M.C.J., Bosma, R.H., 2009. Water withdrawal for brackish and inland
- 528 aquaculture, and options to produce more fish in ponds with present water use. Water
- <sup>529</sup> Policy 11, 52–68. https://doi.org/10.2166/wp.2009.003
- 530 Verdegem, M.C.J.J., Bosma, R.H., Verreth, J.A.J.J., 2006. Reducing water use for animal
- production through aquaculture. Int. J. Water Resour. Dev. 22, 101–113.
- 532 https://doi.org/10.1080/07900620500405544
- 533 Weitzman, J., 2019. Applying the ecosystem services concept to aquaculture: A review
- of approaches, definitions, and uses. Ecosyst. Serv. 35, 194–206.
- 535 https://doi.org/10.1016/j.ecoser.2018.12.009
- 536 Willot, P.-A.A., Aubin, J., Salles, J.-M.M., Wilfart, A., 2019. Ecosystem service

- framework and typology for an ecosystem approach to aquaculture. Aquaculture 512,
- 538 734260. https://doi.org/10.1016/j.aquaculture.2019.734260
- Yang, Q., Liu, G., Casazza, M., Campbell, E.T., Giannetti, B.F., Brown, M.T., 2018.
- 540 Development of a new framework for non-monetary accounting on ecosystem services
- valuation. Ecosyst. Serv. 34, 37–54. https://doi.org/10.1016/j.ecoser.2018.09.006
- 542 Yang, Q., Liu, G., Casazza, M., Hao, Y., Giannetti, B.F., 2019. Emergy-based accounting
- 543 method for aquatic ecosystem services valuation: A case of China. J. Clean. Prod. 230,
- 544 55–68. https://doi.org/10.1016/j.jclepro.2019.05.080

## CAPÍTULO 3: Multi-Criteria Sustainability Assessment of Lambari Aquaculture in Brazil

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## 546 Multi-Criteria Sustainability Assessment of Lambari Aquaculture in Brazil

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## 556 Abstract

Lambari aquaculture has the opportunity to promote sustainable development for 557 rural populations. For that, efforts on the assessment, monitoring and planning of 558 sustainability are a mandatory pathway. This study applies the five sector multi-559 criteria sustainability assessment (5SEnSU) on lambari aquaculture farms, as a 560 model for investigating the sustainability of small-scale freshwater aquaculture in 561 Brazil. We studied nine aquaculture farms, located in southeast region. Indicators 562 of environmental, social and economic sustainability were selected according to 563 the 5SEnSU model and analyzed by goal programming. The aquaculture 564 systems studied have different performances of sustainability, depending on the 565 level of control of major farming variables. Low control farms performed better at 566 the social dimension, contributing mainly for the local socio-economic 567 development and for the maintenance of rural livelihoods. On the other hand, 568 moderate and high control farms performed better at the environmental and 569 economic dimensions as they use natural resources more efficiently and achieve 570 higher productivity and profitability. Results indicated that a systemic and multi-571 criteria approach can be useful for the strategic planning towards the sustainable 572 development of smallholder aquaculture systems. 573

574 **Keywords:** smallholder aquaculture; sustainable development; indicators;

575 multi-criteria assessment.

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## 576 **1. Introduction**

Aquaculture is the fastest growing food production sector in the recent 577 decades, and is expected to grow 37% by 2030 (Garlock et al., 2020). The world 578 aquaculture production reached 82.1 million tonnes of fish and 32.4 million tonnes 579 of aquatic plants in 2018, and has already surpassed the wild catch in 35 580 countries (FAO, 2020). Brazil has exceptional environmental and social 581 conditions to produce aquatic organisms, and currently occupies the 8th position 582 in the world ranking of the largest producers of fish by aquaculture. In 2019, the 583 production was ~800,000 tonnes, which represents a gross revenue of about US\$ 584 1 billion, mostly commercialized at the domestic market (IBGE, 2020; Valenti et 585 al., 2021). The freshwater fish species are the most produced, and the culture of 586 native species is widespread in Brazil, playing a very important role on 587 socioeconomic development. Most of the Brazilian aquaculture production comes 588 from small rural properties (<2 ha), where the production is carried out mainly in 589 freshwater ponds (95%). Currently, more than 200 thousand freshwater fish 590 farms are active in Brazil (Valenti et al., 2021). 591

The global major challenges for the freshwater aquaculture are currently 592 related to government regulation, the lack of a structured production chain, 593 environmental pollution, the scape of alien and interspecific hybrid species, 594 animal welfare and health, adequate nutrition and technical support (Boyd et al., 595 2020; Brugère et al., 2019; Henares et al., 2019). The scientific community has 596 developed several technologies for addressing some of these challenges. Among 597 others examples are: technology for effluents treatment, native species breeding 598 and rearing methods, suitable drugs and probiotics, genetic improvements, 599 decrease of the feed conversion ratio, replacement of fish meal in aquafeed by 600 vegetal protein, and recirculating systems (Antonucci and Costa, 2020; Dawood 601 et al., 2019; Humphries et al., 2019; Lulijwa et al., 2019; Tacon, 2020). Despite 602 the significant improvements achieved, solutions are costly and frequently 603 unattainable by small farmers. Moreover, current technology could prove 604 ineffective at mitigating the complex threats caused by SARS-CoV-2 and climate 605 change crises in the near future. Therefore, innovative technologies that can 606 adjust to all these challenges will be vital for its sustainability in the long term. 607

Sustainability is a fundamental goal for the development of human society. 608 and has been a major concern for the future of aquaculture (Boyd et al., 2020; 609 FAO, 2020; United Nations, 2015). The recent discussions on the topic have 610 focused on assessing the sustainability of different production systems, 611 management practices and levels of intensification. For this, several assessment 612 methods have been used, such as life cycle assessment (LCA), set of 613 sustainability indicators and emergy synthesis (David et al., 2020; Mungkung et 614 al., 2013; Valenti et al., 2018). Nevertheless, small rural producers, who represent 615 the majority of systems in operation in Brazil, have received little attention in 616 strategic planning for the sustainable development of the activity worldwide 617 (Brugère et al., 2019). Even though research had investigated the role of rural 618 aquaculture systems on poverty reduction (Béné et al., 2016), the analysis lack 619 620 of a systemic approach to investigate the systems' functioning and the ways they relate to the social, economic and environmental spheres in which they are 621 inserted. 622

Lambari (Astyanax lacustris) is a freshwater small native fish cultured for 623 the market of live bait for sport fishing (Fonseca et al., 2017). The production is 624 concentrated in the southern and southeastern region, mainly in São Paulo State. 625 The current production is estimated to be above 1,000 tonnes per year, which 626 most part is commercialized alive at the farm gate, although the market for human 627 consumption and a dedicated processing plant exist in Brazil (Valenti et al., 628 2021). The bulk of production comes from small ponds (0.03 to 0.03 ha), in small 629 farms, where the fish is cultured in fed monoculture systems although a 630 631 commercial feed designed for the nutritional requirements of the specie is inexistent. Moreover, the activity faces the same challenges as the majority of 632 small rural aquaculture systems in Brazil: poor management practices, lack of a 633 structured production chain, adequate nutrition, environmental pollution and 634 technical support (Valenti et al., 2021). Nevertheless, as a native low-trophic 635 specie, lambari aquaculture has the opportunity to promote sustainable 636 development for rural populations (Fonseca et al., 2017). For that, efforts on the 637 assessment, monitoring and planning of sustainability are a mandatory pathway. 638 639 Therefore, lambari aquaculture systems are a good model for the study of the

sustainability of small freshwater pond aquaculture, since their prospects andchallenges are similar to the overall small aquaculture farms worldwide.

The multicriteria assessment of sustainability based on the five sectors 642 (model 5SEnSU) can be a useful tool for measuring and establishing goals for 643 644 the sustainable development of aquaculture in Brazil. The model is based on the premise that human systems are thermodynamically open, which demands 645 energy and materials from nature that will be transformed into goods and services 646 by means of human work (Bastianoni et al., 2016; Giannetti et al., 2019b). 647 According to the conceptual model of 5SEnSU, the environment acts as a 648 supplier of natural resources (sector 1) and as a receiver of by-products and 649 wastes (sector 2) from the productive sector (sector 3). On the other hand, society 650 acts as a supplier of labor and technology (sector 4), and as a receiver of goods 651 and services (sector 5) produced by sector 3 (Giannetti et al., 2019b). This model 652 allows a holistic view of the production system, which occurs through the 653 quantification of physical exchanges that occur between the environmental, 654 economic and social sectors, which provides a scientifically robust measurement 655 of the sustainability of the production systems. Therefore, this study applies the 656 5SEnSU multi-criteria assessment on lambari aquaculture farms as a model for 657 investigating the sustainability of small-scale freshwater aquaculture in Brazil. 658

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## 2. Materials and Methods

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## 2.1 - Farms description and data source

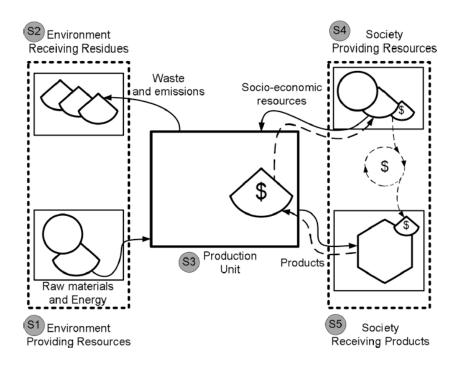
We studied nine aquaculture farms that produce the yellowtail lambari 663 (Astyanax lacustris) in intensively fed monoculture using earthen ponds. The 664 farms differ by land and pond sizes, management strategies, capital for economic 665 investments in infrastructure and equipment, technification, and control of the 666 system by the operators. These dissimilarities resulted in the three categories, or 667 levels of control: low control, moderate control, and high control. The details on 668 the practices of the control levels can be found at chapter 1. Data on natural and 669 economic inputs, management practices and landscape features were obtained 670 in situ through a semi-structured survey applied to the nine farmers. The 671 questionnaire focused on accounting for the total amount of materials, 672

equipment, and infrastructure purchased, as well as labor, taxes, and depreciation. Additional information was obtained on the available literature. All inflows of materials, energy and money were accounted in unities/hectare, and they correspond to one year (i.e. 3 production cycles) of farm operation.

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## 678 2.2 – Five sector sustainability (*5SEnSU*) conceptual model

Human-driven systems are biophysically open systems, which demand 679 energy and materials inputs that are transformed by means of human labor efforts 680 into goods, services and wastes (Bastianoni et al., 2016). Assuming that, the 681 5SEnSU model considers the social, economic and environmental dimensions of 682 sustainability, based on their functions and on the biophysical exchanges among 683 them (Figure 1). According to the model, the environment dimension act as a 684 provider of raw materials and energy (Sector 1), and as a receiver of wastes and 685 emissions (Sector 2) from the production system (Sector 3). On the other hand, 686 society acts as a provider of human capital (Sector 4), and as a receiver of the 687 products or services and externalities (Sector 5). The first step of the 5SEnSU 688 assessment is the selection of indicators. In the present study, two indicators 689 were selected for each of the five sectors of the model, allowing a balance among 690 the environmental, social, and economic dimensions of sustainability (Giannetti 691 et al., 2019b). They were selected according to their relevance, sensitivity, and 692 representativeness of the sector, and account for a time-window of one year of 693 farm operation. 694



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Figure 2- FIVE SEctor SUstainability (5SEnSU) model from Giannetti et. al. (2019). The
symbology used here comes from the emergy accounting method as available in Odum (1996),
where circles mean energy sources, "water tank" means storage, hexagon means consumers,
continuous arrows mean material and energy flows and dashed arrows mean money flows.
S=sector.

For the sector 1, Unity Emergy Value (UEV - sej/kg) and use of water (m<sup>3</sup>/kg) 702 were selected as indicators of the efficiency of the aquaculture system for 703 converting natural resources into fish (UEV), and of the systems' reliance upon 704 the water resources. For the sector 2, the flow of positive and negative ecosystem 705 functions (sei/ha/year) were selected as indicators of the roles of the aquaculture 706 system on restoring and depleting the functioning of the ecosystem in which the 707 system is inserted. Productivity (t/ha) and Internal Rate of Return (%) were the 708 indicators selected to demonstrate the economic performance of the sector 3. For 709 the sector 4, the local jobs generation (%) and the time of permanence of the 710 farmers at the business (years) are indicators of the contribution of the local 711 community to the enterprise. For the sector 5, the fraction of the lambari 712 production that is consumed locally (%), and the fraction of inputs for production 713 that is purchased locally (%) are the indicators of the benefits received by the 714 local community. 715

## 2.3 - Multi Criteria Decision Making: goal programming

The sustainability assessment of the 5SEnSU model was based on the goal 718 programming method for the multi criteria decision making (Giannetti et al., 719 2019b). This approach allows handling multiple, normally conflicting goals 720 721 represented by indicators. In summary, the logic is to measure the distance of the assessed system, represented by an indicator, from achieving the desired goal. 722 The deviation can be either positive or negative, depending on the objective of 723 maximizing or minimizing the chosen indicator. In the present study, no weigh 724 was assigned as the sectors were assumed to have equal relevance. The goals 725 established were the maximum value found among systems for the objective of 726 maximizing, and the minimum value for the objective of minimizing. The 727 equations of the GP procedure are as follows: 728

729 (1) 
$$ISG_{ijk}^{+} - \sum_{ijk} \frac{N_{ijk}^{+}}{W_{jk'}^{+}G_{jk}^{+}} + \sum_{ijk} \frac{P_{ijk}^{+}}{W_{jk'}^{-}G_{jk}^{+}}$$

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$$\forall i \in \{1, 2, \dots NE\} \forall j \in \{1, 2, \dots NS\} \forall k \in \{1, 2, \dots NI\}$$

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732 (2) 
$$ISG_{ijk}^- - \sum_{ijk} \frac{N_{ijk}^-}{W_{jk}^- G_{jk}^-} + \sum_{ijk} \frac{P_{ijk}^-}{W_{jk}^+ G_{jk}^-}$$

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$$\forall i \in \{1, 2, \dots NE\} \forall j \in \{1, 2, \dots NS\} \forall k \in \{1, 2, \dots NI\}$$

735 ISG= index of sustainability goal of indicator

Nijk+ and Nijk- = positive and negative indicators for the negative deviation
 variables, respectively;

Pijk+ and Pijk- =positive and negative indicators for the positive deviation
 variables, respectively;

Gjk+ and Gjk- = goals for the positive or negative indicators;

741 Wjk+ and Wjk- = the weight for each indicator;

NE, NS, and NI are the amount of evaluated systems, sectors, and indicators
 per sector, respectively;

i, j, and k represents the system being evaluated, the correspondent sector to 744 the 5SEnSU model, and the indicator(s) for each sector, respectively. 745 When added, the ISGs (Eqs. (1) and (2)) provide the sector sustainability 746 indicator (SSI), representing the sum of the differences between the positive and 747 748 negative deviations for each sector of 5SEnSU model: (3)  $SSI_{ii} - WS \sum_{iik} (ISG^+_{iik} - ISG^-_{iik})$ 749  $\forall i \in \{1, 2, \dots NE\} \forall j \in \{1, 2, \dots NS\} \forall k \in \{1, 2, \dots NI\}$ 750 751 in which: 752 WS = the weight established for each sector. 753 754 Finally, the sustainability synthetic indicator of system (SSIS) can be obtained 755 by adding the SSI of each sector: 756 757 (4)  $SSIS_i - \sum_{i}^5 SSI_{ii}$ 758  $\forall i \in \{1, 2, \dots NE\} \forall j \in \{1, 2, \dots NS\}$ 759 760

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# 3. Results

The indicators of environmental, social and economic sustainability for the 762 nine lambari aquaculture farms are presented at table 1. Unity emergy value 763 (UEV) and use of water were the indicators of sector 1. For UEV, the lower value 764 indicated higher efficiency; therefore, the UEV achieved by HC2 was selected as 765 a goal. For water use, HC1 had the lowest use of water, which was selected as 766 a goal. In the sector 2, the values of positive ecosystem services were slightly 767 different among the systems assessed, but the highest value were achieved by 768 HC1. For the negative ecosystem functions, LC1 had the lowest value, which was 769 selected as the goal. For the sector 3, HC1 and HC2, had the best performance 770 for productivity and I.R.R. respectively. In the sector 4, the systems had equal 771 performance for the local jobs indicator, and the HC2 had the highest value for 772 time of permanence. Finally, LC1, LC2 and LC3 had the highest value for local 773 consumption and for local purchases. 774

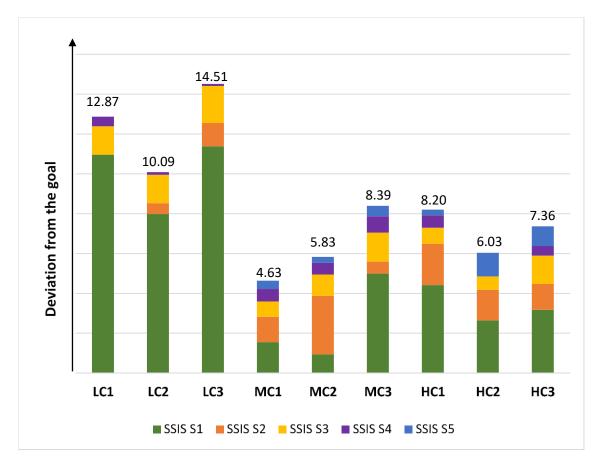
Table 2- Indicators of environmental, social and economic sustainability for the nine lambari aquaculture farms. UEV = Unity Emergy Value; LC= low control farm; MC = moderate control farm; HC = high control farm. Numbers represent different farms within the same control level.

	Sector 1 – Environment as a provider		Sector 2 – Environment as a receiver		Sector 3 – Production system		Sector 4 – Society as a provider		Sector 5 – Society as a receiver	
Systems	UEV (sej/j)	Water use (m³/kg)	Positive ecosyste m functions (sej/ha)	Negative ecosyste m functions (sej/ha)	Productivit y (t/ha)	Internal Rate of Return* (%)	Local jobs generatio n ratio (%)	Farmer permanenc e on the activity (years)	Fraction of production consumed locally (%)	Fraction of inputs purchased locally (%)
LC1	2.84E+06	36,330	1.88E+15	4.30E+12	1.4	29	100	15	100	60
LC2	1.89E+06	29,273	1.88E+15	6.67E+12	3.0	20	100	25	100	60
LC3	3.07E+06	36.910	1.88E+15	9.36E+12	1.0	4	100	26	100	60
MC1	1.55E+06	6.622	1.88E+15	9.77E+12	7.7	37	100	11	75	50
MC2	9.11E+05	7.107	1.88E+15	1.69E+13	7.5	19	100	12	70	60
MC3	2.17E+06	16.843	1.88E+15	6.91E+12	3.1	18	100	5	65	50
HC1	4.68E+06	3.758	1.88E+15	1.33E+13	12.0	13	100	11	100	43
HC2	8.64E+05	13.702	1.88E+15	1.09E+13	4.0	62	100	29	10	43
HC3	2.47E+06	8.671	1.88E+15	9.88E+12	4.8	11	100	15	30	43
Objectiv						Maximiz				
e	Minimize Lowest	Minimize Lowest	Maximize Highest	Minimize Lowest	Maximize Highest	e Highest	Maximize Highest	Maximize Highest	Maximize Highest	Maximize Highest
Goal	value	value	value	value	value	value	value	value	value	value

\*Annual discount rate of 12%.

According to the goal programming analysis, the system with the best sustainability performance is MC1, followed by MC2, HC2, HC3, HC1, MC3, LC2, LC1 and LC3 (Figure 2). Overall, the low control systems performed poorer at Sector 1 and Sector 3, but had a better performance at the sector 5. The MC2 and HC2 had the worst performance at sector 2. The HC2 had the best performance at sector 4, followed by LC3 and LC2.

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Figure 3 – Goal programming analysis accounting for the distance of each lambari farm studied 786 from the goals established at each sector of the Five Sector Sustainability model (5SEnSU). 787 Higher values indicate higher distance from the goal and lower sustainability. LC= low control 788 farm: MC = moderate control farm: HC = high control farm: numbers 1, 2 and 3 represent different 789 790 farms within the same control level. Numbers over the bars represent the Sustainability Synthetic Indicator of Systems (SSIS); SSIS S1 = Sustainability Synthetic Indicator of System 1; SSIS S2 791 = Sustainability Synthetic Indicator of System 2; SSIS S3 = Sustainability Synthetic Indicator of 792 System 3; SSIS S4 = Sustainability Synthetic Indicator of System 4; SSIS S5 = Sustainability 793 Synthetic Indicator of System 5. . 794

## 796 **4. Discussion**

The multi-criteria assessment of the sustainability of lambari aquaculture 797 indicates that the lower control systems are less sustainable compared to the 798 moderate and high control systems, which the MC1 achieved the highest 799 sustainability. These systems had a poorer performance at sector 1, which is 800 demonstrated by the larger deviation of the obtained indicators from the chosen 801 802 goal. On the other hand, the deviation was lower for the sector 2. The LC systems also had a large deviation from the goals of system 3, which evidences the poorer 803 performance of the system on the economic sphere. Nevertheless, the LC 804 systems had a better performance in the sectors 4 and 5, which indicates a higher 805 level of sustainability on the social sphere compared to the higher control 806 systems. 807

The sector 1 represents the natural environment as a provider of resources 808 for the production system (Giannetti et al., 2019b). The indicator Unity Emergy 809 Value (UEV) accounts the systems efficiency on converting natural resources into 810 811 products, in which the lower the UEV, the higher the efficiency (Giannetti et al., 2019a; Odum, 1996). The second indicator selected was water use (m<sup>3</sup>/kg), 812 which evidences the systems' dependence on the water resources, in which the 813 higher the value, the lower the sustainability (Boyd et al., 2007; Valenti et al., 814 2018). Freshwater is a core natural resource for rural aquaculture systems, that 815 has been threatened by irrational consumption, pollution and climate change 816 (Ahmed et al., 2018; Goddard and Delghandi, 2020). The LC systems are low 817 efficient in the use of water and other natural resources, which could be a result 818 of poorer management practices of production. As demonstrated in Chapter 1, 819 these systems lack of technical assistance on simple management practices that 820 greatly influences their sustainability, such as the replacement of groundwater by 821 superficial sources and a proper feeding regime. The absence of technology 822 transfer policies that matches producers' realities is one of the greatest 823 bottlenecks for the sustainable development of the activity in Brazil (Henares et 824 al., 2019; Valenti et al., 2021). Therefore, extension programs devoted for 825 smallholder farms are urgent, and should focus on low-cost alternative practices 826 that could benefit farmers on the long term. 827

The environment represented by sector 2 act as a receiver of the pollution 828 that may be generated by the production activity (Giannetti et al., 2019b). The 829 discard of wastes compromises the natural functioning of the environment, 830 demanding energy and resources for its recovery, which results in the depletion 831 of natural stocks and, sometimes, the collapse of the ecosystem (Almeida et al., 832 2010; Bastianoni et al., 2016). On the other hand, the production system may 833 interact in a positive way with the surrounding environment, by improving 834 ecosystems functions that restore natural stocks (Coscieme et al., 2013; Willot et 835 al., 2019). The lambari pond systems interact in both ways with the adjacent 836 ecosystem (see Chapter 2). The positive functions generate services such as 837 water regulation and microclimate regulation, which are inherent of pond systems 838 and thus have a low variation among farms. The negative functions are caused 839 840 mainly by the discard of eutrophic effluents in the watershed. This disservice is higher at moderate and high-control farms, which were demonstrated by their 841 larger deviation from the goal at sector 2 (figure 2). 842

The indicators of the sector 3 demonstrate the economic performance of the 843 lambari farm systems. The low control systems are less productive compared to 844 MC and HC systems, which may be a result of the poorer management practices, 845 but also the lower quantity and quality of inputs, such as commercial feed. 846 Contrariwise, the I.R.R., which is a measure of the systems profitability on the 847 long term, was lower at LC3 (3.8%), HC3 (10.8%) and HC1 (12.6%), and higher 848 at HC2 (62%) and MC1 (37%). This high variation is expected in small farms, 849 since they usually lack of accurate accounting records and business plans 850 851 (Valenti et al., 2021). Besides, rural credit and business opportunities offers are hardly achievable by small producers. Consequently, most of the aquaculture 852 farms in Brazil are not economically sustainable (Valenti et al., 2021). 853 Nevertheless, many opportunities exist to overcome this issue. The precise 854 record of financial and production data, the compliance with the regulations for 855 obtaining credit, diversification of markets and products, and equity of 856 opportunities among small and big farms would strengthen the activity (Valenti et 857 al., 2021). 858

The aquaculture of low-trophic level fish is very important for the subsistence of rural populations worldwide (Ali et al., 2016; Béné et al., 2016;

Zhang et al., 2011). Lambari production started as an alternative income source 861 for the tilapia producers who were outcompeted by the larger investors, and 862 remains a social relevant activity (Valenti et al., 2021). All systems studied rely 863 on local labor and eight farms are over 10 years in the activity, which indicates 864 that lambari aquaculture contributed to the construction of self-identity of the 865 farmers as aquaculture producers, allowing them to maintain their rural 866 livelihoods (Goswami et al., 2017; Silva et al., 2020). Nevertheless, LC farmers 867 contributes larger to the local economy as they purchase their inputs and sell their 868 products locally. This fact indicates that the increase in technification of the 869 activity may increase the economic performance and attract larger investors, but 870 at a cost of contributing less to the local development and excluding smallholder 871 farms which can be an issue for sustainability. For facing that, public policies 872 873 should focus on diverse solutions for attending the problems of small, medium and large producers, instead of a single-oriented development plan (Brugère et 874 al., 2019; Valenti et al., 2021). 875

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## **5.** Conclusion

The lambari aquaculture has different performances of sustainability, 878 depending on the control of the key variables by the systems operators. Low 879 control farms performed better at the social dimension, contributing mainly for the 880 local socio-economic development and for the maintenance of rural livelihoods. 881 On the other hand, moderate and high control farms performed better at the 882 environmental and economic dimensions, because they use natural resources 883 more efficiently and achieve higher productivity and profitability. Results indicate 884 that the issues of smallholder aquaculture systems are variable among the three 885 dimensions of sustainability, which claims for strategies towards the sustainable 886 development that consider the diversity of problems in a holist perspective. For 887 that, a systemic approach and a multi-criteria assessment, based on the 888 889 biophysical aspects of the systems, such as the five sector sustainability assessment, can be a useful tool. 890

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#### 892 **References**

- Ahmed, N., Ward, J.D., Thompson, S., Saint, C.P., Diana, J.S., 2018. Blue-893 Green Water Nexus in Aquaculture for Resilience to Climate Change. Rev. 894 Fish. Sci. Aquac. 26, 139–154. 895 896 https://doi.org/10.1080/23308249.2017.1373743 Ali, H., Murshed-e-Jahan, K., Belton, B., Dhar, G.C., Rashid, H.O., 2016. 897 Factors determining the productivity of mola carplet (Amblypharyngodon 898 mola, Hamilton, 1822) in carp polyculture systems in Barisal district of 899 Bangladesh. Aquaculture 465, 198-208. 900 https://doi.org/10.1016/j.aguaculture.2016.09.017 901 Almeida, C.M.V.B., Jr, D.B., Bonilla, S.H., Giannetti, B.F., 2010. Resources, 902 Conservation and Recycling Identifying improvements in water 903 management of bus-washing stations in Brazil. "Resources, Conserv. 904 Recycl. 54, 821-831. https://doi.org/10.1016/j.resconrec.2010.01.001 905 Antonucci, F., Costa, C., 2020. Precision aquaculture: a short review on 906 engineering innovations. Aquac. Int. 28, 41–57. 907 https://doi.org/10.1007/s10499-019-00443-w 908
- Bastianoni, S., Coscieme, L., Pulselli, F.M., 2016. The input-state-output model
- and related indicators to investigate the relationships among environment,
- society and economy. Ecol. Modell. 325, 84–88.
- 912 https://doi.org/10.1016/j.ecolmodel.2014.10.015

Béné, C., Arthur, R., Norbury, H., Allison, E.H., Beveridge, M., Bush, S.,

- Campling, L., Leschen, W., Little, D., Squires, D., Thilsted, S.H., Troell, M.,
- 915 Williams, M., 2016. Contribution of Fisheries and Aquaculture to Food
- 916 Security and Poverty Reduction: Assessing the Current Evidence. World
- 917 Dev. 79, 177–196. https://doi.org/10.1016/j.worlddev.2015.11.007
- Boyd, C.E., D'Abramo, L.R., Glencross, B.D., Huyben, D.C., Juarez, L.M.,
- Lockwood, G.S., McNevin, A.A., Tacon, A.G.J., Teletchea, F., Tomasso,
- J.R., Tucker, C.S., Valenti, W.C., 2020. Achieving sustainable aquaculture:
- <sup>921</sup> Historical and current perspectives and future needs and challenges. J.
- 922 World Aquac. Soc. 51, 578–633. https://doi.org/10.1111/jwas.12714

923	Boyd, C.E., Tucker, C., Mcnevin, A., Bostick, K., Clay, J., 2007. Indicators of
924	Resource Use Efficiency and Environmental Performance in Fish and
925	Crustacean Aquaculture. Rev. Fish. Sci. 15, 327–360.
926	https://doi.org/10.1080/10641260701624177
927	Brugère, C., Aguilar-Manjarrez, J., Beveridge, M.C.M., Soto, D., 2019. The
928	ecosystem approach to aquaculture 10 years on – a critical review and
929	consideration of its future role in blue growth. Rev. Aquac. 11, 493–514.
930	https://doi.org/10.1111/raq.12242
931	Coscieme, L., Pulselli, F.M., Jørgensen, S.E., Bastianoni, S., Marchettini, N.,
932	2013. Thermodynamics-based categorization of ecosystems in a socio-
933	ecological context. Ecol. Modell. 258, 1–8.
934	https://doi.org/10.1016/j.ecolmodel.2013.02.031
935	David, L.H., Pinho, S.M., Agostinho, F., Kimpara, J.M., Keesman, K.J., Garcia,
936	F., 2020. Emergy synthesis for aquaculture: A review on its constraints and
937	potentials. Rev. Aquac. raq.12519. https://doi.org/10.1111/raq.12519
938	Dawood, M.A.O., Koshio, S., Abdel-Daim, M.M., Van Doan, H., 2019. Probiotic
939	application for sustainable aquaculture. Rev. Aquac. 11, 907–924.
940	https://doi.org/10.1111/raq.12272
941	FAO, 2020. The State of World Fisheries and Aquaculture 2020. Sustainability
942	in action. FAO, Rome. https://doi.org/10.4060/ca9229en
943	Fonseca, T., Costa-Pierce, B.A., Valenti, W.C., 2017. Lambari Aquaculture as a
944	Means for the Sustainable Development of Rural Communities in Brazil.
945	Rev. Fish. Sci. Aquac. 25, 316–330.
946	https://doi.org/10.1080/23308249.2017.1320647
947	Garlock, T., Asche, F., Anderson, J., Bjørndal, T., Kumar, G., Lorenzen, K.,
948	Ropicki, A., Smith, M.D., Tveterås, R., 2020. A Global Blue Revolution:
949	Aquaculture Growth Across Regions, Species, and Countries. Rev. Fish.
950	Sci. Aquac. 28, 107-116. https://doi.org/10.1080/23308249.2019.1678111
951	Giannetti, B.F., Marcilio, M.D.F.D.F.B., Coscieme, L., Agostinho, F., Liu, G.,
952	Almeida, C.M.V.B., 2019a. Howard Odum's "Self-organization, transformity
953	and information": Three decades of empirical evidence. Ecol. Modell. 407,

Orientador Wagner C. Valenti

954	108717. https://doi.org/10.1016/j.ecolmodel.2019.06.005
955	Giannetti, B.F., Sevegnani, F., Almeida, C.M.V.B., Agostinho, F., Moreno
956	García, R.R., Liu, G., 2019b. Five sector sustainability model: A proposal
957	for assessing sustainability of production systems. Ecol. Modell. 406, 98–
958	108. https://doi.org/10.1016/j.ecolmodel.2019.06.004
959	Goddard, S., Delghandi, M., 2020. Importance of the Conservation and
960	Management of Freshwater to Aquaculture, in: Encyclopedia of the World's
961	Biomes. Elsevier, pp. 35–44. https://doi.org/10.1016/B978-0-12-409548-
962	9.11954-2
963 964 965	Goswami, R., Saha, S., Dasgupta, P., 2017. Sustainability assessment of smallholder farms in developing countries. Agroecol. Sustain. Food Syst. 41, 546–569. https://doi.org/10.1080/21683565.2017.1290730
966 967 968 969	Henares, M.N.P., Medeiros, M. V., Camargo, A.F.M., 2019. Overview of strategies that contribute to the environmental sustainability of pond aquaculture: rearing systems, residue treatment, and environmental assessment tools. Rev. Aquac. 1–18. https://doi.org/10.1111/raq.12327
970 971 972	Humphries, F., Benzie, J.A.H., Morrison, C., 2019. A systematic quantitative literature review of aquaculture genetic resource access and benefit sharing. Rev. Aquac. 11, 1133–1147. https://doi.org/10.1111/raq.12283
973	IBGE - Instituto Brasileiro de Geografa e Estatística, 2020. Pesquisa da
974	Pecuária Municipal 2017, 2018 e 2019.
975	Lulijwa, R., Rupia, E.J., Alfaro, A.C., 2019. Antibiotic use in aquaculture,
976	policies and regulation, health and environmental risks: a review of the top
977	15 major producers. Rev. Aquac. 2000, 1–24.
978	https://doi.org/10.1111/raq.12344
979	Mungkung, R., Aubin, J., Prihadi, T.H., Slembrouck, J., Van Der Werf, H.M.G.,
980	Legendre, M., 2013. Life cycle assessment for environmentally sustainable
981	aquaculture management: A case study of combined aquaculture systems
982	for carp and tilapia. J. Clean. Prod. 57, 249–256.
983	https://doi.org/10.1016/j.jclepro.2013.05.029

984	Odum, H.T. (Ed.), 1996. Environmental Accounting: Emergy and Environmental
985	Decision Making, 1st ed. John Wiley & Sons, New York.
986	Silva, J.R. da, Mauad, J.R.C., Domingues, C.H. de F., Marques, S.C.C.,
987	Borges, J.A.R., 2020. Understanding the intention of smallholder farmers to
	adopt fish production. Aquac. Reports 17, 100308.
988	
989	https://doi.org/10.1016/j.aqrep.2020.100308
990	Tacon, A.G.J., 2020. Trends in Global Aquaculture and Aquafeed Production:
991	2000–2017. Rev. Fish. Sci. Aquac. 28, 43–56.
992	https://doi.org/10.1080/23308249.2019.1649634
993	United Nations, 2015. Transforming our World: the 2030 Agenda for
994	Sustainable Development.
995	Valenti, W.C., Barros, H.P., Moraes-Valenti, P., Bueno, G.W., Cavalli, R.O.,
996	2021. Aquaculture in Brazil: past, present and future. Aquac. Reports 19,
997	100611. https://doi.org/10.1016/j.aqrep.2021.100611
998	Valenti, W.C., Kimpara, J.M., Preto, B. de L., Moraes-Valenti, P., 2018.
999	Indicators of sustainability to assess aquaculture systems. Ecol. Indic. 88,
1000	402-413. https://doi.org/10.1016/j.ecolind.2017.12.068
1001	Willot, PA.A., Aubin, J., Salles, JM.M., Wilfart, A., 2019. Ecosystem service
1002	framework and typology for an ecosystem approach to aquaculture.
1003	Aquaculture 512, 734260.
1004	https://doi.org/10.1016/j.aquaculture.2019.734260
1005	Zhang, L.X., Ulgiati, S., Yang, Z.F., Chen, B., 2011. Emergy evaluation and
1006	economic analysis of three wetland fish farming systems in Nansi Lake
1007	area, China. J. Environ. Manage. 92, 683–694.
1008	https://doi.org/10.1016/j.jenvman.2010.10.005
1000	

# 1010 CONCLUSÕES GERAIS

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- Os sistemas de produção de lambari dependem majoritariamente de recursos não-renováveis, ou de baixa renovabilidade, independentemente do nível de controle adotado pelo produtor (baixo, moderado ou alto).
- Os principais recursos utilizados são ração comercial e água, sendo que os sistemas de baixo e médio controle dependem principalmente de água subterrânea, que é ambientalmente mais custoso comparado ao uso da água superficial.
- Estratégias de produção que incluem: mudança na fonte de abastecimento
   hídrico, controle de fertilização, aumento de sobrevivência na produção de
   alevinos, e substituição da fonte de proteína animal por fontes vegetais na
   ração, resultam em aumento na performance em emergia e na resiliência
   dos sistemas.
- Esforços ainda são necessários para melhores práticas de manejo da produção da lambaricultura, mas os resultados encontrados no capítulo 1 destacam que mudanças simples fazem a diferença na sustentabilidade dos sistemas.
- Além da provisão de peixes, os viveiros escavados de aquicultura de água
   doce fornecem serviços e desserviços ecossistêmicos que dependem das
   funções ecológicas desse tipo de sistema.
- Os serviços de regulação hídrica e regulação do microclima são
   características positivas inerentes ao sistema, e dependem das funções
   ecossistêmicas de recarga das águas subterrâneas e evaporação.
- A absorção e as emissões de gases do efeito estufa variam de acordo com as práticas de manejo aplicadas ao sistema produtivo, e neste estudo, apresentou um efeito neutro no serviço de regulação climática global.
- O efluente de qualidade eutrófica das fazendas estudadas causa o maior
   desserviço, uma vez que provoca o desvio de recursos para a recuperação
   da qualidade da água.
- O balanço entre as externalidades positivas e negativas na aquicultura do
   lambari indicam que os benefícios superam os prejuízos, desde que o
   sistema opere sob a capacidade de suporte da natureza.

- A lambaricultura tem diferentes desempenhos de sustentabilidade,
   dependendo do controle de variáveis produtivas pelos operadores dos
   sistemas.
- As fazendas de baixo controle apresentaram melhor desempenho na dimensão social, contribuindo principalmente para o desenvolvimento socioeconômico local e para a manutenção do modo de vida rural.
- As fazendas de moderado e alto controle apresentaram melhor desempenho
   nas dimensões ambiental e econômica, pois utilizam os recursos naturais de
   forma mais eficiente e alcançam maior produtividade e lucratividade.
- Os resultados encontrados no capítulo 3 indicam que os problemas dos pequenos produtores aquícolas são variáveis entre as três dimensões da sustentabilidade, o que demanda uma abordagem sistêmica e uma avaliação multicritério, com base nos aspectos biofísicos dos sistemas.

# 1057 **REFERÊNCIAS COMPLEMENTARES**

- Antonucci, F., Costa, C., 2020. Precision aquaculture: a short review on engineering innovations. Aquac.
   Int. 28, 41–57. https://doi.org/10.1007/s10499-019-00443-w
- Béné, C., Arthur, R., Norbury, H., Allison, E.H., Beveridge, M., Bush, S., Campling, L., Leschen, W., Little,
  D., Squires, D., Thilsted, S.H., Troell, M., Williams, M., 2016. Contribution of Fisheries and
  Aquaculture to Food Security and Poverty Reduction: Assessing the Current Evidence. World Dev.
  79, 177–196. https://doi.org/10.1016/j.worlddev.2015.11.007
- Boyd, C.E., D'Abramo, L.R., Glencross, B.D., Huyben, D.C., Juarez, L.M., Lockwood, G.S., McNevin, A.A.,
   Tacon, A.G.J., Teletchea, F., Tomasso, J.R., Tucker, C.S., Valenti, W.C., 2020. Achieving sustainable
   aquaculture: Historical and current perspectives and future needs and challenges. J. World Aquac.
   Soc. 51, 578–633. https://doi.org/10.1111/jwas.12714
- Brugère, C., Aguilar-Manjarrez, J., Beveridge, M.C.M., Soto, D., 2019. The ecosystem approach to
   aquaculture 10 years on a critical review and consideration of its future role in blue growth. Rev.
   Aquac. 11, 493–514. https://doi.org/10.1111/raq.12242
- 1072 Costa-pierce, B.A., Page, G.G., 2010. Sustainability Science in Aquaculture. Ocean Farmign Sustain.
   1073 Aquac. Sci. Technol. Encycl. Sustain. Sci. Technol. 1–28. https://doi.org/10.1007/978-1-4614-5797 1074 8\_175
- Dawood, M.A.O., Koshio, S., Abdel-Daim, M.M., Van Doan, H., 2019. Probiotic application for sustainable
   aquaculture. Rev. Aquac. 11, 907–924. https://doi.org/10.1111/raq.12272
- 1077 FAO, 2020. The State of World Fisheries and Aquaculture 2020. Sustainability in action. FAO, Rome.
   1078 https://doi.org/10.4060/ca9229en
- Fonseca, T., Costa-Pierce, B.A., Valenti, W.C., 2017. Lambari Aquaculture as a Means for the Sustainable
   Development of Rural Communities in Brazil. Rev. Fish. Sci. Aquac. 25, 316–330.
   https://doi.org/10.1080/23308249.2017.1320647
- Giannetti, B.F., Sevegnani, F., Almeida, C.M.V.B., Agostinho, F., Moreno García, R.R., Liu, G., 2019. Five
   sector sustainability model: A proposal for assessing sustainability of production systems. Ecol.
   Modell. 406, 98–108. https://doi.org/10.1016/j.ecolmodel.2019.06.004
- Goswami, R., Saha, S., Dasgupta, P., 2017. Sustainability assessment of smallholder farms in developing
   countries. Agroecol. Sustain. Food Syst. 41, 546–569.
   https://doi.org/10.1080/21683565.2017.1290730
- Henares, M.N.P., Medeiros, M. V., Camargo, A.F.M., 2019. Overview of strategies that contribute to the
   environmental sustainability of pond aquaculture: rearing systems, residue treatment, and
   environmental assessment tools. Rev. Aquac. 1–18. https://doi.org/10.1111/raq.12327
- Humphries, F., Benzie, J.A.H., Morrison, C., 2019. A systematic quantitative literature review of
   aquaculture genetic resource access and benefit sharing. Rev. Aquac. 11, 1133–1147.
   https://doi.org/10.1111/raq.12283
- 1094 IBGE Instituto Brasileiro de Geografia e Estatística, 2020. Produção da Pecuária Municipal. Disponível
   1095 em: https://www.ibge.gov.br/estatisticas/economicas/agricultura-e-pecuaria/9107-producao-da 1096 pecuaria-municipal.html?=&t=publicacoes. Acesso em fevereiro de 2021.
- Jung, S., Rasmussen, L.V., Watkins, C., Newton, P., Agrawal, A., 2017. Brazil's National Environmental
   Registry of Rural Properties: Implications for Livelihoods. Ecol. Econ. 136, 53–61.
   https://doi.org/10.1016/j.ecolecon.2017.02.004
- Lee, D.J., Brown, M.T., 2021. Estimating the Value of Global Ecosystem Structure and Productivity: A
   Geographic Information System and Emergy Based Approach. Ecol. Modell. 439, 109307.
   https://doi.org/10.1016/j.ecolmodel.2020.109307
- 1103 Lulijwa, R., Rupia, E.J., Alfaro, A.C., 2019. Antibiotic use in aquaculture, policies and regulation, health

- and environmental risks: a review of the top 15 major producers. Rev. Aquac. 2000, 1–24.
  https://doi.org/10.1111/raq.12344
- Odum, H.T. (Ed.), 1996. Environmental Accounting: Emergy and Environmental Decision Making, 1st ed.
   John Wiley & Sons, New York.
- Silva, N.J., Lopes, M.C., Fernandes, J.B.K., Henriques, M.B., 2011. Caracterização Dos Sistemas De Criação
   E Da Cadeia Produtiva Do Lambari No Estado De São Paulo, Brasil. Informações Econômicas 41,
   117–28.
- 1111 Tacon, A.G.J., 2020. Trends in Global Aquaculture and Aquafeed Production: 2000–2017. Rev. Fish. Sci.
   1112 Aquac. 28, 43–56. https://doi.org/10.1080/23308249.2019.1649634
- 1113 United Nations, 2020. UN Research Roadmap for the COVID-19 Recovery. New York.
- 1114 Valenti, W.C., Barros, H.P., Moraes-Valenti, P., Bueno, G.W., Cavalli, R.O., 2021. Aquaculture in Brazil:
  1115 past, present and future. Aquac. Reports 19, 100611.
- 1116 https://doi.org/10.1016/j.aqrep.2021.100611